## ASSESSMENT OF REARING CAPACITY

 FOR CONSIDERATION OF RE-INTRODUCING SOCKEYE SALMON TO THE COQUITLAM RESERVOIR
## Prepared for:

BC Hydro Bridge Coastal Fish and Wildlife Restoration Program 6911 Southpoint Drive (E14),
Burnaby, BC V3N 4X8

BCRP Report No. \#05.Co. 13

July 2006

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## EXECUTIVE SUMMARY

An assessment of fish abundance and biomass, and limnological characteristics in Coquitlam Reservoir was initiated in 2004, to determine whether re-introducing anadromous sockeye salmon to Coquitlam would be rearing-limited. This report presents the 2005 study results with comparisons with the 2004 findings.

The Coquitlam Reservoir is oligotrophic (unproductive), and is a major source of high quality water for the Greater Vancouver Regional District (GVRD). The limnological characteristics of the reservoir were similar during 2004-2005. This lake is characterized by low nutrient concentrations (is phosphorous limited), low phytoplankton biomass, and good water clarity. Its relatively cool water temperature regime, high dissolved oxygen levels, and favourable water quality conditions make it suitable for resident cold-water fishes. It also has low zooplankton stocks ( $1.2 \mathrm{ug} / \mathrm{L}$ ) compared with other west coast oligotrophic lakes, which may be limiting fish production.

Since Coquitlam Reservoir is an important source of drinking water for the GVRD, there is concern whether the introduction of fishes, such as anadromous sockeye salmon, will have negative effects on water quality for human consumption. Previous research has found strong interactions among planktivorous fishes, zooplankton, and phytoplankton and water quality. However, there is no research reviewed here that demonstrates direct linkages between sockeye introduction and water quality.

The food chain structure, especially as it relates to the zooplankton standing crop and its relationship to fisheries production, was examined. Analyses of fish stomach contents and stable isotope levels in fish tissue indicated low pelagic (open water) productivity and the importance of nutrients and foods from nearshore and terrestrial areas in fish foraging. The data showed that different fish species rely on quite different food sources. Only kokanee and threespine stickleback forage in the pelagic habitat and obtain their carbon inputs essentially from pelagic food sources.

A total fish population of 404,177 and 194,604 was estimated using hydroacoustics in May and November 2005, respectively. Of these, approximately $37-40 \%$ was estimated to be kokanee. In May, the total biomass of fish was estimated to be approximately $18,014 \mathrm{~kg}$ (range $15,516 \mathrm{~kg}$ to $20,511 \mathrm{~kg}$ ), and estimated fish production ranged between $290 \mathrm{fish} / \mathrm{ha}(13 \mathrm{~kg} / \mathrm{ha}$ ) and $384 \mathrm{fish} / \mathrm{ha}$ ( $17 \mathrm{~kg} / \mathrm{ha}$ ). In November, the total biomass of fish was estimated to be approximately $42,260 \mathrm{~kg}$ (range $25,095 \mathrm{~kg}$ to $59,425 \mathrm{~kg}$ ), and estimated fish production ranged between 96 fish/ha (21 $\mathrm{kg} / \mathrm{ha}$ ) and 228 fish/ha ( $50 \mathrm{~kg} / \mathrm{ha}$ ).

Empirical data on current zooplankton and kokanee production indicate that Coquitlam Reservoir is likely to support a relatively small sockeye smolt population (i.e. less than 1 million smolts). Available lake shore spawning habitat in Coquitlam Reservoir could potentially support between 3,000 and 5,000 female sockeye spawners and 5,000 female sockeye would produce a smolt population of approximately $400,0004.5 \mathrm{~g}$ sockeye smolts. This is consistent with zooplankton biomass and observed fall $0+1+$ kokanee biomass in the fall of 2004.

Predicted sockeye smolt biomass (kg) using a relationship with seasonal mean zooplankton biomass was 1853 in 2004 and 1129 in 2005. These may underestimate the total zooplankton
biomass for the reservoir because only one site was sampled and this was outside of the main area of pelagic fish abundance. However, these predicted sockeye smolt biomasses were similar to estimates of smolt biomass using hydroacoustics estimates of fall juvenile ( $0+$ and $1+$ ) kokanee abundance in 2004 and 2005 after accounting for $30 \%$ mortality to smolting (2184 and 529 , respectively).

Based on results from the first two years of study we conclude that, should re-introduction of sockeye proceed, a reasonable interim production target would be $400,0004.5 \mathrm{~g}$ sockeye smolts derived from a spawning population of 10,000 sockeye adults. This would balance with presumed available spawning habitat within the reservoir and would account for 1800 kg of sockeye/kokanee smolt biomass. If juvenile sockeye are introduced, it is likely that the kokanee population will decline due to direct competition for the same prey species and similar prey sizes. Any introduction of sockeye to the reservoir will require a pre-cautionary approach and a detailed limnological and fish monitoring program.

## INTRODUCTION

The restoration of anadromous fish runs, where practical, is a key objective of the Bridge-Coastal Fish and Wildlife Restoration Program (BCRP). In the Coquitlam-Buntzen BC Hydro system, numerous interested parties including government agencies, the Kwikwetlem First Nation, stewardship groups, environmental Non-Government Organizations (NGOs), and concerned citizens have an interest in restoring anadromous salmon runs in the Coquitlam Reservoir.

In 2002, LGL Limited developed a framework for evaluating fish passage issues in the Bridge Coastal hydro operating area (Bocking and Gaboury 2002). Following this, the BCRP commissioned an evaluation of the feasibility of restoring anadromous fish stocks into the Coquitlam Reservoir (Bocking and Gaboury 2003). Bocking and Gaboury (2003) estimated the rearing capacity for sockeye salmon of Coquitlam Reservoir using two models; the Euphotic Volume (EV) model of Koenings and Burkett (1987) and the Photosynthetic Rate (PR) model of Shortreed et al. (2000). Both these models estimate sockeye biomass based on the amount of physical space available to juvenile sockeye during the primary growing season of May - October. Bocking and Gaboury (2003) also made several recommendations for future study on issues pertaining to anadromous fish passage and water quality in Coquitlam Reservoir.

This 2005 study provides:

- A limnological characterization of the reservoir;
- Information on the abundance and biomass of the existing kokanee population and other fish species in Coquitlam Reservoir; and
- An assessment of rearing capacity for Kokanee/sockeye and potential effects of reintroducing sockeye in Coquitlam Reservoir.

An assessment of spawning for kokanee in the Coquitlam Reservoir is addressed in a parallel report by Gaboury and Murray (2006).

## Goals and objectives

Three specific objectives related to the limnology and fish populations of Coquitlam Reservoir were developed in cooperation with a number of agencies, institutions and concerned groups. Basically, these objectives adhere to the approach outlined by Bocking and Gaboury (2002) as follows:

1. Assess the current limnological features (physical, chemical, phytoplankton and zooplankton) of Coquitlam Reservoir to assist decision making on the potential effects of re-introducing sockeye salmon to the reservoir,
2. Determine the abundance, biomass and age structure of the existing kokanee population in Coquitlam Reservoir using hydroacoustic and various netting techniques, and characterize the fish community, and
3. Evaluate the current biomass of kokanee in the reservoir against the potential sockeye biomass at capacity reported by Bocking and Gaboury (2003), and identify limiting factors and potential initial stocking levels of sockeye in the reservoir.

## COQUITLAM RESERVOIR STUDY AREA

The Coquitlam Reservoir, located in southwest British Columbia and comprising an area of approximately 1200 ha (Figure 1), is a major source of domestic water for the Greater Vancouver region. The area is characterized by west coast maritime air with cool wet winters and warm dry summers. The reservoir has mean and maximum depths of approximately 87 and 187 m , respectively, at a pool elevation of 152 m , with complete mixing occurring between November and March. It is approximately 12 km long with an average width of roughly 1 km , and is classified as a monomictic body of water with an ultra-oligotrophic status (Wetzel 2001). Some physical details of the reservoir are presented in Table 1 (Nordin and Mazumder 2005; James 2000).


Figure 1. Map of Coquitlam watershed showing local communities and features.

Table 1. Morphological characteristics of Coquitlam Reservoir (Nordin and Mazumder 2005; James 2000).

| Attribute | Measure |
| :---: | :---: |
| Lake Volume ( $\mathrm{m}^{3}$ ) | 1,044,000,000 |
| Mean depth (m) | 87 |
| Surface area ( $\mathrm{km}^{2}$ ), (ha) | 12 (1200) |
| Watershed area ( $\mathrm{km}^{2}$ ) | 212 |
| Watershed area contributing to reservoir ( $\mathrm{km}^{2}$ ) | 191 |
| Watershed to Reservoir area ratio ${ }^{\text {a }}$ | 15.9:1 |
| Normal operating elevation (m) | 137.48-154.86 |
| Normal operating elevation range (m) | 17.4 |
| Average annual precipitation (rain) (mm) | 3576.8 |
| Average annual precipitation (snow) (mm) | 158.2 |
| Inflow ( $\mathrm{m}^{3} / \mathrm{yr}$ ) | 725,000 |
| Mean inflow ( $\mathrm{m}^{3} / \mathrm{s}$ ) | 23 |
| Water Residence time (yr) | 1.44 |
| Sedimentation rates - 1967-1997 ( $\mathrm{g} / \mathrm{m}^{2} / \mathrm{yr}$ ), ( $\mathrm{t} / \mathrm{km}^{2} / \mathrm{yr}$ ) | $192\left(1.92 \times 10^{2}\right)$ |
| Sedimentation rates - 1990-2002 (g/m ${ }^{2} / \mathrm{yr}$ ) | 267 |
| Sedimentation rates - 1905-2002 (mm/year) ${ }^{\text {b }}$ | $1.8 \mathrm{~mm} / \mathrm{year}$ |
| ${ }^{\text {a }}$ to the mouth of lower Coquitlam |  |
| ${ }^{\mathrm{b}}$ over the period |  |

In November 2002, a preliminary assessment of the fish population in the reservoir and potential salmonid spawning habitat in the upper Coquitlam watershed was conducted by LGL Limited (Bocking and Gaboury, 2003). The Coquitlam system supports several species of salmonids including kokanee (Oncorhynchus nerka) and cutthroat trout (Salmo clarki), and a number of coarse fishes including peamouth chub (Mylocheilus caurinus), northern pike minnow (Ptychocheilus oregonensis), and largescale sucker (Catostomus macrocheilus).

In September 2004, a study was conducted to assess fish abundance and biomass (Bussanich et al. 2005). The total population of fish in Coquitlam Reservoir in October was estimated at 648,000 fish ( $\pm 13 \%$ ) using hydroacoustics. Of this total, approximately $40 \%$ were kokanee. Total kokanee biomass was estimated at $7,700 \mathrm{~kg}$ (range 4,500 to 15,800 ) and of these; $3,100 \mathrm{~kg}$ were $0+(240$ kg ) and $1+$ kokanee ( 2880 kg ). By numbers, eighty-seven percent of the kokanee were in their first or second year of life (termed age-0 and age-1 fish) while $13 \%$ were in their third or fourth year (termed age-2 and age-3). The observed sex ratio among the age- 2 and- 3 kokanee was five males to one female.

Gaboury and Murray (2006) estimated that there is sufficient spawning habitat in the lake (1500 $\mathrm{m}^{2}$ ) below the 140 m contour and also in Cedar and Beaver creeks ( $1000 \mathrm{~m}^{2}$ ) to support a kokanee population of 4500 females or a sockeye population of 1500 females.

If the maximum reservoir drawdown is maintained at the 144 m elevation as proposed by Bocking and Gaboury (2003), this would potentially support a kokanee population of about 10,200 females or an equivalent sockeye population of roughly 3400 females. Sediment core data suggest that Coquitlam Reservoir did not have a large population of resident salmonids prior to construction of the dam in 1905 (Nordin \& Mazumder 2005). Spawning habitat in the Coquitlam tributaries is currently limited, and kokanee or sockeye (O. nerka) are more likely to spawn in the fluvial fans along the lakeshore rather than in the tributaries.

Additional habitat ( $25,000 \mathrm{~m}^{2}$ ) has also been identified in the Upper Coquitlam River and could support another 40,000 female spawners.

## METHODS

## Field Schedule

Sampling of the reservoir's physical and chemical properties as well as phytoplankton and zooplankton was conducted monthly (January 2004-December, 2005) by the University of Victoria (Field et al. 2005). Juvenile and adult fish populations were sampled in May, September and November 2005, using a combination of hydroacoustics, gillnetting, minnow trapping, trawling, and visual surveys.

## Limnology Assessment (Year 2: January to December 2005)

## Sampling Sites

Four sampling stations were used for limnological study. Site 1 was located at $49^{\circ} 21.40 \mathrm{~N}, 122^{\circ}$ $47.18^{\prime} \mathrm{W}$; Site L2 was located at $49^{\circ} 22.57^{\prime} \mathrm{N}, 122^{\circ} 47.97^{\prime} \mathrm{W}$; Site 3 was located at $49^{\circ} 23.58^{\prime} \mathrm{N}$, $122^{\circ} 47.60^{\prime} \mathrm{W}$; and Site 4 was located at $49^{\circ} 24.88^{\prime} \mathrm{N}, 122^{\circ} 46.43^{\prime} \mathrm{W}$ using a global positioning system and marked fixed rafts. These sites were located in areas of depth ranging from 9 m to 187 m , covering the length of the lake (Figure 2).

## Physical Measurements

Temperature, conductivity, turbidity, and chlorophyll a measurements were collected hourly, and downloaded monthly throughout the year, at a depth of 4 m at all sites using a YSI model 6600 Sonde (Hoskin Scientific). Turbidity measurements were also collected hourly and downloaded monthly throughout the year at depths of $15 \mathrm{~m}, 30 \mathrm{~m}$ and 40 m at Sites 2L, 3L and 4L, respectively, using YSI model 600 Sonde (Hoskin Scientific). Generally, the turbidity and chlorophyll $a$ data provided by the YSI rafts did not match well with values from more reliable laboratory methods, and thus the turbidity and chlorophyll data from the YSI rafts are not discussed in this report. For more information on these turbidity data, refer to Field and Mazumder (2006). Monthly (from January 2004-December 2005) vertical profiles of temperature, turbidity, chlorophyll $a$, specific conductivity and dissolved oxygen were taken in situ with a YSI model 6600 Sonde (Hoskin Scientific) at each site at 1-m intervals from the surface to a maximum depth of 18 m .


Figure 2. Bathymetric map of Coquitlam Reservoir, showing limnological and fish sampling sites, 2004-2005.

Secchi disk readings of water transparency were measured monthly (time period as above) at each site using a standard 20 cm Secchi disk, to the nearest 0.25 m , without a viewing chamber. Light attenuation measurements were collected monthly from January to December 2005 at 1-m intervals from 0 to 9 m , using a Li-Cor-1400 Light meter and Li-Cor underwater sensor (Hoskin Scientific). Euphotic Zone Depth (EZD), defined as the depth at which light is $1 \%$ of the intensity of that at the surface, was determined for each sampling site and date by best-fit linear regression of light intensity with depth

Water samples were also collected monthly (time period as above) in epilimnetic (surface), metalimnetic (middle) and hypolimnetic (bottom) waters at each site using a Van Dorn water bottle for analysis of turbidity and chlorophyll $a$.

Air temperature, wind direction, wind speed, relative humidity and light saturation measurements were collected hourly at each site from January 2004 to December 2005 using a Li-Cor-190 SA Quantum Sensor (Hoskin Scientific), and downloaded monthly. Atmospheric data were collected for an ancillary Greater Vancouver Regional District (GVRD) project and are not reported in this document.

## Water Chemistry \& Nutrients

Water samples were collected monthly, from January 2004 to December 2005 for water chemistry and nutrient analysis at all sites (L1-L4 inclusive) in epilimnetic (surface) waters, and at Site L2 in metalimnetic (middle) and hypolimnetic (bottom) waters, using a Van Dorn water bottle.

Carbon samples were collected monthly in epilimnetic (surface) waters, metalimnetic (middle) and hypolimnetic (bottom) waters at all sites except Site L2 where carbon was only sampled at a depth of 4 m .

Handling and analysis of samples followed standard water sampling procedures of the NSERCIRC laboratory at the University of Victoria. Water samples were analysed for total phosphorous (TP), total dissolved phosphorous (TDP), soluble reactive phosphorous (SRP), ammonia-N $\left(\mathrm{NH}_{4}{ }^{+}\right)$, total Kjeldahl nitrogen (TKN), nitrate- $\mathrm{N}\left(\mathrm{NO}_{3}\right)$, nitrite- $\mathrm{N}\left(\mathrm{NO}_{2}\right)$, and sulphate- $\mathrm{S}\left(\mathrm{SO}_{4}\right)$. For carbon measurement, samples were collected monthly at a depth of 4 m at each of the four sites from January 2004 to December 2005.

All water samples to be analysed for nutrients and chemistry were collected in clean dark 2-L Nalgene bottles pre-rinsed with sample water, stored on ice and processed within 2 hours of collection. Processing consisted of filtering samples through nitrocellulose filters into acid-washed bottles for later analysis. Carbon samples, including total organic and dissolved organic carbon were filtered through ashed GFF glass microfibre filters and stored in glass vials with no head space. All samples were kept cool during transport to the University of Victoria, where nutrient samples were frozen and carbon samples were refrigerated until analysed.

All water samples were analyzed at the NSERC-IRC laboratory at the University of Victoria. Carbon samples were processed immediately using a Shimadzu Total Organic Carbon Analyzer. Nutrient samples (TP, TDP, TN, $\mathrm{NO}_{2}, \mathrm{NO}_{3}$ and $\mathrm{SO}_{4}$ ) were analyzed using a Zellweger Analytics Lachat QuickChem autoanalyser. The SRP and chemical anion samples were processed using a

High Performance Liquid Chromatography (HPLC) analyzer made by Dionne Industries. The $\mathrm{NH}_{4}{ }^{+}$samples were analyzed using a Pharmacia photospectrometer.

## Phytoplankton

To assess the seasonal fluctuation of algal populations and predict potential bloom problems, phytoplankton cells in whole water samples were identified and enumerated. Samples were settled in Utermohl settling chambers. An Olympus IMT-2 inverted research microscope was used to view the samples. Individual cells were identified to genus or species, measured and counted. QA/QC was done on randomly chosen samples monthly.

## Chlorophyll a

Chlorophyll $a(\mathrm{Chl} a)$ was measured hourly at a depth of 4 m at each site from January 2004 to December 2005 using a YSI 6600 Sonde (Hoskin Scientific), and downloaded monthly. To validate monitoring by the Sonde, Chl a samples were collected monthly in epilimnetic (surface), metalimnetic (middle) and hypolimnetic (bottom) waters at each site. Within 2 hours of sample collection, 1 L of water from each sample site was filtered through an ashed 47 mm diameter, 0.45 $\mu \mathrm{m}$ Whatman GFF glass microfibre filter. Samples were filtered and kept in the dark to prevent chlorophyll from degrading in light. Filters were then folded and placed in 15 ml conical tubes and kept cool during transport to the NSERC-IRC laboratory at the University of Victoria, where they were frozen until analysis.

Chlorophyll- $a$ samples were analysed with a Turner Designs Model Trilogy Fluorometer. Extraction of samples was carried out 18-24 hours prior to analysis using $95 \%$ ethanol added to the conical tubes. The ethanol extracts the chlorophyll from the filter and after being centrifuged the sample can be decanted off and analysed using the fluorometer.

## Zooplankton

Macrozooplankton (excluding nauplii and rotifers) density and biomass were monitored in Coquitlam Reservoir monthly from January to December in 2004 and 2005 at limnological sampling Site L2 only. Vertical plankton hauls were conducted at sampling Site L2 using a $64-\mu \mathrm{m}-\mathrm{mesh}$ net with a mouth diameter of 30 cm . A standard downrigger was used to lower and retrieve the plankton net from a depth of 27 m to the surface at a constant speed of $1 \mathrm{~ms}^{-1}$. Two years of hydroacoustic investigations of fish populations in Coquitlam Reservoir have shown that virtually all fish targets are within the top 30 m of the reservoir (Bussanich et al. 2005). All zooplankton collected were emptied into a 60 mL plastic bottle and kept cool until preserved using a $10 \%$ sugared formalin solution.

Zooplankton samples were analysed for species composition, abundance, and length at the NSERC-IRC laboratory at the University of Victoria. Samples were passed through a $64-\mu \mathrm{m}$ Nitex mesh sieve and carefully rinsed with tap water to remove preservative. Whole samples were then either enumerated, or if zooplankton density indicated, diluted and a sub-sample analysed. Sub-samples had to yield a minimum of 150 enumerated organisms, or additional sub-samples were enumerated. Sub-samples were always counted in their entirety. Samples were returned to preservative once they were enumerated, and stored at the NSERC-IRC laboratory.

When possible, zooplankton was identified to species using the keys of Pennak (1978) and Clifford (1991). Rotifers were not enumerated. Samples were enumerated at 12X to 16X magnification. The counting tray the zooplankton samples were placed in was a clear plastic block with six parallel, interconnected channels cut into its surface. This counting tray held approximately 5 mL of liquid. Enumeration was conducted using a binocular compound microscope connected to a CCD video monitor. This apparatus was run through a PC using ZCount software developed for the NSERC-IRC laboratory at the University of Victoria. This software counts and determines lengths, and from this calculates zooplankton biomass ( $\mathrm{mg} / \mathrm{m}^{3}$ ) based on empirical sampling of reservoirs (J. Edmundson, Alaska Dept. of Fish and Game, Soldotna, AK, pers. comm.). Lengths were converted to biomass using species-specific regression equations relating wet length to mean dry mass. The software also calculates zooplankton density based on the volume of the sub-samples counted, the size of the plankton net, and the length of the tow.

## Fish Populations

## Hydroacoustic Surveys

Acoustic sampling was conducted at night 16-17 May, 31 October, and 1 November, and also during the day on 1 November 2005 on the Coquitlam Reservoir, using a $200-\mathrm{kHz}$ and $420-\mathrm{kHz}$ frequency BioSonics DTX echo processor and two transducers (Photo 1). Both the $200-\mathrm{kHz}$ split beam (starboard mounted) and $420-\mathrm{kHz}$ split beam (port mounted), $6^{\circ}$ circular transducers were positioned 0.5 m below the water surface and aimed vertically to sample from 1.5 m below the surface to near bottom.


Photo 1. Echoprocessor used for hydroacoustic sampling on Coquitlam Reservoir, 2005.

A total of 21 transects was sampled twice on Coquitlam Reservoir at night during the May and October-November surveys, and a total of 21 transects were sampled once during the day in the October-November survey (Figure 3).

The threshold for the $200-\mathrm{kHz}$ transducer was set to -75 dB with a 0 dB power level setting, and the $420-\mathrm{kHz}$ transducer was set to -75 dB with a 0 dB power level setting. The sample rate for each transducer was 12 pings per second, and pinged alternately between the two transducers. Data were geo-referenced with a GPS and written to file with the hydroacoustic sample data.

The acoustic system was calibrated in the field using a standard tungsten carbide calibration sphere following data acquisition. A standard $21 \mathrm{~mm}(420 \mathrm{kHz}$ calibration) and $36 \mathrm{~mm}(200 \mathrm{kHz}$ calibration) tungsten carbide sphere was used to calibrate the acoustic system. The calibration sphere was lowered to 10 m below the transducer, positioned in the beam, and 2000 to 2500 pings were recorded to estimate target strength and align the aim of the system. In situ tests indicated that the transducers were calibrated correctly; therefore, no offset was done for either transducer.

Acoustic data were processed to estimate the mean acoustic size of fish for scaling the echo integration relative densities. Vertical and horizontal data files were processed to output split beam target size. Data were output with Echoview V3.2 to ASCII format files and filtered by location in the beam to use only those targets less than 4 dB off-axis, and a pulse length of 0.5 to 2 x the transmitted pulse length of 0.2 msec . Average TS and sigma were calculated for the survey.

The total reflected voltages from echo integration were converted to absolute areal densities (number/hectare) by scaling the voltages by the average density sigmas by report and transect. All data files were processed by echo integration with Echoview 3.2 software. Each vertical data file was processed with $5-\mathrm{m}$ vertical strata from 2 m below the surface to near bottom, with bottom removed by manual bottom-editing in the software. Areal densities were provided for transects 2 23 , and areal densities were provided for each $250-\mathrm{m}$ segment at transect 1 (North Basin) using Arc GIS ArcView 8.2 ${ }^{\circledR}$.

We assumed Love's (1977) equation for all aspects was representative of the target strength distribution:

$$
\mathrm{TS}=20 \log \mathrm{~L}-69.23 \text { (all aspects); }
$$

where,
TS $=$ target strength in decibels; and
$\mathrm{L}=$ fork length in centimeters.
Target strengths were used to estimate fish lengths, which were compared with lengths of fish in gillnet catches.


Figure 3. Map of Coquitlam Reservoir showing arbitrary basin boundaries and hydroacoustic survey transects, 2005.

## Fish Sampling Operations

A three-person crew conducted the fish gill-netting operation from 15 May to 17 May, 31 October to 2 November and 24 November 2005. Net sampling sites were located throughout the reservoir (Figure 2). A site was deemed suitable for sampling if it met the following criteria:
(1) Historical catch sites (Bocking and Gaboury 2003);
(2) Close proximity to alluvial fans;
(3) Relatively high densities of fish identified in hydroacoustic surveys; and
(4) Free from rocks and other debris on the bed bottom that might damage the nets or jeopardize crew safety.

## Gillnetting

Gillnetting (set nets and drift nets) was the primary fish collection method used from 15 May to 17 May, and from 31 October to 2 November to develop an inventory of the fish stocks present in accordance with Resource Inventory Committee (RIC) standards (Anon. 2001). Set netting was the primary technique used and involved anchoring a net from the shore to fish a perpendicular/lateral length nearshore with soak times greater than 1 h . Sunken and floating surface nets were located in littoral and limnetic areas to determine the vertical distribution of fish.

Multi- and single-panel gillnets were used to sample kokanee (juveniles and adults), trout, and coarse fish (peamouth chub, sculpins, suckers) (Appelberg 2000). All nets were constructed of double knotted, light green monofilament nylon (Miracle R-13 L) mesh, and a hang ratio of 2:1. The multi-panel net consisted of six nets or panels 15.2 m long, thread thickness ranging from 0.2 to 0.25 mm , and of mesh sizes $25,38,51,64,76$, and 89 mm , strung together in a "gang" to form a net 91.2 m long and 2.4 m deep.

Daily gillnetting effort was measured as mean fishing time in minutes, and was calculated as follows:
where,

$$
\mathrm{MT}=(\mathrm{SI}-\mathrm{FO})+\frac{[(\mathrm{FO}-\mathrm{SO})+(\mathrm{FI}-\mathrm{SI})]}{2}
$$

$\mathrm{MT}=$ Mean fishing time (min);
$\mathrm{SO}=$ time the gillnet first entered the water;
$\mathrm{FO}=$ time the gillnet was fully deployed;
$\mathrm{SI}=$ time the gillnet retrieval began; and
$\mathrm{FI}=$ time the gillnet retrieval was completed.
Catch-per-unit-effort (CPUE), $\mathrm{C}_{j}$, the number of fish caught per hour, was computed for set $j$ as:

$$
\mathrm{C}_{j}=\frac{\mathrm{N}}{\mathrm{MT}}
$$

where:
$\mathrm{N}=$ number of fish caught. All CPUE estimates were standardized to a fishing area of $90 \mathrm{~m}^{2}$.

## Trawling

Mid-water trawl sampling was conducted during the night of 24 November 2005 to validate the pelagic species composition of the fish observed acoustically, and collect age composition data of juvenile sockeye. Trawling in May was discontinued following a trial tow and loss of gear on 17 May. The trawl net ( 3 mx 7 mx 18 m ) was constructed of knotless, nylon multi-mesh (3.0, 13, 19, 51 and 102 mm ) (Photo 2) and towed behind the boat (Photo 3 ) at a speed of approximately 1 $\mathrm{m} / \mathrm{s}$, for 15 min in a northerly-southerly direction. The net was positioned using a bridle attached to the mainline of a portable winch, with the mouth of the net set agape using horizontally mounted 3 m steel pipes (attached to the head rope and footrope) and a pair of 15 kg cannon balls (attached to the footrope). Selected trawling depths (from near surface to 20 m depth) were achieved by varying the length of the main lines; trawling depths were monitored using a Vemco ${ }^{\circledR}$ data logger attached to the head rope. Catch-per-unit-effort was computed as the number of fish caught per hour trawling.


Photo 2. Mid-water trawl net of multi-mesh sizes used to collect juvenile kokanee in Coquitlam Reservoir, November 2005.


Photo 3. Boat used to pull the mid-water trawl assembly to collect juvenile kokanee in Coquitlam Reservoir, November 2005.

## Minnow Trapping

Gee-minnow trapping was the primary fish collection method used from 15 May to 17 May and from 31 October to 2 November to inventory juvenile coarse fish in accordance with Resource Inventory Committee (RIC) standards (Anon. 2001). Catch-per-unit-effort was computed as the number of fish caught per hour of minnow trapping.

## Fish Handling

All crew members were experienced with the handling techniques necessary to minimize stress on captured fish. The standard procedure following capture was to assess fish condition. Only fish deemed to be in good condition were released. The remaining fish were sacrificed, placed in labeled, aseptic plastic bags, stored in a cooler, and moved to a central location at the GVRD security building for processing or frozen storage within 3 h .

## Adult Kokanee Spawner Surveys

The lower 500 m of the Upper Coquitlam River and Cedar, Harmony and Falls creeks were walked, beginning at the streams' confluences with the reservoir, to determine the presence or absence of kokanee spawners. Each stream surveyed involved a crew of two people, with surveys conducted in September, October and November 2005 during low stream flows.

## Biological Sampling (Sex, Age, Size, Genetics, Diet)

All fish captured in gillnets, trawls and minnow traps were identified to species, classified as adult or juvenile using RIC standards (Anon. 2001), and enumerated. On-site sampling information recorded included date, time, gear type and set number, fish species, life stage and sex. The fish were bagged, stored in a cooler and transferred to the field processing station. The samples were then processed and, fork length (FL, mm) and wet weight (g) were recorded for each fish. The fish were sub-sampled by life stage (juvenile, adult) in both the nearshore and offshore samples from each of the three basins (North, Central and South). For each fish sub-sampled, stomach fullness was recorded as empty, $0-25 \%, 26-50 \%, 51-75 \%$, or $76-100 \%$ full, and the stomach was then removed, bagged and frozen for later dietary analysis. Muscle tissue samples ( $25 \times 25 \mathrm{~mm}$ ) from fish $>120 \mathrm{~mm}$ long, and the whole fish for specimens $<120 \mathrm{~mm}$ long, were bagged and frozen for later isotopic C:N analysis at the University of Victoria. Additional information for kokanee sampled in May, October, and November included age (from scale and otoliths), flesh colour, and DNA for stock discrimination purposes. Scale samples were interpreted by Birkenhead Scale Analysis (Lone Butte, BC) and reported as age-1-, 2- or 3-yr old fish. For DNA analysis a 5-mm diameter of tissue was taken from the adipose fin of adult kokanee, stored in ethanol, and processed and analyzed by Seastar Biotech Inc (Victoria, BC), and reported by Bussanich et al. (2006). Additional information for the October and November kokanee samples included total gonad weight (for both sexes) and egg diameter (mean of 10 eggs), external coloration and evidence of fraying of the caudal fin.

The contents of individual stomachs were placed in a Petri dish, and viewed using a binocular compound microscope at 8 to 12.5 X magnification. Any large insects ( $>2 \mathrm{~mm}$ ) were identified (terrestrial or aquatic), counted, and discarded. Gut parasites, small fish, spiders, and other large taxa were also identified, counted and discarded. The remaining organisms ( $<2 \mathrm{~mm}$ : zooplankton, small insects, immature parasites, and other small organisms) were enumerated using standard methods for zooplankton. The samples were diluted and sub-sampled, if warranted, so that a minimum of 150 organisms were counted per sample. The results from counting the zooplankton samples were recorded and the total number of organisms in the entire sample (if sub-sampled) was computed.

## Food Web Inter-relations

Samples of frozen fish muscle and stomach contents were freeze-dried and ground into a fine, homogenous powder with a mortar and pestle. The powder was weighed into tin cups, combusted and analyzed in a Costech 4010 Elemental Analyzer coupled to a Thermo Delta Advantage continuous flow isotope ratio mass spectrometer. Isotope ratios are reported in $\%$ ratios referenced against peedee belemnite carbonate (PDB) for $\sigma^{13} \mathrm{C}$ and atmospheric nitrogen (Air) for $\sigma^{15} \mathrm{~N}$. Analysis of replicate lab reference material indicated a standard error of $0.15 \%$ for $\sigma^{13} \mathrm{C}$ and $0.31 \%$ for $\sigma^{15} \mathrm{~N}$ (Shapna Mazumder, pers. comm.).

## Fish Community Standing Stock

A stratified random sampling design was used to estimate total fish standing stock (MacLennan and Simmonds 1992) in Coquitlam Reservoir. The surveyed area was stratified into 20 depth regions at $5-\mathrm{m}$ intervals from 0 m to 100 m below the surface. Strata volume for each of the three basins was derived using a bathymetric map, and Arc GIS ArcView 8.2 ${ }^{\circledR}$. For each depth stratum,
mean fish density was expanded in proportion to the volume sampled, and these were summed to estimate total fish standing stock for each basin (North, Central, and South; Figure 3). The variance and $95 \%$ confidence intervals were calculated for a stratified random sample as per Cochran (1977) for each of the standing stock estimates.

## Kokanee Standing Stock

Kokanee standing stock was based on all acoustic tracked fish at 5-m depth intervals from 0 m to 100 m below the surface. For each depth stratum, kokanee acoustic targets were classified by three size classes (length of each acoustic target was estimated using Love's equation (1977)). The three size classes ranged from $30 \mathrm{~mm}(-53 \mathrm{~dB})$ to $300 \mathrm{~mm}(-37 \mathrm{~dB})$ as identified by Bussanich et al. (2005). These acoustic size classes are related to kokanee sizes as sub-yearlings ( $30-80 \mathrm{~mm}$ ), yearlings ( $81-170 \mathrm{~mm}$ ), and adults (171-300 mm) (Teuscher et al. 1994). This size classification approximates age- 0 , age- 1 , and age- 2 and older kokanee, respectively. Acoustic targets below 30 mm and above 300 mm were not considered within the size range of juvenile or adult kokanee in the Coquitlam Reservoir. It was assumed that $100 \%$ of all acoustic targets between 30 and 170 mm were kokanee. Of the total fish standing stock between 170 and 300 mm , the catch data were used to estimate age-2 and older kokanee.

Biomass, $\mathrm{B}_{j}$, the kg of kokanee, was computed for age class $j$ as:

$$
\mathrm{B}_{j}=\quad N \mathrm{x} m
$$

where:
$N=$ estimated number of kokanee; and $m=$ estimated mean wet weight of each age-class.
A least squares regression of length and weight was used to predict the biomass of kokanee for each age class using data collected in 2005 (Ricker 1975).

## Lake Rearing Capacity and Kokanee Standing Crop

Bocking and Gaboury (2003) estimated the rearing capacity for sockeye in Coquitlam Lake using the EV model (Koenings and Burkett 1987) and PR model (Shortreed et al. 2000). In these models, euphotic zone depth (EZD) for Coquitlam was estimated from limited information on secchi depths. Two years of study of physical limnology at Coquitlam Reservoir now enables the calculation of EZD from light penetration data. As well, the model predictions of total number of sockeye were based on 3.5 g and 4.5 g average weights of smolts for the EV and PR models, respectively. In 2005, an experimental flow release by BC Hydro resulted in several thousand kokanee emigrating from the reservoir. Average fork length of these mostly $1+$ kokanee was between 80 and 90 mm (Decker and Lewis 2006). These lengths are consistent with 3.5 to 4.5 g sockeye smolts (Baxter and Bocking 2006).

With the availability of empirical data on euphotic zone depth (EZD) we were able to revise the original EV and PR estimates of total lake rearing capacity as well as estimate total sockeye smolt biomass (standing crop) from mean seasonal macro zooplankton (excluding nauplii) biomass using an equation from Koenings and Kyle (1997). This was done for both 2004 and 2005.

## RESULTS

## Limnological Assessment

## Water Temperature and Dissolved Oxygen

Coquitlam Reservoir has an annual pattern of thermal stratification. Using Site L2 as an example, in January through March 2005, the lake was $4-5^{\circ} \mathrm{C}$ throughout the water column (Figure 4;
Appendix 1). In April 2005 the lake began to warm and by the end of April the surface waters had warmed to $13{ }^{\circ} \mathrm{C}$ although the deeper waters were still at $4-5^{\circ} \mathrm{C}$. The lake continued to warm through to August when the maximum surface water temperatures were seen $\left(22^{\circ} \mathrm{C}\right)$ along with the maximum thermal stratification. The mixed surface warm layer (epilimnion) extended down to 6 m ; a wide thermocline (zone of rapid temperature decrease) occupied the depths from 6-12 m, where temperature decreased from 21 to $7{ }^{\circ} \mathrm{C}$. The deep-water zone (hypolimnion) was at depths below 12 m and had temperatures of 6 to $7{ }^{\circ} \mathrm{C}$. Water cooled after August and by October surface waters were $12^{\circ} \mathrm{C}$ and the lake mixed, likely in November as by 3 December the lake was again isothermal at $6{ }^{\circ} \mathrm{C}$.

Dissolved oxygen (DO) concentrations at all stations generally remained near the saturation concentration (Figure 5; Appendix 2). In surface waters, the minimum concentrations were measured in summer when temperatures were highest - when oxygen saturation capacity in water is reduced. There was no evidence of oxygen depletion or that fish would be limited by oxygen concentrations anywhere in the lake.


Figure 4. Variation in temperature with depth at Site L2 in Coquitlam Reservoir, 2005.


Figure 5. Variation in dissolved oxygen (DO) concentration with depth at Site L2 in Coquitlam Reservoir, 2005.

## Secchi Depth \& Turbidity

Water clarity as Secchi depth was very good in Coquitlam Reservoir throughout 2005, although somewhat lower than 2004. Secchi depth ranged from 2.5 to 10 m at the four sampling sites throughout the year (Figure 6; Appendix 3). The lowest water clarity was in January at Sites L3 and L4 and highest in August at Site L4. There seemed to be a general trend with water being slightly clearer, on average, with distance from the dam.

Turbidity data from lab analysis of grab water samples indicate that Coquitlam Reservoir waters are very clear. Turbidity was generally less than 1.0 nephelometric turbidity unit (NTU) at all depths, at all sites, on all dates, with the exception of January, when turbidity at a depth of 4 m ranged from just over 1 NTU (Sites L1, L2, and L4) to 1.6 NTU (Site L3) (Figure 7).


Figure 6. Water transparency as Secchi depth, at four sites in Coquitlam Reservoir, January to December, 2005.

## Water Chemistry \& Nutrients

The pH of Coquitlam Reservoir ranged from around neutral to slightly acidic with a range of 6-8 in 2005 (Appendix 4). Nutrient samples were collected at Site L2 from January to December 2005. Total phosphorous (TP) concentrations were very low and generally in the range from 1 to $5 \mu \mathrm{~g} / \mathrm{L}$; typical of ultra-oligotrophic (unproductive) lakes (Figure 8; Appendix 5). Two anomalously high samples in December 2005 are suspected of being inaccurate but included here.


Figure 7. Water transparency as turbidity at a depth of 4 m , at four sites in Coquitlam Reservoir, January to December, 2005.


Figure 8. Total Phosphorus (TP) at Site L2 in surface (epilimnetic), middle (metalimnetic) and bottom (hypolimnetic) waters in Coquitlam Reservoir, January to December, 2005.

Mean epilimnetic (surface) TP was $3.0 \mu \mathrm{~g} / \mathrm{L}$; including the high December value. Without the December value it was $2.6 \mu \mathrm{~g} / \mathrm{L}$.

Total nitrogen (TN) concentrations were also very low, ranging from $100-230 \mu \mathrm{~g} / \mathrm{L}$, typical of ultra-oligotrophic lakes and similar to 2004 (Figure 9; Appendix 6). TN remained below $200 \mu \mathrm{~g} / \mathrm{L}$ at most depths throughout the year and yearly mean epilimnetic TN was $136 \mu \mathrm{~g} / \mathrm{L}$, again similar to 2004.


Figure 9. Total Nitrogen (TN) at Site L2 in surface (epilimnetic), middle (metalimnetic) and bottom (hypolimnetic) waters in Coquitlam Reservoir, January to December, 2005.

Nitrate-nitrite was also measured with concentrations in the 50-110 ug/L range with no obvious annual pattern (Figure 10; Appendix 7).


Figure 10. Nitrogen as nitrate plus nitrite at Site L2 in surface (epilimnetic), middle (metalimnetic) and bottom (hypolimnetic) waters in Coquitlam Reservoir, January to October, 2005.

The nitrogen to phosphorus ratio (TN: TP) is a good indicator of which of these two nutrients is the limiting factor in lake productivity (phytoplankton growth); with TN: TP >20 indicating phosphorus limitation, while TN : $\mathrm{TP}<10$ indicates potential nitrogen limitation. TN: TP was always $>25$ throughout the year (phosphorus limitation) and the mean N : P ratio for the epilimnion was 59:1 (Figure 11; Appendix 8).


Figure 11. Ratio of Total Nitrogen to Total Phosphorus (TP:TN) in surface (epilimnetic), middle (metalimnetic) and bottom (hypolimnetic) waters at Site L2 in Coquitlam Reservoir, January to December, 2005.

Total organic carbon (TOC) and dissolved organic carbon (DOC) concentrations were also very low, typical of ultra-oligotrophic lakes (Figure 12 and Figure 13; Appendix 9). Total organic carbon varied only within a narrow range $(1.5-2.5 \mathrm{mg} / \mathrm{L})$ at all depths at all sites throughout the year. Yearly mean TOC in epilimnetic waters was $1.8 \mathrm{mg} / \mathrm{L}$, which was slightly higher than the 2004 data. Dissolved organic carbon concentrations were in the range of $1.6-2.1 \mathrm{mg} / \mathrm{L}$, again very low and indicative of the ultra-oligotrophic nature of Coquitlam Reservoir.

## Chlorophyll $a$

Chlorophyll a concentrations remained below $1.5 \mu \mathrm{~g} / \mathrm{L}$ all year at Site L2; somewhat lower than 2004, with peak readings in May and October (Figure 14; Appendix 10). As with nutrients, values this low (epilimnetic mean of $1.6 \mathrm{ug} / \mathrm{L}$ ) are typical of ultra-oligotrophic lakes.


Figure 12. Total Organic Carbon (TOC) in surface (epilimnetic), middle (metalimnetic) and bottom (hypolimnetic) waters at Site L2 in Coquitlam Reservoir, January to December, 2005.


Figure 13. Dissolved Organic Carbon (DOC) in surface (epilimnetic), middle (metalimnetic) and bottom (hypolimnetic) waters at Site L2 in Coquitlam Reservoir, January to December, 2005.


Figure 14. Chlorophyll $a$ in surface (epilimnetic), middle (metalimnetic) and bottom (hypolimnetic) waters at Site L2 in Coquitlam Reservoir, January to December, 2005.

## Phytoplankton

Like nutrient and Chlorophyll $a$ concentrations, phytoplankton biomass was low, following the pattern seen in 2004. Total phytoplankton biomass varied through the year, with peaks in March ( $>500 \mu \mathrm{~g} / \mathrm{L}$ ), July ( $400 \mathrm{ug} / \mathrm{L}$ ) and September ( $400 \mu \mathrm{~g} / \mathrm{L}$ ), somewhat higher than was seen in 2004. In the early part of the year, the phytoplankton community was dominated by Chrysophytes and Chlorophtes and in late summer and early fall by Cyanophyta (blue-green algae) and Dinophyceae (dinoflagellates) (Figure 15).

## Zooplankton

The zooplankton community (excluding rotifers) within Coquitlam Reservoir was numerically dominated by small taxa, principally Bosmina, cyclopoid copepods and copepod nauplii (larvae), with densities ranging from one to four animals per litre. Total densities of all other species combined remained below one animal per litre throughout the year (Figure 16). These low densities are typical of ultra-oligotrophic systems.

Zooplankton biomass exhibited seasonal variation, with minimum biomass in January ( $<0.25$ $\mathrm{ug} / \mathrm{L}$ ) and peaks in April ( $>3.5 \mathrm{ug} / \mathrm{L}$ ) and July (approximately $2.5 \mathrm{ug} / \mathrm{L}$ ) (Figure 17). Small taxa (nauplii, cyclopoid copepods, Bosmina) comprised the majority of zooplankton biomass all year except in July and August, when the larger Daphnia species contributed at least $50 \%$ to total biomass (Figure 17). The low values observed for zooplankton biomass are typical of ultraoligotrophic systems.


Figure 15. Phytoplankton biomass by major groups at Site L2 in Coquitlam Reservoir, January to December, 2005.


Figure 16. Zooplankton density at Site L2 in Coquitlam Reservoir, January to November, 2005.


Figure 17. Zooplankton biomass at Site L2 in Coquitlam Reservoir, January to November, 2005.

## Fish Assessments

## Fish Distribution and Abundance Using Hydroacoustics

The total mean fish target strength in Coquitlam Reservoir was -52.6 dB during the May survey and -46.0 dB during the October survey (Appendices 13-18). The mean size of fish tended to be smaller in the top 5 m and at depths exceeding 30 m in Coquitlam Reservoir during the May survey (ranging from -53.0 dB to -57.8 dB ) and in the top 10 m and 30 m interval during the November survey (ranging from -52.1 dB to -56.8 dB ). The mean size of fish ranged from -55.1 dB to -59.6 dB in May and from -55.1 dB to -59.6 dB in October. In May, a mean fish length of 113 mm was estimated using Love's equation (1977), while target distribution ranged between 41 mm and 179 mm (Figure 18). In November, a mean fish length of 223 mm was estimated using Love's equation (1977), while target distribution ranged between 46 mm and 281 mm .


Figure 18. Fish size distribution based on acoustic targets, Coquitlam Reservoir, 27 September, 2004, 15-17 May 2005, and 1 November 2005.

Fish densities (N/ha) in Coquitlam Reservoir varied by location and by season (Appendix 19 and 20). Fish were most abundant at the central region (between $49^{\circ} 23.20^{\prime} \mathrm{N}$ and $49^{\circ} 25.00^{\prime}$ ) during the May survey (Figure 19). In November fish were most abundant at the south region (between $49^{\circ} 21.70^{\prime} \mathrm{N}$ and $49^{\circ} 22.70^{\prime} \mathrm{N}$ ) and the central region (between $49^{\circ} 24.50^{\prime} \mathrm{N}$ and $49^{\circ} 25.00^{\prime} \mathrm{N}$ ) (Figure 20). The surface to 25 m interval contributed $95 \%$ of the total fish and densities ranged from $42 \mathrm{fish} /$ ha to $122 \mathrm{fish} /$ ha in May (Figure 21). The surface to 5 m depth interval of Coquitlam Reservoir contributed $48 \%$ of the total fish densities. Densities ranged from 7 fish/ha to 97 fish/ha between the surface and 35 m in November with fewer kokanee targets detected below 35 m .


Figure 19. Fish acoustic density surface model (inverse distance weighted) derived from sampling at Coquitlam Reservoir, 15 - 17 May 2005.


Figure 20. Fish acoustic density surface model (inverse distance weighted) derived from density sampling at Coquitlam Reservoir, 1 November 2005.


Figure 21. Fish densities (N/ha), by depth strata, using echo integration, Coquitlam Reservoir, 27 September 2004, 15-17 May 2005, and 1 November 2005.

## Gillnetting Spring Survey

Gillnetting in the Coquitlam Reservoir during the night on 15-17 May for a total 187-h effort resulted in a total catch of 257 fish: 137 kokanee ( 53 \%), 55 Northern pikeminnow ( 21 \%), 28 cutthroat trout ( $11 \%$ ), 27 peamouth chub ( $11 \%$ ), 7 sucker ( $3 \%$ ), and 3 redside shiner ( $1 \%$ ) (Figure 22, Appendix 21, Appendix 26, Photo 4).

A total CPUE (catch per hour standardized for a $90 \mathrm{~m}^{2}$ set area) of 3.23 fish was estimated for the gillnet operation. The proportion of the total CPUE was highest and similar in the South (48\%) and North ( $47 \%$ ) basins, with the Central Basin constituting only $4 \%$ the catch. CPUE ranged from a low 0.02 for redside shiner to a high 1.86 for kokanee. CPUE for kokanee was 0.86 ( $46 \%$ ), $0.14(8 \%)$, and $0.86(46 \%)$ in the North, Central, and South basins, respectively (Appendix 3). Of a total 41.7 kg of fish, 20.3 kg was Northern pikeminnow (49\%), 10.9 kg kokanee ( $26 \%$ ), 6.4 kg cutthroat trout ( $15 \%$ ), 2.1 kg sucker ( $5 \%$ ), 1.9 kg peamouth chub ( $5 \%$ ), and 0.05 kg redside shiner ( $<1 \%$ ).


Figure 22. Summary of the proportional catch, catch per unit effort (CPUE), and biomass by species of fish collected using gillnets, Coquitlam Reservoir, 15-17 May 2005.


Photo 4. Juvenile coarse fish from Coquitlam Reservoir - Northern pikeminnow (top), threespine stickleback (center), and peamouth chub (bottom).

## Gillnetting Fall Survey

Gillnetting was conducted at Coquitlam Reservoir during the nights of 31 October - 2 November (134-h effort). The total catch constituted 93 kokanee (43 \%), 52 peamouth chub ( $24 \%$ ), 30 Northern pikeminnow ( $14 \%$ ), 25 cutthroat trout ( $12 \%$ ), 12 sucker ( $6 \%$ ), 3 sculpin ( $1 \%$ ), and 1 coho ( $<1 \%$ ) (Figure 23, Appendix 23, Appendix 28).

A total CPUE (catch per hour standardized for a $90 \mathrm{~m}^{2}$ set area) of 4.04 fish was estimated for the fall gillnetting operation. Proportionally, CPUE was highest in the Central Basin (57\%), followed by the South Basin ( $28 \%$ ), and the North Basin ( $21 \%$ ). Total CPUE ranged from 0.04 for coho and sculpins to 2.02 for kokanee. Of the total CPUE for kokanee, 0.45 ( $22 \%$ ), 0.97 ( $48 \%$ ), and 0.60 ( $30 \%$ ) were taken in the North, Central, and South basins, respectively. Of the total 35.0 kg of fish caught, kokanee constituted $36 \%$, cutthroat trout $25 \%$, Northern pikeminnow $19 \%$, peamouth chub $13 \%$, sucker $7 \%$, and both coho and sculpin $<1 \%$.




Figure 23. Summary of proportional catch, CPUE, and biomass of fish species collected using gillnets, Coquitlam Reservoir, 31 October-2 November 2005.

## Minnow Trapping Fall Survey

Minnow trapping was conducted in Coquitlam Reservoir during the nights of 15-18 September for a total of 176 -h effort. A total of 112 fish were caught. The catch comprised 38 Northern pikeminnow ( $34 \%$ ), 30 redside shiner ( $27 \%$ ), 30 three-spine stickleback ( $27 \%$ ), 7 sculpin ( $6 \%$ ), 5 cutthroat trout, and 1 peamouth chub (1\%) (Appendix 22, Appendix 27). CPUE (standardized as number of fish caught per trap per hour) for the minnow trapping survey was 0.66 fish. The proportion of the total CPUE was highest in the North Basin (56\%), and intermediate in the South ( $23 \%$ ), and Central ( $21 \%$ ) basins. CPUE ranged from 0.02 for both sculpins and cutthroat trout to 0.38 for Northern pikeminnow. Of a total 0.45 kg of fish, Northern pikeminnow constituted $46 \%$, sculpin $16 \%$, redside shiner $15 \%$, cutthroat trout $14 \%$, three-spine stickleback $7 \%$, sucker $1 \%$, and peamouth chub $<1 \%$.

Minnow trapping was also conducted in the Coquitlam Reservoir during the nights of 31 October2 November for a total 99-h effort. Total catch consisted of 26 redside shiner ( $67 \%$ ) 9 Northern pikeminnow ( $23 \%$ ) and 4 sculpin ( $10 \%$ ) (Appendix 24, Appendix 29). Total CPUE (catch standardized by trap per hour) was 0.38 fish, with the highest proportion of fish captured in the Central Basin ( $40 \%$ ) followed by the South ( $32 \%$ ) and North ( $28 \%$ ) basins. Total CPUE ranged from 0.03 for sculpin to 0.21 for redside shiner with Northern pikeminnow comprising $55 \%$, redside shiner $33 \%$, and sculpin $12 \%$.

## Mid-water Trawl Fall Survey

Mid-water trawling was operated in Coquitlam Reservoir during the night on 24 November for a total 0.6-h effort, resulting in three juvenile kokanee captured, yielding a CPUE of 4.80 fish per trawl hour (Appendix 25).

## Adult Kokanee Spawner Distribution and Maturation

The Coquitlam Reservoir was test netted during 31 October- 2 November 2005 to obtain a better understanding of the distribution of kokanee spawners. Both sexually mature and spent kokanee were captured in gillnets in the South and Central basins of the reservoir. A total of 41 female and 52 male kokanee were caught. Of the total number of females caught, $70 \%$ were spawning, $10 \%$ were spent, and $20 \%$ were immature. Of the males caught, $80 \%$ were spawning, $10 \%$ were spent, and $10 \%$ were immature. Of 24 aged $2+$ and 26 age- $3+$ male kokanee, $83 \%$ and $96 \%$, respectively, were spawning in November 2005. For the same time period, of 11 age- $2+$ and 30 age- $3+$ female kokanee, $45 \%$ ) and $2893 \%$, respectively, were spawning. The sex ratio of mature kokanee in 2005 was 1:1.2.

It was suspected that kokanee were broadcast spawning over large boulders as the majority of spawners showed no evidence of caudal fraying, or abrasion marks on their bellies. Kokanee were most abundant in four likely beach spawning areas: Site I (Falls Creek), Site C, Site D (Harmony Creek), and Site E (Figure 24). Peak kokanee spawning was estimated to occur in the first two weeks of November.


Figure 24. Map of Coquitlam Reservoir showing fluvial fans and potential kokanee beach spawning sites (Bocking and Gaboury 2003).

During spawning surveys done on foot in the lower 500 m of the Upper Coquitlam River and Cedar, Harmony and Falls creeks between September and November, 2005, no kokanee were seen, nor was there any evidence of redds.

## Biological Characteristics of Coquitlam Kokanee

Of the 137 kokanee sampled in May, lengths (and weights) ranged between $128 \mathrm{~mm}(22 \mathrm{~g})$ and $225 \mathrm{~mm}(123 \mathrm{~g})$ (Photo 5, Appendix 30 and 32). The sex ratio among age- 2 and age- 3 kokanee was $0.8: 1.0$. Age-1 fish constituted $1 \%$, age- $243 \%$, and age- $356 \%$ (Figure 25). The dominant age class among female and male kokanee was age-3 and age- 2 , respectively. Mean length and weight for age-2 females and males, respectively, were 179.3 mm and 176.7 mm , and 58.2 g and 60.9 g .


Photo 5. An example of kokanee length measurement and scales collected, Coquitlam Reservoir, 2005.

The lengths (and weights) of 96 kokanee sampled in November ranged between $128 \mathrm{~mm}(27 \mathrm{~g})$ and $251 \mathrm{~mm}(182 \mathrm{~g})$ (Appendix 31, Appendix 34). The sex ratio among age-2 and age-3 kokanee was 1.0:1.2. Age-0 fish comprised $3 \%$, age- $12 \%$ ), age-2 ( $37 \%$ ) and age-3 $58 \%$ of the sample of fish that was aged (Figure 25). Age-3 was the dominant age for both male and female kokanee. Mean length (and weight) for age-0 and age- 1 kokanee were $52.7 \mathrm{~mm}(1.4 \mathrm{~g})$ and 136.5 mm ( 36.1 g ), respectively. Mean length and weight for age- 2 female and male kokanee, respectively, were 210.2 mm and 215.3 mm , and 116.4 g and 120.2 g .

Growth rates of kokanee collected in Coquitlam Reservoir were relatively high for age- 1 fish, but otherwise typical for age-2 and age-3. Specific growth rates based on Ricker's growth equation (1975) were $1.36 \% /$ day, $0.30 \% /$ day and $0.07 \% /$ day for $1-, 2-$ and $3-y r-o l d$ fish. Absolute growth rates were $0.11 \mathrm{~g} /$ day, $0.23 \mathrm{~g} /$ day and $0.09 \mathrm{~g} /$ day, respectively, for $1-$, 2- and $3-\mathrm{yr}$-old fish.



Figure 25. Length frequency and age distribution of kokanee in Coquitlam Reservoir, 15-17 May, 31 October-2 November, and 24 November, 2005.

## Diet of Kokanee, Cutthroat Trout, and Peamouth Chub

In both 2004 and 2005, the diet of kokanee consisted primarily of zooplankton. In autumn 2004, the dominant food comprised larger prey such as Daphnia. In 2005, the analysis of stomach contents was done in two time periods, May and November (Appendix 36). In May, kokanee consumed primarily cyclopoid copepods, with small numbers of other zooplankton. In September, the dominant prey consumed was the cladoceran Holopedium, with variable amounts of other zooplankton such as copepods, Daphnia, Leptodora and Polyphemus. Kokanee did not appear to feed selectively, but rather fed on whatever taxa of zooplankton was dominant.

Cutthroat trout (Appendix 37) appeared to be generalist feeders, eating substantial numbers of several zooplankton taxa, as well as aquatic and terrestrial insects.

The diets of several species of coarse fish were also examined. From an analysis of their gut contents, it was found that in May, peamouth chub fed on a mixture of zooplankton (Bosmina) and aquatic insects, Northern pikeminnow and redside shiner consumed almost entirely Bosmina, and longnose sucker consumed a mixture of cyclopoid copepods and aquatic insects. For the November samples, the only species for which dietary data are complete is peamouth chub, for which the main prey consisted of Daphnia, Daphnia ephippia and Holopedium.

## Coquitlam Reservoir Fish Food Web

The diets of fish in Coquitlam Reservoir were also examined indirectly by examining the compositions of stable isotopes of nitrogen and carbon in the muscle tissue of fish. The compositions of fish carbon $\left(\delta^{13} \mathrm{C}\right)$ and nitrogen $\left(\delta^{15} \mathrm{~N}\right)$ were used to assess the structure of the fish food web. The carbon stable isotope signature gives an indication of whether the food (e.g., zooplankton, aquatic insects, other fish) is from the open-water part of the ecosystem or of watershed origin and originating in the nearshore littoral, shallow-water areas (e.g., terrestrial insects). For carbon isotopes, the enrichment during a trophic step is low ( $0-1 \%$ ) and usually regarded as negligible. Hence, the consumer's $\delta^{13} \mathrm{C}$ is close to that of its carbon source. In lakes, terrestrial carbon sources usually display higher $\delta^{13} \mathrm{C}$ values than aquatic carbon sources. Fish $\delta^{13} \mathrm{C}$ values provide information on which of these carbon sources support the bulk of fish production. On the other hand, the nitrogen stable isotope signature provides information on where in the food chain a fish species belongs and what the dominant source of food might be. For example, a fish that eats other fish would have a higher nitrogen stable isotope signature than a fish which only eats zooplankton. The consumer's $\delta^{15} \mathrm{~N}$ is usually $3-4 \%$ higher than that of its prey, resulting in a ${ }^{15} \mathrm{~N}$-enrichment for species that occupy higher trophic positions within the food web.

The $\delta^{15} \mathrm{~N}$ data of the Coquitlam fish suggest a two-level food web, with cutthroat trout and sculpins as top predators. Intra-species variability in $\delta^{13} \mathrm{C}$ signature is generally high in this system and not related to the sample lipid content. The high variability suggests that in this food web there are no clearly pelagic versus littoral/terrestrial food chains but rather an intricate and interconnected food web, with a gradient in the use of pelagic and littoral/terrestrial carbon sources by fish (Figure 26).

There are no noticeable differences in the food web configuration between 2004 and 2005 (Figure 26). Carbon isotope values $\left(\delta^{13} \mathrm{C}\right)$ values of $\mathrm{CT}, \mathrm{NSC}, \mathrm{PCC}$, RSS and SC are centered around $27 \%$ suggesting a major use of terrestrial or littoral carbon sources. CT showed a significant decrease in $\delta^{13} \mathrm{C}$ and a significant increase in ${ }^{15} \mathrm{~N}$ between 2004 and 2005 , but no changes in $\delta^{13} \mathrm{C}$ variability, which is consistent with the significant increase in the proportion of pelagic carbon on which CT rely. NSC and PCC did not exhibit significant changes in their $\delta^{13} \mathrm{C}$ and $\delta^{15} \mathrm{~N}$ values, but a significant decrease in their $\delta^{13} \mathrm{C}$ variability suggests more specific foraging on terrestrial and littoral carbon sources for these species.

Only kokanee and the threespine stickleback foraged in the pelagic habitat and obtained their carbon essentially from pelagic sources. Kokanee sampled in 2004 and 2005 showed a significant decrease in $\delta^{13} \mathrm{C}$ variability in 2004, consistent with more specific carbon foraging on pelagic sources (Figure 27).


Figure 26. Comparison of bi-plot diagrams of the C - and N -isotope compositions of fish species and of their potential food sources in 2004 and 2005. Species codes are: $\mathrm{KO}=$ kokanee, $\mathrm{CT}=$ cutthroat trout, $\mathrm{PCC}=$ peamouth chub, NSC = Northern pikeminnow, LSU = largescale sucker, RSS = redside shiner, $\mathrm{SC}=$ sculpin, TSS $=$ threespine stickleback, $\mathrm{AQ}=$ aquatic prey, $\mathrm{TERR}=$ terrestrial prey.

## Fish Standing Stock

On 16 May 2005, the total fish population in Coquitlam Reservoir was estimated at 404,177 fish ( 348,136 to 460,218 fish at $95 \%$ CI) (Table 2). The total fish population within each of the basins was estimated at 31,281 fish ( $\pm 51 \%$ at $95 \%$ CI) for the North Basin, 275,323 fish ( $\pm 17 \%$ at $95 \%$ CI) for the Central Basin, and 97,573 fish ( $\pm 33 \%$ at $95 \%$ CI) for the South Basin (Appendix 38). Total fish biomass in the reservoir was approximately $18,014 \mathrm{~kg}$ (range 15,516 to $20,511 \mathrm{~kg}$ ) (Appendix 40). Fish production in the reservoir ranged between 290 fish $/ \mathrm{ha} \mathrm{( } 13 \mathrm{~kg} / \mathrm{ha}$ ) and 384 fish/ha ( $17 \mathrm{~kg} / \mathrm{ha}$ ) in 2005. An estimated $39 \%$ of the fish population in the reservoir was kokanee ( 156,912 kokanee; 95 CI of 92,136 to 221,688 ). Total biomass of kokanee was $7,266 \mathrm{~kg}$ (range $4,846-13,346 \mathrm{~kg})$.

On 1 November 2005, the total fish population in Coquitlam Reservoir was estimated at 194,604 ( 115,560 to 273,648 at $95 \% \mathrm{CI}$ ). An estimated 13,893 fish were in the North Basin, 95,213 fish $( \pm 40 \%$ at $95 \% \mathrm{CI})$ in the Central Basin, and 85,498 fish ( $\pm 87 \%$ at $95 \% \mathrm{CI}$ ) in the South Basins (Appendix 39). Total fish biomass was approximately $42,260 \mathrm{~kg}$ (range 25,095 to $59,425 \mathrm{~kg}$ ) (Appendix 41).


Figure 27. Frequency distributions of individual $\delta^{13} \mathrm{C}$-values for cutthroat trout, kokanee, northern pikeminnow, and peamouth chub in Coquitlam Reservoir.

Table 2. Summary of fish population characteristics, Coquitlam Reservoir, 2004-2005.

|  | Sep-04 | May-05 | Nov-05 |
| :--- | ---: | ---: | ---: |
| Fish Population Estimate | 648,420 | 404,177 | 194,604 |
| $\quad$ Lower 95 \% CI | 564,678 | 348,136 | 115,560 |
| $\quad$ Upper 95 \% CI | 732,162 | 460,218 | 273,648 |
| Mean acoustic density (\# / ha) | 698 | 382 | 200 |
| Fish standing crop (kg/ha) | 31.3 | 15.0 | 35.2 |
| Mean length (mm) | 68 | 179 | 223 |
| Vertical distribution (m) | $0-30$ | $0-30$ | $0-35$ |
| Percent Kokanee (\%) | 39 | 39 | 37 |

Fish production in the reservoir ranged between 96 fish $/ \mathrm{ha} \mathrm{( } 21 \mathrm{~kg} / \mathrm{ha}$ ) and 228 fish $/ \mathrm{ha}(50 \mathrm{~kg} / \mathrm{ha})$ in 2005. Kokanee comprised $37 \%$ of the total fish population in the reservoir ( 71,159 kokanee; 95 CI of 27,429 to 197,645 ). Total kokanee biomass was 3212 kg (range 1667 to $14,643 \mathrm{~kg}$ ).

## Kokanee Standing Stock

Of the estimated 156,912 kokanee present in May, $45 \%$ were age- $0+$, $26 \%$ were age- $1+$ and $29 \%$ were age- 2 and older (Table 3). The estimated biomass of age- 0 kokanee was 99 kg , while age- 1 and age-2 and older comprised 981 kg and 6187 kg , respectively (Appendix 42). Production estimates were as follows: age- $0=59$ fish $/ \mathrm{ha}(0.1 \mathrm{~kg} / \mathrm{ha})$; age-1 = 34 fish $/ \mathrm{ha}(0.8 \mathrm{~kg} / \mathrm{ha})$; and age- 2 and older $=38$ fish $/ \mathrm{ha}$ ( $5.1 \mathrm{~kg} / \mathrm{ha}$ ) (Appendix 42).

Of the estimated 72,159 kokanee population in November, $34 \%$ were age- $0+, 42 \%$ were age- $1+$, and $24 \%$ were age- 2 and older (Table 3). Biomass of age- 0 , age- 1 , and age- 2 and older fish was $34 \mathrm{~kg}, 735 \mathrm{~kg}$, and 2352 kg , respectively (Appendix 43). Production estimates were as follows: age- $0=20 \mathrm{fish} / \mathrm{ha}(<0.1 \mathrm{~kg} / \mathrm{ha})$; age- $1=25 \mathrm{fish} / \mathrm{ha}(0.6 \mathrm{~kg} / \mathrm{ha})$; and age- 2 and older $=14 \mathrm{fish} / \mathrm{ha}$ $(2.0 \mathrm{~kg} / \mathrm{ha})$ (Appendix 43). A least squares regression of weight by age was used to compute mean fish weight per age class: age- $0+=1.4 \mathrm{~g}$; age- $1+=24 \mathrm{~g}$; and age- $2+$ and older $=135 \mathrm{~g}$.

Table 3. Kokanee population parameters, Coquitlam Reservoir, 2004-2005.

|  | Sep-04 |  |  |  | May-05 |  |  |  | Nov-05 |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Parameter ${ }^{\text {a }}$ | Age-0+ | Age-1+ | Age-2+ Age-3+ | Total | Age-0+ | Age-1+ | Age-2+ Age-3+ | Total | Age-0+ | Age-1+ | Age-2+ Age-3+ | Total |
| Population Estimate | 106,357 | 117,969 | 34,067 | 258,393 | 70,388 | 40,694 | 45,830 | 156,912 | 24,221 | 30,514 | 17,424 | 72,159 |
| Lower 95 \% CI | 59,480 | 69,749 | 4,133 | 133,362 | 33,198 | 3,752 | 7,776 | 44,726 | 4,538 | 6,897 | 11,073 | 22,508 |
| Upper 95 \% CI | 153,234 | 166,188 | 84,450 | 403,872 | 107,578 | 77,636 | 83,884 | 269,098 | 91,084 | 109,950 | 87,895 | 288,929 |
| Mean densities (\#/ha) | 88 | 98 | 28 | 214 | 59 | 34 | 38 | 131 | 20 | 25 | 14 | 59 |
| Standing crop (kg/ha) | 0.2 | 2.4 | 3.8 | 6.4 | 0.1 | 0.8 | 5.1 | 6.0 | 0.03 | 0.6 | 2.0 | 2.6 |
| Length / weight (n) | - | 5 | $38 \quad 2$ | 45 | - | 1 | 5977 | 137 | 3 | 2 | $35 \quad 56$ | 96 |
| Mean length (mm) | - | 158.6 | $224.5 \quad 254.5$ | 218.5 | - | 128 | 175.1208 .4 | 193.5 | 52.7 | 136.5 | $213.7 \quad 233.7$ | 218.7 |
| SD length (mm) | - | 5.5 | $18.0 \quad 17.7$ | 27.9 | - | - | $7.9 \quad 9.9$ | 19.6 | 5.5 | 12.0 | $14.9 \quad 8.5$ | 35.8 |
| Mean weight (g) | - | 47.0 | $136.0 \quad 177.0$ | 128.0 | - | - | 59.397 .9 | 80.6 | 1.4 | 36.1 | $119.0 \quad 149.7$ | 131.5 |
| SD weight (g) | - | 3.7 | $28.8 \quad 32.5$ | 40.4 | - | - | 7.414 .5 | 23.1 | 0.5 | 13.0 | 26.418 .6 | 37.8 |

${ }^{\text {a }} \quad$ Age- $2+$ and Age-3+ estimates pooled for population, mean densities, and standing crop.

Coquitlam kokanee exhibit fairly constant circuli patterns for the entire growing period up to age-3 with very weak winter checks (Carol Lidstone, Birkenhead Scales, Lone Butte, BC, pers. comm.). This pattern seems to be consistent between different brood years so that we can rule out effects of environmental variation. The increased juvenile growth and early spawning traits of Coquitlam kokanee (i.e., younger age at sexual maturity) appear to be similar to those found in other studies of kokanee (Alm 1959 and Nilsson 1990).

## Rearing Capacity and Potential Sockeye Production

Estimates of the rearing capacity of juvenile sockeye were updated from the original estimates of Bocking and Gaboury (2003) using empirical estimates of Euphotic Zone Depth (EZD). Bocking and Gaboury (2003) used an estimated EZD of 15.6 m in their original model estimates of 15,000 kg smolt biomass (EV Model) and 13,000 kg smolt biomass (PR model). Note that in Bocking and Gaboury (2003) photosynthetic rate was approximated since there are no empirical estimates of photosynthetic rate for Coquitlam Reservoir.

The three year (2003-05) average EZD for Coquitlam Reservoir was determined as 12 m based on light penetration. The re-calculated smolt biomass estimate at capacity for Coquitlam Reservoir
was then $11,500 \mathrm{~kg}$ using the EV Model (Table 4). No estimate is presented using the PR model as there are still no empirical data for photosynthetic rate in Coquitlam Reservoir. A request to conduct radio isotope analysis to determine photosynthetic rate was not approved by the GVRD.

It is likely that the EV Model over-predicts sockeye smolt biomass for the ultra-oligotrophic Coquitlam Reservoir. Hyatt (pers. comm.) suggests that non-glacial, coastal oligotrophic lakes could produce between 1.5 and $5.0 \mathrm{~kg} / \mathrm{ha}$ of sockeye fry biomass in the fall with typical values between 2.0 and $3.0 \mathrm{~kg} / \mathrm{ha}$. Using this Oligotrophic Lake Productivity Model we estimated that Coquitlam could produce between 1800 and $6000 \mathrm{~kg} / \mathrm{ha}$ of fall sockeye fry (Table 4).

Total sockeye smolt biomass and seasonal mean macrozooplankton biomass for the growing season (May - Oct) preceding smolting have been shown to be highly correlated for sockeye lakes ( $\mathrm{r}^{2}=0.92$; Koenings and Kyle 1997). Koenings and Kyle (1997) determined that, for 18 BC and Alaskan sockeye lakes, sockeye smolt biomass $\left(\mathrm{kg} / \mathrm{km}^{2}\right)$ was 2.11 times seasonal mean macrozooplankton biomass ( $\mathrm{mg} / \mathrm{m}^{2}$ ). Applying this relationship (Zooplankton Biomass Model) to Coquitlam zooplankton data for 2004 and 2005, it is possible to predict total sockeye biomass at smolting. Estimated sockeye smolt biomass using the Zooplankton Biomass Model for 2005 and 2006 smolt years were 1,853 and 1,129 , respectively (Table 4).

These three model estimates of sockeye smolt biomass were compared with actual Fall Fry Standing Crop of 0+ and $1+$ kokanee as measured using hydroacoustics and accounting for $30 \%$ winter mortality as suggested by Koenings and Kyle (1997). To relate these actual biomass estimates to sockeye we assumed that the forage area that would be available to sockeye is currently occupied primarily by $0+$ and $1+$ kokanee. The same $30 \%$ winter mortality assumption was applied to the model estimates to estimate Fall Standing Crop of sockeye.

Each estimate of smolt biomass was converted to numbers of smolts by dividing by the mean weight of smolts. Although there are no data available for what the average weight of a Coquitlam sockeye smolt might be, a suitable proxy is the average weight of kokanee $1+$ 'smolts' that emigrated from the reservoir in 2005. Unfortunately weights were not taken from the thousand kokanee measured for length in 2006 (Alf Leake, BC Hydro, pers. comm.). However, Bussanich et al. (2005) determined a length-weight relationship for Coquitlam kokanee and this can be used to derive weights for kokanee that emigrated as $1+$ fish in 2005. Assuming a mean length of 80 mm (actual mean not available at this time but mean length of Alouette $1+$ kokanee emigrants in 2005 was also 80 mm (Baxter and Bocking 2006), the mean weight of emigrating 1+ kokanee at Coquitlam in 2005 would have been estimated at 7.6 g . In contrast, the mean weight of Alouette $1+$ kokanee is 4.4 g .

Until such time as empirical data for the weight of $1+$ kokanee emigrating in 2006 from Coquitlam Reservoir are available, we recommend using a mean weight of 4.5 g per kokanee smolt and 4.0 g as an approximate weight for fall fry. The 7.6 g estimate from the weight-length relationship seems on the high side.

The required number of spawners to produce the estimated number of sockeye smolts was determined using an egg-to-smolt survival of $3 \%$, a fecundity of 2500 and a sex ratio of 1:1. Excluding the EV Model, estimates of the required number of spawners ranged from 3,136 (Observed Fall Standing Crop for 2004) to 24,889 (High Oligotrophic Lake Productivity).

Table 4. Comparison of different model estimates of smolt biomass, smolt numbers and required spawners for Coquitlam Reservoir.

| Method | Year | Euphotic Zone Depth (m) | Lake Area ( $\mathrm{km}^{2}$ ) | EV Units ${ }^{2}$ | Seasonal Mean Zooplankton Biomass ( $\mathrm{mg} / \mathrm{m}^{2}$ ) | Fall Standing $\text { Crop (kg) }{ }^{3}$ | Estimated Smolt Biomass $(\mathrm{kg})^{4}$ | Estimated <br> Smolt <br> Number | Estimated Eggs Required ${ }^{6}$ | Estimated <br> Spawners <br> Required ${ }^{7}$ | $\begin{array}{r} \text { Potential Re } \\ \text { Spawl } \\ \hline \end{array}$ | turn per |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| EV (Koenings and Burkett 1987) |  | 12 | 12 | 144 |  | 16,522 | 11,566 | 2,570,151 | 85,671,704 | 68,537 | $\begin{aligned} & \hline 1: 1 \\ & 68,537 \end{aligned}$ | $\begin{gathered} 5: 1 \\ 342,687 \end{gathered}$ |
| Oligotrophic Lake Productivity | Low <br> High |  | $\begin{aligned} & 12 \\ & 12 \end{aligned}$ |  |  | $\begin{aligned} & 1,800 \\ & 6,000 \end{aligned}$ | $\begin{aligned} & 1,260 \\ & 4,200 \end{aligned}$ | $\begin{array}{r} 280,000 \\ 933,333 \end{array}$ | $\begin{array}{r} 9,333,333 \\ 31,111,111 \end{array}$ | $\begin{array}{r} 7,467 \\ 24,889 \end{array}$ | $\begin{array}{r} 7,467 \\ 24,889 \end{array}$ | $\begin{array}{r} 37,333 \\ 124,444 \end{array}$ |
| Zooplankton Biomass (Koenings and Kyle 1997) | $\begin{aligned} & 2004 \\ & 2005 \end{aligned}$ |  | $\begin{aligned} & 12 \\ & 12 \end{aligned}$ |  | $\begin{aligned} & 73.17 \\ & 44.57 \end{aligned}$ | $\begin{aligned} & 2,647 \\ & 1,612 \end{aligned}$ | $\begin{aligned} & 1,853 \\ & 1,129 \end{aligned}$ | $\begin{aligned} & 411,703 \\ & 250,781 \end{aligned}$ | $\begin{array}{r} 13,723,440 \\ 8,359,351 \end{array}$ | $\begin{array}{r} 10,979 \\ 6,687 \end{array}$ | $\begin{array}{r} 10,979 \\ 6,687 \end{array}$ | $\begin{aligned} & 54,894 \\ & 33,437 \end{aligned}$ |
| Observed Fall Standing Crop | $\begin{aligned} & 2004 \\ & 2005 \end{aligned}$ |  |  |  |  | $\begin{gathered} 3,120 \\ 756 \end{gathered}$ | $\begin{gathered} 2,184 \\ 529 \end{gathered}$ | $\begin{aligned} & 485,333 \\ & 117,600 \end{aligned}$ | $\begin{array}{r} 16,177,778 \\ 3,920,000 \end{array}$ | $\begin{array}{r} 12,942 \\ 3,136 \end{array}$ | $\begin{array}{r} 12,942 \\ 3,136 \end{array}$ | $\begin{aligned} & 64,711 \\ & 15,680 \end{aligned}$ |
| Mean of all estimates excluding EV model |  |  |  |  |  |  |  |  |  | 11,017 | 11,017 | 55,083 |

## ${ }^{1}$ pers. comm. Kim Hyatt

${ }^{2}$ EV Units $=$ Euphotic Zone Depth $x$ Lake Area
${ }^{3}$ Estimated fall standing crop using EV Model is calculated as: fall standing crop $=1.3 \times$ Smolt Biomass
Oligotrophic Lake Productivity Model is calculated as: low productive $=1.5 \times$ Lake Area; high productive $=5 \times$ Lake Area
Zooplankton Biomass Model is calculated as: Fall Standing Crop $=1.3 \times$ Smolt Biomass
Observed Fall Standing Crop is combined $0+$ and $1+$ kokanee biomass estimated from hydroacoustics
${ }^{4}$ Estimated smolt biomass using EV Model is calculated as: Smolt Biomass $=-130+81.22 \times$ EVunits
Oligotrophic Lake Productivity Model is calculated as: Smolt Biomass $=0.7 \times$ Fall Standing Crop
Zooplankton Biomass Model is calculated as: Smolt Biomass $=2.11 \mathrm{x}$ Zooplankton Biomass x Lake Area
Observed Fall Standing Crop is calculated as: Smolt Biomass $=0.7 \times$ Fall Standing Crop
${ }^{5}$ Assumes 4.5 g smolts
${ }^{6}$ Assumes 3\% egg to smolt survival
${ }^{7}$ Assumes female fecundity of 2500 and $1: 1$ sex ratio

## DISCUSSION

## Limnological Assessment

Based on two years of monitoring data, plus previously gathered data, Coquitlam Reservoir displays limnological and water quality characteristics typical of oligotrophic (unproductive) lakes (Wetzel 2001). The 2005 data agree well with previous data from 2004 (Field et al. 2005) as well as 2000 and 2001 (Basu 2001), which showed very low concentrations of phosphorus ( P ) and nitrogen ( N ), and low phytoplankton (algal) biomass, indicative of the unproductive state of Coquitlam Reservoir. An initial review done by Stockner (2003) also concluded that Coquitlam was very unproductive. In addition, our data support sediment-coring (paleolimnological) data which show that the reservoir has been an unproductive lake since at least before construction of the dam in 1905 (Nordin \& Mazumder 2005).

Phosphorus (P) and nitrogen (N) have been shown to be significant factors limiting phytoplankton (algal) biomass and water clarity in surface waters (Schindler et al. 1971; Dillon and Rigler 1974). Low nitrogen and especially phosphorus concentrations likely play a major role in the low algal biomass and good water clarity observed in Coquitlam Reservoir. The reservoir has a circumneutral pH . It had low concentrations of total organic carbon (TOC), which in combination with low algal biomass and good water clarity, indicate good water quality for drinking water (low colour, few organic and inorganic particulates). Good quality source water is crucial as it generally requires lower treatment intensities than poor source water, and results in healthier, better-tasting finished drinking water with fewer toxic disinfection by-products (DBP's) (Davies and Mazumder 2003; Davies et al. 2004).

The 2005 Secchi and turbidity data showed some interesting trends. It appears that winter water clarity can be quite poor, depending on winter runoff conditions, with the January and February 2005 values showing poorer clarity than summer, when phytoplankton growth should be highest. This pattern would imply that the winter runoff and inorganic particulates may have a larger effect on water clarity than summer organic biological particulates (phytoplankton growth).

Turbidity is an important indicator of water quality, especially for drinking water supplies. High levels of turbidity can protect bacteria and viruses from disinfection, stimulate the growth of bacteria, and result in the need for an increased dosage of disinfectants (World Health Organization 2004; Health Canada 2003). Higher dosages of disinfectants generally result in higher levels of toxic disinfection by-products (DBPs), such as Trihalomethanes (THMs) (Health Canada 2003). Both inorganic and organic suspended particles can impart taste and odor problems, and some biological organisms produce toxic substances (e.g. species of the blue-green algae Microcystis produce the liver toxin, microcystin-LR) (Health Canada 2003).

For the benefit of water managers, guidelines for turbidity levels have been set by several governing bodies, including the Province of British Columbia, the Government of Canada and the World Health Organization. As of this writing, the British Columbia provincial guideline for induced turbidity is 1.0 NTU in reservoirs such as Coquitlam (i.e. waters of exceptional clarity where background turbidity is less than or equal to 5.0 NTU) (Singleton 2001). Although turbidity in Coquitlam is generally below this 1.0 NTU guideline, our data show that winter values can occasionally exceed this threshold. In a recent study of the water distribution systems of the

GVRD, Aramini et al. (2000) showed that the probability of gastroenteritis increased as turbidity increased. According to Health Canada (2003), turbidity levels as low as 0.83 NTU may result in increased microorganism growth, and increases in turbidity of only 0.2 to 0.3 NTU are associated with increased concentrations of Giardia cysts. During periods where turbidity exceeds the 1.0 NTU guidelines, more extensive monitoring of water quality may be necessary.

Although overall algal biomass in Coquitlam Reservoir is low, a substantial portion of the algal community consists of cyanobacteria (blue-green algae). Several types of blue-greens that were observed at low biomass in Coquitlam Reservoir (e.g., Microcystis, Anabaena) have the potential to cause problems of taste and odour, and/or toxicity at higher biomass (Davies and Mazumder 2003). The low water temperatures and nutrient concentrations normally observed in the reservoir make the likelihood of a large bloom of blue-greens unlikely. However, continued regular monitoring of key nutrients and algae are indicated, because an increase in nutrients has the potential to trigger a shift in the algal community to greater biomass of obnoxious species such as blue-greens (Reynolds, 1984; Carmichael, 2001). Should blooms of blue-greens or other noxious algae be observed in future, sampling for taste and odour compounds may be warranted.

The conclusion that can be drawn from the fish stomach content and stable isotope results is that fish productivity (at least for the species that were sampled and assuming they represent all of the fish community) relies essentially on littoral/terrestrial carbon sources. This pattern is consistent between summer (data from September 2004) and winter/spring (May 2005). Only some species (kokanee being the dominant one) rely indirectly on phytoplankton production. This configuration of the Coquitlam Reservoir food web is consistent with its low trophic status, and scarcity of the pelagic resource. We know that fish in Coquitlam Reservoir are rearing-limited, and if yearling sockeye are introduced, it is likely that the kokanee population will decline due to direct competition for the same prey and selectivity for similar prey sizes.

## Fish Abundance

Inter-annual variation in fish abundance in Coquitlam Reservoir was apparent between 2004 and 2005 (Table 3, Bussanich et al. 2005). From acoustic surveys, the total fish population was estimated at 648,000 (565,000-732,000; 95\% CI), 404,177 (348, 136 to 460,218; $95 \%$ CI) and 194,604 (115,560 to 273,648; 95 \% CI) in September, 2004, May 2005 and November 2005, respectively. Mean densities in fall 2005 (200 fish/ha) were lower than fall 2004 ( 698 fish $/ \mathrm{ha}$ ). Whole lake fish abundance in fall was lower in $2005(194,604)$ than in $2004(648,420)$, while whole lake biomass in fall was higher in $2005(42,260 \mathrm{~kg})$ than in $2004(37,700 \mathrm{~kg})$. A higher proportion of larger fish ( $>200 \mathrm{~mm}$ ) was detected in 2005. Kokanee juvenile abundance was significantly lower in fall 2005 (25,000 (Age-0+), 30,000 (Age-1+)) than in fall 2004 (106,000 (Age-0+), 118,000 (Age-1+). The average size of fish targets in fall was larger in 2005 ( 79 mm ) compared to $2004(68 \mathrm{~mm})(2004)$. Fall fish population estimates were 3 times greater in 2004 than in 2005.

Population estimates may vary depending on fish distribution, behaviour, background acoustic noise levels, and other environmental factors (MacLennan and Simmonds 1992, Freon et al. 1993, Mitson 1993, McAllister 1998). We were unable to apportion age-0 and age-1 kokanee population estimates from other fish species, therefore, our kokanee population estimates for these age classes are likely to be over-estimated.

Representative samples from the pelagic fish community were needed to identify the species present and adjust the hydroacoustic population estimate for the different depth strata. Only age-0 kokanee were captured using mid-water trawling, however, larger resident fish were intermingled with the smaller kokanee in the surface to 30 m depth range. In addition to collecting age, size and sex data on populations, errors associated with species distribution and composition can be reduced by intensive mid-water trawling and multi-panel, floating gillnets (mesh sizes ranging from 3 mm to 91 mm ). As estimates of fish populations based on trawling and gillnetting are influenced by several variables, it is important to quantify and explain as much of the variation as possible (He 1993, Walsh and Hickey 1993, Wardle 1993).

Additional years of fish population assessment are required to more rigorously examine annual variability as occurred for example between 2004 and 2005. Factors potentially affecting fish population estimates include:

- sampling error between years;
- fish behavioural differences - e.g., adult kokanee distribution in September 2004 versus November 2005 - there was greater probability of acoustically detecting adult fjsh offshore in 2004 versus adults near-shore in 2005;
- reservoir operational differences that affected juvenile kokanee abundance (e.g. variability in reservoir drawdown);
- variation in the zooplankton community; and
- changes in predation rates between years.


## Kokanee Age Structure, Growth \& Survival

Egg-to-fry survival in 2005 was relatively low compared with other kokanee stocks in reservoirs (Fredricks et al. 1995). Bussanich et al. (2005) estimated that 6000 female spawners in 2004 deposited approximately 2.28 million eggs (assumed 380 eggs per female). In fall 2005, there was an estimated 24,200 age- $0+$ kokanee present, arising from $1.1 \%$ egg-to-fry survival kokanee eggto fry survival has ranged from $1.0 \%$ to $7.7 \%$ in other reservoirs (Fredricks et al. 1995).

Over-winter fry survival for kokanee in the Coquitlam Reservoir in 2005 was similar to that for kokanee stocks in other reservoirs (Fredricks et al. 1995). Bussanich et al. (2005) estimated that there were 106,400 age- $0+$ kokanee present in 2005. In the present study, the estimate of 40,700 age- $1+$ kokanee in spring 2005 indicates an over-winter fry survival of $38 \%$. The range in overwinter fry survival for kokanee has been considerable in other reservoirs (Fredricks et al. 1995).

## Kokanee Spawner Abundance, Distribution \& Potential Egg Deposition

Gaboury and Murray (2006) indicated that tributaries to Coquitlam Reservoir that are accessible to adfluvial salmon do not appear to provide a significant amount of suitable spawning habitat. Spawner surveys revealed that streams accessible to adfluvial fish, such as Cedar Creek, and the fans areas of Harmony and Falls creeks did not appear to be utilized by kokanee for spawning in 2005. Similarly, based on the quality and quantity of spawning habitat in the accessible tributaries, it is not expected that significant numbers of re-introduced sockeye would utilize these streams for spawning. The spawning period for kokanee likely began in mid-October, peaked in early-

November, and was completed by late-November. Recorded water temperatures in late October and early November ranged from $6^{\circ} \mathrm{C}$ to $12{ }^{\circ} \mathrm{C}$ over which kokanee are known to spawn.

Gaboury and Murray (2006) indicated that there is sufficient lake spawning habitat ( $1500 \mathrm{~m}^{2}$ ) below the 140 m elevation and in Cedar and Beaver creeks ( $1000 \mathrm{~m}^{2}$ ) to support a kokanee population of approximately 4500 females or a sockeye population of 1500 females. If the maximum reservoir drawdown is maintained at the 144 m elevation as proposed by Bocking and Gaboury (2003), this would potentially support a kokanee population of about 10,200 females or an equivalent sockeye population of roughly 3400 females. In addition, sediment core data suggest that the Coquitlam Reservoir never had a relatively large population of salmonids prior to construction of the dam in 1905 (Nordin \& Mazumder 2005).

There appears to currently be sufficient habitat in the reservoir basin to support between 4500 and 10,200 kokanee female spawners. Using the $1.2: 1$ sex ratio observed in 2005, the reservoir could support an adult kokanee population as large as $10,000-22,600$ based on available spawning habitat. The population of mature kokanee estimated to be present in the reservoir in November 2005 was 17,400 . In $2005,24 \%$ of the kokanee population was in their third or fourth year of life. The observed sex ratio in the age- 2 and age- 3 mature kokanee was 1.2 males to 1 female $(\mathrm{n}=91)$. Of the age- 2 and age- 3 observed, $70 \%$ of the age- 2 and $100 \%$ age- 3 kokanee would spawn in November 2005. The estimated total female spawners for 2005 would be approximately 6900 females ( 4400 to $34,80095 \% \mathrm{CI}$ ). The estimated total eggs deposited for 2005 would be approximately 2.62 million (based on 380 eggs per female observed in 2005).

## Lake Productivity \& Implications of Re-Introducing Sockeye

There have been concerns raised about the effects of the introduction of sockeye on the zooplankton community, mainly related to increased grazing pressure. An adequate understanding of the trophic structure is of key importance in managing the resource. Food chain pyramids are useful in understanding the inter-relationships of energy flow in lake systems (Carpenter et al. 1985, Persson 1999). The amount of phytoplankton produced is directly related to the amount of nutrients supplied to a lake (Sakamoto 1966; Pridmore and McBride 1984). There is also considerable documentation that zooplankton productivity is directly related to phytoplankton productivity (Rublee 1992, Canfield and Jones 1996). Moreover, there is evidence that the production of obligate planktivorous fish, like kokanee or juvenile sockeye is heavily dependent on zooplankton abundance (Baldwin et al 2000). The Coquitlam Reservoir with its simple food chain in the pelagic zone would result in that sockeye production would be influenced by zooplankton, phytoplankton and nutrient levels in the lake.

In considering the biological pyramid of the reservoir, it is essential to establish the phosphorus ( P ) characteristics of the lake, since biological productivity is directly related to P . Total P is estimated at 3132 kg , using a typical concentration of about $3 \mathrm{ug} / \mathrm{L} \mathrm{TP}$ and a lake volume of $1044 \times 10^{6} \mathrm{~m}^{3}$. Preliminary estimates of P load (range between 4.97 tonnes/year and 6.64 tonnes/year) used the Vollenweider (1968) model that relates loading to flushing rate and lake depth (Stockner 2003). Adjustments to lake TP concentration (from $1.5 \mathrm{ug} / \mathrm{L}$ to $3 \mathrm{ug} / \mathrm{L} \mathrm{TP}$ ) and the flushing rate (from 0.3 year to 1.4 years exchange time), resulted in a P loading estimate of 2.24 tonnes/year (or 0.187 $\mathrm{g} / \mathrm{m}^{2} / \mathrm{yr}$ ). Findings of the present study and others (Field et al. 2005) indicate that Coquitlam Reservoir in 2004 and 2005 was strongly P-limited.

Phosphorus input to the reservoir from marine-derived nutrients via sockeye returning from the sea need to be considered. Would sockeye at carrying capacity of the reservoir result in an appreciable rise in nutrient levels in the lake? Based on estimates of fish production, this does not seem likely. Assuming 3000 adult sockeye returned to Coquitlam and spawned, each weighing 2.7 kg and contributing approximately $0.5 \%$ phosphorus (Larkin and Slaney 1997, Mathieson et al 1988), the estimated P input would be $40 \mathrm{~kg} /$ year - this would be less than $5 \%$ of the annual load and well within the range of natural variation.

Reliable estimates of phytoplankton and zooplankton biomass are important to understanding the trophic structure in Coquitlam Reservoir. Phytoplankton biomass for the reservoir was estimated using the regression equation of Desortova (1981): Chl $a(\mathrm{ug} / \mathrm{L})=-1.69+6.38 \mathrm{~B}$, where B is $\mathrm{mg} / \mathrm{L}$ phytoplankton. The mean summer concentration of $0.4 \mathrm{ug} / \mathrm{L}$ in the top 16 m of the lake $(80 \mathrm{~kg}$ chlorophyll $a$ in the euphotic zone) would equate to about $330 \mathrm{ug} / \mathrm{L}$ wet weight phytoplankton biomass, and is within the 2005 estimated range ( 140 to $530 \mathrm{ug} / \mathrm{L}$ ). Phytoplankton chlorophyll content can vary widely, but typically comprises $0.5-5 \%$ of the biomass (Nichols and Dillon 1978, Lewis 1991). A total algal biomass production of $20,000 \mathrm{~kg}$ biomass in the epilimnion, or 66,000 kg in the euphotic zone, was estimated for the lake (assuming $330 \mathrm{ug} / \mathrm{L}$ average biomass and an epilimnetic volume of $60 \times 10^{6} \mathrm{~m}^{3}$ ).

A total 264 kg (dry weight) zooplankton biomass was estimated for the top 20 m of the lake $\left(220 \times 10^{6} \mathrm{~m}^{3}\right)$, based on a mean standing crop of $1.2 \mathrm{ug} / \mathrm{L}$ (dry weight) (approximate range 0.5 to $3.5 \mathrm{ug} / \mathrm{L}$ ). The biomass would double ( 528 kg ) if the top 40 m of the lake were used as the productive volume for zooplankton. The volumetric biomass estimate ( $1.2 \mathrm{ug} / \mathrm{L}$ ) is relatively low compared to west coast, oligotrophic lakes assessed for the Fisheries and Oceans Lake Enhancement Program. Simpson et al (1981) reported zooplankton standing crops of $6.7 \mathrm{mg} / \mathrm{m}^{3}$ ( $\mathrm{ug} / \mathrm{L}$ ) and $3 \mathrm{mg} / \mathrm{m}^{3}$ for Woss and Nimpkish lakes, respectively on Vancouver Island. Similarly, Stockner et al. (1980) reported zooplankton standing crops of $9 \mathrm{mg} / \mathrm{m}^{3}, 7.5 \mathrm{mg} / \mathrm{m}^{3}$, and $5.4 \mathrm{mg} / \mathrm{m}^{3}$ for Great Central, Henderson, and Kennedy lakes, respectively. These lakes produce two to five times greater zooplankton standing crops than Coquitlam.

## Rearing Capacity and Potential Sockeye Production

The question of estimating fish productivity in general has received considerable attention since the initial work of Ryder (1982) who proposed the Morpho-edaphic index (MEI) based on the relationship he found between fish yield and two lake characteristics: mean depth and the concentration of total dissolved solids. This question has been followed up on by many researchers and the general idea modified in a variety of ways using phosphorus concentration, primary production, temperature and other factors. Rearing capacity models for the management of BC, Washington, and Alaska sockeye stocks have been tested for over 20 years. Alaskan models rely on seasonal euphotic zone depth (EV model) (Koening and Kyle 1997), while BC models have modified the Alaskan model, and use photosynthetic rates (PR model) (Shortreed et al. 2000). However, as mentioned previously, none of these models may be particularly useful in determining sockeye rearing capacity in the ultra-oligotrophic Coquitlam Reservoir.

The reservoir is an unproductive lake ecosystem as is reflected in its low concentrations of nutrients, phytoplankton and zooplankton compared with other lakes in British Columbia, Washington, and Alaska (Figure 28, Koenings and Burkett 1987; Hume et al. 1996; Costella et al
1983). In eutrophic lakes, coldwater fishes may be subjected to a temperature-oxygen squeeze, where warm surface waters push fish to deeper water, while depleted oxygen in deep waters force fish to the surface. As an unproductive lake, Coquitlam Reservoir exhibits cool temperatures and well-oxygenated waters that are very favourable for coldwater sport-fishes, such as salmon, and the resident cutthroat trout and kokanee. As a result, the entire water column is likely available as usable habitat for these fishes.

Coquitlam Reservoir is considered a rearing (forage) limited system. According to Koenings and Burkett (1987), if approximately $85 \%$ of the out-migrant smolts were age-1 (and of threshold size), this would suggest density-dependent forage limitation in one growing season. If Coquitlam Reservoir is rearing-limited as a result of lack of forage, we need to examine whether densitydependence is regulating the kokanee population. We suspect that years with high kokanee stock abundance will be associated with low abundance of recruits, whereas years with low stock abundance will produce high numbers of recruits.


Figure 28. Relation between areal TP load at 1.4 year residence time and average Chlorophyll $a$ concentrations in Coquitlam Reservoir and other coastal BC lakes.

Table 4 illustrates that sockeye smolt biomass predicted from seasonal mean macrozooplankton biomass and from fall standing crop of $0+$ and $1+$ kokanee is substantially less than predicted from the EV Model. There are a number of plausible explanations for this difference:

1. The EV model over-estimates rearing capacity for sockeye in Coquitlam Reservoir,
2. Seasonal mean zooplankton biomass estimates from Site L2 underestimate the zooplankton biomass for the reservoir, and
3. The entire forage area available to all age classes of kokanee would be available for sockeye hence the fall standing crop prediction of sockeye smolt biomass using just $0+$ and $1+$ fall biomass is an underestimate of potential sockeye smolt biomass.

Unfortunately, there is no way to determine if explanation 1 (EV model overestimates) is valid since the model relies solely on EZD and lake area. The exact determination of photosynthetic rate might further refine the estimate of rearing capacity but this is unlikely given restrictions on the use of radio isotopes in this drinking water reservoir. As well, Shortreed et al. (2000) indicated that the PR model is more closely correlated in lakes and years when grazing pressure is minimal. In the presence of continuous high grazing pressure, a lake (like Coquitlam) may develop a predator-resistant, less productive zooplankton community even though PR remains the same.

With respect to zooplankton, when comparing the location of limnology Site L2 (Figure 2) with the location of kokanee in the lake during May and November of 2005 (Figure 19 and Figure 20) and September 2004 (Bussanich et al. 2005), it is evident that the kokanee are patchy in their distribution. In May of 2005 and September of 2004, they were mostly concentrated over the deepest and centre part of the reservoir. Therefore, it seems reasonable that the zooplankton data collected to date might underestimate the total lake zooplankton density. By how much will not be known unless future zooplankton sampling is conducted at more of the sample sites.

With respect to the forage area, total sockeye smolt biomass would increase from 2184 kg to 5376 kg in September of 2004 if the entire kokanee biomass were replaced by sockeye. The total sockeye smolt biomass using November 2005 estimates would increase from 529 kg to 2184 kg , but the validity of this is not certain.

As can be seen, modeling and predicting sockeye production from a lake system like Coquitlam can be difficult. Additional years of data for, at a minimum, zooplankton biomass and fall kokanee abundance would significantly increase understanding of annual variability in lake productivity and the interrelationship between zooplankton and kokanee. Until such time, it is known that given seasonal mean macrozooplankton biomass and fall standing crop of kokanee observed in 2004 and 2005, between 500 and 2100 kg of sockeye smolts might have been produced from those growing years. Based on a mean smolt size of 4.5 g , this translates into between 117,000 and 485,000 sockeye smolts that might have been produced in those two years.

Determining a target for the number of sockeye spawners, should re-introduction proceed, is even more problematic as two critical assumptions are introduced. These assumptions are; 1) egg-tosmolt survival and 2) sockeye fecundity. Sex ratio can likely be safely assumed to be $1: 1$. For the purpose of determining a potential spawner target based on the various model estimates of smolt production, we assumed an egg-to-smolt survival of $3.0 \%$ and a fecundity of 2500 eggs per female. Both these assumptions seem reasonable and within the ranges observed for other sockeye lakes.

## RECOMMENDATIONS

Any introduction of sockeye to the reservoir will require a pre-cautionary approach and a longterm limnological and fish monitoring program. The following specific recommendations are proposed.

## Limnological Assessment

1. Seasonal sampling is sufficient for the purpose of monitoring physical, chemical, phytoplankton and macrozooplankton trends (i.e. May - October). Of these zooplankton is the most critical.
2. From a drinking water quality perspective, investigate the factors which affect disinfection by-product generation from a relative risk approach (seasonality, precursors, temporal and spatial variation in turbidity).

## Fish Population Assessment

To improve our understanding of regulating factors of $O$. nerka populations in Coquitlam Reservoir:

1. Continue monitoring fall (November) fish abundance using the standard hydroacoustic and trawling techniques in 2005; and
2. Assess trends in population abundance relative to environmental correlates (i.e. reservoir elevations).

## Interim Production Target

Based on the above analysis, it seems reasonable to establish an interim sockeye production goal of 400,000 sockeye smolts (approximately 550,000 fall fry). At 4.0 g per fall fry, this number of sockeye would account for 2750 kg of fish biomass in the reservoir in the fall, on average each year. Assuming a fecundity of 2500 eggs per female sockeye and $3 \%$ survival from egg to smolt, approximately 5000 females or 10,000 total spawners would be required to produce 400,000 sockeye smolts. This suggested target is relatively consistent with the amount of lake spawning habitat that could be available for sockeye in the reservoir $50 \%$ of the time ( 144 m elevation) (Figure 29). If it is determined later that the reservoir can sustain a higher abundance of sockeye juveniles, then additional spawning habitat would need to be found, likely in the Upper Coquitlam River.

Pacific salmon populations go through periods of low and high productivity (Beamish et al. 1997). In terms of adult returns, at current low marine survival rates, returns per spawner for Coquitlam sockeye would likely be in the vicinity of $1: 1$ but could increase to higher returns per spawner (5:1) should marine survival improve. Returns per spawner between 3 and 10 were common for Fraser River sockeye in the 1950s through 1990s (DFO 1999). Although changes in freshwater conditions can affect stock productivity both within and among stocks, marine conditions appear to be the main driver (e.g. Beamish et al. 1997).

## Coquitlam Reservoir Elevation



Figure 29. Coquitlam reservoir elevations modeled for Water Use Plan alternative SY_STP6_PAsha. The median reservoir elevation for the 30 years modeled remained above 144 m during spawning and incubation.

## ACKNOWLEDGMENTS

The cooperation of many people was essential in meeting the objectives of this study. Special thanks to the Greater Vancouver Regional District, University of Victoria, Kwikwetlem First Nation, Chief and Elders, Fisheries and Oceans Canada, and to all whom assisted in the collection, processing, and analysis of the field data for this study. Technical assistance was provided by: Theresa Cotter, Crystal Lawrence, Kelly Young, Rob Newell, Sergei Verenitch, Shapna Mazumder, Blake Matthews and Shane Edmison of the University of Victoria. From GVRD, Dave Dunkley, Ken Juvik, Heidi Walsh, Scott Stuart, Malcolm Schulz, Brian De Gusseme and several Coquitlam Reservoir security personnel offered technical assistance in the field. Thanks to George Chaffee, Nancy Joe (Kwikwetlem First Nation), Howie Wright (Okanagan First Nation Alliance Fisheries), Don Degan (Aquacoustics Inc.), and Carol Lidstone (Birkenhead Scales) for their support. Joe Cherwatti (MOELP) provided assistance with obtaining the fish collection permit. Bruce Murray, Tom Skulstad, Robin Tamasi, Marc Gaboury, and Dorothy Baker (LGL) assisted with data collection, graphics and final reporting. Funding for this project was provided by the BC Hydro-Bridge Coastal Fish \& Wildlife Restoration Program, the National Science and Engineering Research Council of Canada, and the Greater Vancouver Regional District.

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## APPENDICES

Appendix 1. Temperature profile data at Coquitlam Lake Station 2 for 2005.

| Depth (m) | 27-Jan | 23-Feb | 31-Mar | 28-Apr | 5-Jun | 6-Jul | 27-Jul | 25-Aug | 22-Sep | 22-Oct | 5-Dec |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 5.5 | 4.8 | 6.0 | 13.0 | 15.2 | 17.6 | 20.5 | 21.9 | 17.3 | 12.4 | 6.3 |
| 2 | 5.5 | 4.8 | 6.0 | 13.0 | 15.0 | 17.6 | 20.4 | 21.7 | 17.2 | 12.3 | 6.2 |
| 3 | 5.3 | 4.8 | 6.0 | 11.4 | 14.8 | 17.6 | 20.3 | 21.6 | 17.1 | 12.1 | 6.2 |
| 4 | 5.2 | 4.8 | 6.0 | 10.8 | 14.7 | 17.2 | 19.9 | 21.6 | 17.1 | 12.0 | 6.2 |
| 5 | 5.2 | 4.7 | 5.9 | 8.7 | 14.4 | 15.4 | 19.0 | 21.5 | 17.1 | 11.9 | 6.2 |
| 6 | 5.1 | 4.7 | 5.9 | 8.0 | 13.3 | 14.1 | 17.2 | 21.3 | 17.0 | 11.6 | 6.2 |
| 7 | 5.0 | 4.7 | 5.9 | 6.7 | 12.6 | 13.3 | 15.3 | 17.9 | 17.0 | 11.4 | 6.2 |
| 8 | 5.0 | 4.7 | 5.7 | 6.2 | 10.5 | 12.8 | 13.8 | 14.7 | 13.7 | 11.2 | 6.2 |
| 9 | 5.0 | 4.7 | 5.7 | 6.1 | 9.2 | 11.7 | 13.2 | 11.4 | 11.7 | 11.0 | 6.2 |
| 10 | 5.0 | 4.7 | 5.5 | 5.9 | 8.1 | 10.5 | 10.4 | 10.3 | 9.6 | 10.7 | 6.2 |
| 11 | 5.0 | 4.7 | 5.4 | 5.8 | 7.4 | 10.1 | 8.7 | 9.7 | 8.7 | 10.4 | 6.2 |
| 12 | 5.0 | 4.7 | 5.4 | 5.7 | 6.8 | 9.9 | 8.0 | 7.7 | 7.7 | 10.3 | 6.2 |
| 13 | 4.9 | 4.7 | 5.3 | 5.7 | 6.6 | 8.2 | 7.6 | 7.2 | 7.0 | 10.1 | 6.2 |
| 14 | 4.9 | 4.7 | 5.3 | 5.7 | 6.3 | 7.2 | 7.0 | 6.4 | 6.5 | 9.9 | 6.2 |
| 15 | 4.9 | 4.7 | 5.3 | 5.7 | 6.0 | 6.7 | 6.7 | 6.3 | 6.3 | 9.8 | 6.2 |
| 16 | 4.9 | 4.7 | 5.2 | 5.6 | 5.9 | 6.5 | 6.5 | 6.2 | 6.2 | 9.3 | 6.2 |
| 17 | 4.9 | 4.7 | 5.2 | 5.6 | 5.8 | 6.3 | 6.4 | 6.1 | 6.0 | 8.5 | 6.2 |
| 18 | 4.8 | 4.7 | 5.2 | 5.6 | 5.6 | 6.1 | 5.9 | 6.1 | 5.9 | 8.2 | 6.2 |

Appendix 2. Dissolved oxygen profile data at Coquitlam Lake Station 2 for 2005.

| Depth <br> (m) | 27-Jan | 23-Feb | 31-Mar | 28-Apr | 5-Jun | 6-Jul | 27-Jul | 25-Aug | 22-Sep | 22-Oct | 5-Dec |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1 | 12.1 | 10.6 | 12.2 | 11.4 | 10.6 | 9.7 | 9.1 | 8.7 | 9.2 | 10.7 | 11.2 |
| 2 | 12.1 | 10.6 | 12.2 | 11.5 | 10.6 | 9.7 | 9.1 | 8.7 | 9.2 | 10.7 | 11.2 |
| 3 | 11.9 | 10.6 | 12.2 | 11.8 | 10.6 | 9.7 | 9.1 | 8.7 | 9.2 | 10.7 | 11.1 |
| 4 | 11.9 | 10.5 | 12.1 | 11.9 | 10.7 | 9.8 | 9.1 | 8.7 | 9.2 | 10.7 | 11.1 |
| 5 | 11.8 | 10.5 | 12.1 | 12.2 | 10.7 | 10.1 | 9.3 | 8.7 | 9.1 | 10.7 | 11.1 |
| 6 | 11.7 | 10.5 | 12.1 | 12.3 | 11.0 | 10.3 | 9.7 | 8.7 | 9.1 | 10.7 | 11.1 |
| 7 | 11.6 | 10.5 | 12.1 | 12.4 | 11.1 | 10.5 | 10.0 | 9.8 | 9.1 | 10.8 | 11.1 |
| 8 | 11.5 | 10.5 | 12.1 | 12.3 | 11.5 | 10.6 | 10.2 | 10.4 | 10.1 | 10.8 | 11.1 |
| 9 | 11.4 | 10.5 | 12.1 | 12.2 | 11.7 | 10.8 | 10.3 | 11.0 | 10.5 | 10.8 | 11.1 |
| 10 | 11.4 | 10.5 | 12.1 | 12.1 | 11.9 | 10.9 | 10.8 | 11.1 | 10.9 | 10.9 | 11.1 |
| 11 | 11.4 | 10.5 | 12.1 | 12.0 | 12.0 | 11.0 | 11.1 | 11.3 | 11.0 | 11.1 | 11.0 |
| 12 | 11.3 | 10.5 | 12.1 | 12.0 | 11.9 | 11.0 | 11.2 | 11.6 | 11.3 | 11.0 | 11.1 |
| 13 | 11.3 | 10.5 | 12.0 | 11.9 | 11.9 | 11.3 | 11.2 | 11.6 | 11.4 | 10.9 | 11.0 |
| 14 | 11.3 | 10.5 | 12.0 | 11.9 | 11.9 | 11.5 | 11.3 | 11.8 | 11.5 | 11.0 | 11.0 |
| 15 | 11.3 | 10.5 | 12.0 | 11.9 | 12.0 | 11.5 | 11.4 | 11.8 | 11.5 | 11.0 | 11.0 |
| 16 | 11.3 | 10.5 | 11.9 | 11.9 | 11.9 | 11.5 | 11.4 | 11.7 | 11.5 | 11.1 | 11.0 |
| 17 | 11.3 | 10.5 | 11.9 | 11.9 | 11.9 | 11.5 | 11.5 | 11.7 | 11.5 | 11.3 | 11.0 |
| 18 | 11.3 | 10.5 | 11.9 | 11.9 | 11.9 | 11.5 | 11.5 | 11.7 | 11.5 | 11.4 | 11.0 |

Appenidix 3. Secchi disc (water clarity) data at Coquitlam Reservoir 2005

| Date | Site 1 | Site 2 | Site 3 | Site 4 |
| :--- | ---: | ---: | ---: | ---: |
| 27-Jan | 4.3 | 4.3 | 2.5 | 3.0 |
| 23-Feb | 4.8 | 5.0 | 5.3 | 4.8 |
| 31-Mar | 5.0 | 5.5 | 4.8 | 6.8 |
| 28-Apr | 6.8 | 5.5 | 6.5 | 6.0 |
| 2-Jun | 5.5 | 6.5 | 6.0 | 5.5 |
| 6-Jul | 5.5 | 5.5 | 6.0 | 8.0 |
| 27-Jul | 6.8 | 7.0 | 9.0 | 9.8 |
| 25-Aug | 7.3 | 7.5 | 8.0 | 10.0 |
| 22-Sep | 5.3 | 6.3 | 7.0 | 8.5 |
| 20-Oct | 5.0 | 6.5 | 7.5 | 7.5 |
| 3-Dec | 6.0 | 7.3 | 8.0 | 7.0 |

Appendix 4. Site 2 pH data at Coquitlam Reservoir 2005.

| Date | Epilimnetic | Metalimnetic Hypolimnetic |  |
| :--- | ---: | ---: | ---: |
| 27-Jan | 8.56 | 7.29 | 7.13 |
| 23-Feb | 6.55 | 6.64 | 6.53 |
| 31-Mar | 7.63 | 7.45 | 7.37 |
| 28-Apr | 7.33 | 7.02 | 6.73 |
| 2-Jun | 6.71 | 6.46 | 6.21 |
| 6-Jul | 6.38 | 6.37 | 6.27 |
| 27-Jul | 6.59 | 6.31 | 6.54 |
| 25-Aug | 6.45 | 6.29 | 6.2 |
| 22-Sep | 6.4 | 6.07 | 6.22 |
| 20-Oct | 6.45 | 6.36 | 6.29 |
| 3-Dec | 6.27 | 6.28 | 6.23 |

Appendix 5. Site 3 total phosphorus data at Coquitlam Reservoir 2005.

| Date | Epilimnetic | Metalimnetic | Hypolimnetic |
| :--- | ---: | ---: | ---: |
| 27-Jan | 3.28 | 2.73 | 3.13 |
| 23-Feb | 1.76 | 0.89 | 1.81 |
| 31-Mar | 2.24 | 1.98 | 5.12 |
| 28-Apr | 4.33 | 3.93 | 2.67 |
| 2-Jun | 2.07 | 2.86 | 1.84 |
| 6-Jul | 1.14 | 0.90 | 1.47 |
| 27-Jul | 1.96 | 3.29 | 1.46 |
| 25-Aug | 2.96 | 3.61 | 2.66 |
| 22-Sep | 4.16 | 5.25 | 2.14 |
| 20-Oct | 2.47 | 3.13 | 2.80 |
| 3-Dec | - | - | - |

Appendix 6. TN data for site 2 at Coquitlam Reservoir 2005.

| Date | Epilimnetic | Metalimnetic | Hypolimnetic |
| :--- | ---: | ---: | ---: |
| 27-Jan | 163.54 | 161.22 | 182.84 |
| 23-Feb | 151.80 | 160.75 | 165.92 |
| 31-Mar | 162.30 | 150.52 | 143.00 |
| 28-Apr | 171.77 | 153.65 | 149.47 |
| 2-Jun | 128.70 | 142.71 | 140.03 |
| 6-Jul | 106.23 | 110.04 | 172.40 |
| 27-Jul | 109.77 | 114.25 | 119.49 |
| 25-Aug | 110.46 | 150.92 | 232.19 |
| 22-Sep | 98.37 | 115.37 | 124.09 |
| 20-Oct | 163.25 | 148.67 | 143.34 |
| 3-Dec | - | - | - |

Appendix 7. Nitrate-nitrite data for site 2 at Coquitlam Reservoir 2005.

| Date | Epilimnetic | Metalimnetic | Hypolimnetic |
| :--- | ---: | ---: | ---: |
| 27-Jan | 95.31 | 96.08 | 101.55 |
| 23-Feb | 85.52 | 85.71 | 83.39 |
| 31-Mar | 79.02 | 77.84 | 76.95 |
| 28-Apr | 84.75 | 98.37 | 97.20 |
| 2-Jun | 82.72 | 86.43 | 119.96 |
| 6-Jul | 67.84 | 66.39 | 104.77 |
| 27-Jul | 66.76 | 70.27 | 103.35 |
| 25-Aug | 53.40 | 69.76 | 107.14 |
| 22-Sep | 52.72 | 81.31 | 100.33 |
| 20-Oct | 108.49 | 96.76 | 104.96 |
| 3-Dec | - | - | - |

Appendix 8. Nitrogen to Phosphorus data for site 2 at Coquitlam Reservoir 2005.

| Date | Epilimnetic | Metalimnetic | Hypolimnetic |
| :--- | ---: | ---: | ---: |
| 27-Jan | 50 | 59 | 58 |
| 23-Feb | 86 | 182 | 92 |
| 31-Mar | 73 | 76 | 28 |
| 28-Apr | 40 | 39 | 56 |
| 2-Jun | 62 | 50 | 76 |
| 6-Jul | 93 | 122 | 117 |
| 27-Jul | 56 | 35 | 82 |
| 25-Aug | 37 | 42 | 87 |
| 22-Sep | 24 | 22 | 58 |
| 20-Oct | 66 | 47 | 51 |
| 3-Dec | - | - | - |

Appendix 9. TOC/DOC data for site 2 at Coquitlam Reservoir 2005

|  |  | TOC |  |  | DOC |  |  |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | ---: |
| Date | Epilimnetic Metalimnetic Hypolimnetic |  | 4 m |  | Epilimnetic Metalimnetic Hypolimnetic |  |  |
| 27-Jan | 2.3 | 2.2 | 2.3 | 2.3 | 2.1 | 2.0 | 2.1 |
| 23-Feb | 2.0 | 2.2 | 2.2 | 2.1 | 1.8 | 2.0 | 1.9 |
| 31-Mar | 1.9 | 1.8 | 1.9 | 1.9 | 1.9 | 1.8 | 1.8 |
| 28-Apr | - | 2.1 | 2.1 | - | - | 2.0 | 2.0 |
| 2-Jun | - | 1.8 | 1.6 | 1.6 | - | 1.6 | 1.6 |
| 6-Jul | 1.6 | 1.6 | 1.6 | 1.5 | 1.6 | 1.6 | 1.7 |
| 27-Jul | 1.7 | 2.0 | 1.7 | 1.7 | 1.7 | 1.8 | 1.7 |
| 25-Aug | 1.8 | 1.8 | 1.7 | 1.6 | 1.6 | 1.7 | 1.6 |
| 22-Sep | 1.7 | 1.7 | 1.7 | 1.7 | 1.6 | 1.6 | 1.6 |
| 20-Oct | 2.1 | 2.4 | 2.1 | 1.9 | 1.8 | 2.1 | 1.8 |
| 3-Dec | 2.0 | 1.9 | 2.0 | 2.0 | 1.8 | 1.8 | 1.8 |

Appendix 10. Chlorophyll data for site 2 at Coquitlam Reservoir 2005.

| Date | Epilimnetic | Metalimnetic Hypolimnetic |  |
| :--- | ---: | ---: | ---: |
| 27-Jan | 0.08 | 0 | 0 |
| 23-Feb | 0.06 | 0.23 | 0.46 |
| 31-Mar | nc | nc | nc |
| 28-Apr | 0.19 | 0.30 | 0.09 |
| 2-Jun | 0.89 | 0.54 | 0.16 |
| 6-Jul | 0.27 | 0.59 | 0.02 |
| 27-Jul | 0.52 | 0.15 | 0.46 |
| 25-Aug | 0.73 | 1.27 | 0.81 |
| 22-Sep | 0.05 | 1.44 | 0.74 |
| 20-Oct | 0.75 | 0.48 | 0.35 |
| 3-Dec | 0.27 | 0.63 | 0.61 |

Appendix 11. Summary of hydroacoustic sampling at Coquitlam Lake Reservoir in May 16-17, 2005.

| Route No. | Transect ${ }^{\text {a }}$ <br> No. | Date dd-mm | Start hh:mm | End hhimm | Duration <br> hh:mm:ss | Distance <br> m | Bearing <br> - | Cruise <br> Day / Night | Comments |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 21 | 16-May | 21:35:05 | 21:40:50 | 0:05:45 | 520 | 320 | Night | No wind, no precipitation |
| 2 | 20 | 16-May | 21:40:50 | 21:44:30 | 0003:40 | 380 | 260 | Night | No wind, no precipitation |
| 3 | 19 | 16-May | 21:44:30 | 21:56:56 | 0:12:26 | 1,550 | 10 | Night | No wind, no precipitation |
| 4 | 18 | 16-May | 21:56:56 | 22:04:15 | 0:07:19 | 890 | 270 | Night | No wind, no precipitation |
| 5 | 17 | 16-May | 22:04:15 | 22:11:35 | 0:07:20 | 720 | 10 | Night | No wind, no precipitation |
| 6 | 16 | 16-May | 22:11:35 | 22:20:37 | 0:09:02 | 1,000 | 330 | Night | No wind, no precipitation |
| 7 | 15 | 16-May | 22:20:37 | 22:27:53 | 0:07:16 | 790 | 90 | Night | No wind, no precipitation |
| 8 | 14 | 16-May | 22:27:53 | 22:35:31 | 0:07:38 | 800 | 315 | Night | No wind, no precipitation |
| 9 | 13 | 16-May | 22:35:31 | 22:43:58 | 0:08:27 | 1,050 | 60 | Night | No wind, no precipitation |
| 10 | 12 | 16-May | 22:43:58 | 22:50:38 | 0:06:40 | 680 | 310 | Night | No wind, no precipitation |
| 11 | 11 | 16-May | 22:50:38 | 23:04:44 | 0:14:06 | 1,500 | 75 | Night | No wind, no precipitation |
| 12 | 10 | 16-May | 23:04:44 | 23:10:53 | 0:06:09 | 660 | 325 | Night | No wind, light precipitation |
| 13 | 9 | 16-May | 23:10:53 | 23:24:22 | 0:13:29 | 1,610 | 90 | Night | No wind, light precipitation |
| 14 | 8 | 16-May | 23:24:22 | 23:33:01 | 0:08:39 | 1,000 | 290 | Night | No wind, light precipitation |
| 15 | 7 | 16-May | 23:33:01 | 23:44:46 | 0:11:45 | 1,360 | 90 | Night | No wind, no precipitation |
| 16 | 6 | 16-May | 23:44:46 | 23:54:46 | 0:10:00 | 1,180 | 270 | Night | No wind, no precipitation |
| 17 | 5 | 16-May | 23:54:46 | 0:08:52 | 0:14:06 | 1,460 | 70 | Night | No wind, no precipitation |
| 18 | 4 | 17-May | 0:08:52 | 0:19:05 | 0:10:13 | 1,180 | 270 | Night | No wind, no precipitation |
| 19 | 3 | 17-May | 0:19:05 | 0:28:54 | 0:09:49 | 1,170 | 55 | Night | No wind, no precipitation |
| 20 | 2 | 17-May | 0:28:54 | 0:37:24 | 0:08:30 | 1,070 | 270 | Night | No wind, no precipitation |
| 21 | 1 | 17-May | 0:37:24 | 0:54:44 | 0:17:20 | 1,380 | 0 | Night | No wind, light precipitation |
| 22 | 1 | 17-May | 0:54:44 | 1:15:07 | 0:20:23 | 1,380 | 180 | Night | No wind, no precipitation |
| 23 | 2 | 17-May | 1:15:07 | 1:23:26 | 0:08:19 | 1,070 | 90 | Night | No wind, no precipitation |
| 24 | 3 | 17-May | 1:23:26 | 1:34:46 | 0:11:20 | 1,170 | 210 | Night | No wind, no precipitation |
| 25 | 4 | 17-May | 1:34:46 | 1:44:41 | 0:09:55 | 1,180 | 110 | Night | No wind, no precipitation |
| 26 | 5 | 17-May | 1:44:41 | 1:57:21 | 0:12:40 | 1,460 | 220 | Night | No wind, no precipitation |
| 27 | 6 | 17-May | 1:57:21 | 2:09:01 | 0:11:40 | 1,180 | 120 | Night | No wind, no precipitation |
| 28 | 7 | 17-May | 2:09:01 | 2:21:03 | 0:12:02 | 1,360 | 215 | Night | Moderate wind, no precipitation |
| 29 | 8 | 17-May | 2:21:03 | 2:29:13 | 0:08:10 | 1,000 | 120 | Night | Moderate wind, no precipitation |
| 30 | 9 | 17-May | 2:29:13 | 2:43:34 | 0:14:21 | 1,610 | 215 | Night | Moderate wind, no precipitation |
| 31 | 10 | 17-May | 2:43:34 | 2:49:17 | 0:05:43 | 660 | 120 | Night | Moderate wind, no precipitation |
| 32 | 11 | 17-May | 2:49:17 | 3:03:11 | 0:13:54 | 1,500 | 215 | Night | No wind, no precipitation |
| 33 | 12 | 17-May | 3:03:11 | 3:09:25 | 0:06:14 | 680 | 125 | Night | No wind, no precipitation |
| 34 | 13 | 17-May | 3:09:25 | 3:16:29 | 0:07:04 | 1,050 | 210 | Night | Light wind, no precipitation |
| 35 | 14 | 17-May | 3:16:29 | 3:23:20 | 0:06:51 | 800 | 150 | Night | Light wind, no precipitation |
| 36 | 15 | 17-May | 3:23:20 | 3:29:14 | 0:05:54 | 790 | 270 | Night | Light wind, no precipitation |
| 37 | 16 | 17-May | 3:29:14 | 3:37:37 | 0:08:23 | 1,000 | 140 | Night | No wind, no precipitation |
| 38 | 17 | 17-May | 3:37:37 | 3:41:51 | 0:04:14 | 720 | 190 | Night | No wind, no precipitation |
| 39 | 18 | 17-May | 3:57:37 | 4:02:52 | 0:05:15 | 890 | 120 | Night | No wind, no precipitation |
| 40 | 19 | 17-May | 4:02:52 | 4:11:32 | 0:08:40 | 1,550 | 190 | Night | Light wind, no precipitation |
| 41 | 20 | 17-May | 4:11:32 | 4:14:50 | 0:03:18 | 380 | 100 | Night | No wind, no precipitation |
| 42 | 21 | 17-May | 4:14:50 | 4:19:00 | 0:04:10 | 520 | 170 |  | No wind, no precipitation |
| 43 | 22 | 17-May | 3:41:51 | 3:47:51 | 0:06:00 | 520 | 330 | Night | Bunzten Tunnel, no wind, no precipitation |
| 44 | 22 | 17-May | 3:47:51 | 3:57:37 | 0:09:46 | 520 | 160 | Night | Bunzten Tumnel, no wind, no precipitation |
|  |  |  |  | Total | 6:43:55 | 44,940 |  |  |  |

Appendix 12. Surarrary of hydroacoustic sampling at Coguitlam Lake Reservoir in October 31 to November 02, 2005.

| $\begin{aligned} & \text { Route } \\ & \text { No. } \end{aligned}$ | Transect " <br> No. | $\begin{array}{r} \text { Date } \\ \text { dd-mm } \end{array}$ | $\begin{gathered} \text { Start } \\ \text { hhimm } \end{gathered}$ | $\begin{array}{r} \text { End } \\ \text { hh:mm } \end{array}$ | Duration hh:mm:ss | $\begin{aligned} & \text { Distunce } \\ & \mathrm{m} \end{aligned}$ | Bcaring | $\begin{array}{r} \text { Cruise } \\ \text { Day } / \text { Night } \end{array}$ | Comments |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 1 | $31-0 \mathrm{ct}$ | 21:28:00 | 21:45:00 | 0:17:00 | 1,989 | 180 | Night | Light wind, no precipitation, electrical noise detceted / delemmined to be depth sounder on vessel |
| 2 | 2 | $31-\mathrm{Oct}$ | 21:45:00 | 21:57:00 | 0:12:00 | 1.404 | 110 | Night | Light wind, no precipitation, electrical roise detceted/determinced to be deph sourder on vessel |
|  | 3 | 31.0 ct | 21:57:00 | 22:11:58 | 0:14:58 | 1,751 | 250 | Night | Light wind, no precipitaion, electrical noise detected/ /determined to be depth sounder on vessel |
| 4 | 4 | $31-\mathrm{Ct}$ | 22:11:58 | 22:22:45 | 0:10:47 | 1,262 | 100 | Night | No wird, no precinitaion, electrical noise detected / delermined to be depth sounder on vessel |
| 5 | 5 | $31-\mathrm{Cct}$ | 22:22:45 | 22:36:37 | 0.13:52 | 1,622 | 250 | Night | No wird, no precisitation, electrical noise detected / delernined to be deph sounder on vessel |
| 6 | 6 | $31-\mathrm{Oct}$ | 22:36:37 | 22:47:06 | 0:10:29 | 1,227 | 110 | Night | No wind, no precinitation, electrical noise detected / determined to be deph sounder on vessel |
| 7 | 7 | 3 --Cot | 22:47:06 | 22:59:12 | 0:12:06 | 1,446 | 250 | Nipht | Lipht wind, sight precipitition, elcctrical noise detected / determined to be depth sounder on vessel |
| 8 | 8 | 31 -0ct | 22:59:12 | 23:09:05 | 0:09:53 | 1.156 | 125 | Night | Light wind, slight precipitation, electrical noiss detected / cetermined to be depti sounder on vessel |
| 9 | 9 | 3 H -0ct | 23:09:05 | 23:23:58 | 0:14:53 | 1,741 | 245 | Night | No wind, heavy presipitation, electrical noise detected/ /determined to be depth sounder on vessel |
| 10 | 10 | 3 3-Oct | 23:23:58 | 23:30:37 | 0:06:39 | 778 | 130 | Night | No wind, heavy precipitation, electrical noise detected/determmed to be depth sounder on vessel |
| 11 | 11 | 3 3-Ott | 23:30:37 | 23:44:11 | 0:13:34 | 1,587 | 240 | Night | No wind, heavy precinitation, elecrrical noise decected/d delemmined to be depth sounder onvessct |
| 12 | 12 | 3 3--Ot | 23:44:11 | 23:50:53 | 0:06:42 | 784 | 120 | Night | No wind, Lizht rrecipitaticn, electrical hoise deleeced / determined to be deph sounder on vessel |
| 13 | 13 | $31-\mathrm{Ot}$ | 23:50:53 | 0:00:19 | 0:09:26 | 1,104 | 240 | Night | No wind no precipitation |
| 14 | 14 | 1 -Nov | 0.00:19 | 0:26:59 | 0:26:40 | 3.120 | 125 | Night | File 00:22:17 also included in same fransect flocal arca connection failedhocked, transect re-run overe wrea needing coverage |
| 15 | 15 | 1 -Nov | 0:26:59 | 0.33:59 | 0:07:00 | 819 | 270 | Night | No wind, no precipitation, electrical noise detected / determincd to be depth sounder on vessel |
| 16 | 16 | 1 -Nov | 0:33:59 | 0:42:49 | 0:088:50 | 1,034 | 125 | Night | No wind, no precipitaion, electrical noise delected / detternined to be deppts sounder on v cessel |
| 17 | 17 | 1 -Nov | 0:42:49 | 0:50:02 | 0:07:13 | 844 | 180 | Night | No wind, no precipitation, dectrical noise delected / determined to be depth sounder on vessel |
| 18 | 18 | 1 -Nov | 0:50:02 | 0.56:39 | 0:06:37 | 774 | 80 | Nipht | No wind, no precipitation, electrical noise deteeted / determined to be depth sounder on vessel |
| 19 | 19 | 1 -Nov | 0:5639 | 1:99:11 | 0:12:32 | 1,466 | 180 | Night | No wind, no precipitation, electrical noise detected / determined to be depth sounder on vessel |
| 20 | 20 | 1 -Nov | 1:09:31 | 1:12:40 | 0:03:29 | 408 | 90 | Night | No wind, no precipitation, electrical noisis detected/determined to be depth sounder on vessel |
| 21 | 21 | 1 -Nov | 1:12:40 | 1:16:00 | 0:03:20 | 390 | 150 | Night | No wind, nop precipitation, electricel noise detected / determincd to be depth sounder on vessel |
| 22 | 1 | 1 -Nov | 10:31:45 | 10:46:00 | 0:14:15 | 1,667 | 170 | Day | No wind. Iight rrecipitation electrical noise detected/ determined to be depth souncer on vessel |
| 23 | 2 | 1 -Nov | 10:46:00 | 10:56:15 | 0:10:15 | 1,199 | 130 | Day | No wind, lizht precipitation electrical noise detected/determined to be depth sounder on vessel |
| 24 | 3 | 1 -Nov | 10:56:15 | 11:05:34 | 0:09:19 | 1,090 | 240 | Duy | No wind, lizht precipitation clectrical noise detected / determined to bo deph s sunder on vessel |
| 25 | 4 | 1 -Nov | 11:05.34 | 11:15:58 | 0:10:24 | 1,217 | 110 | Day | No wind, light precipitation electrical nosise detected / determined to be depph suurner on vessiol |
| 26 | 5 | 1 -Nov | 11:15:58 | 11:28:48 | 0:12:50 | 1,502 | 245 | Day | No wind, light precipitation |
| 27 | 6 | 1 -Nov | 11:28:48 | 11:38:49 | 0:10:01 | 1,172 | 120 | Day | No wind. ligh precipitation |
| 28 | 7 | 1 -Nov | 11:38:49 | 11:50:08 | 0:11:19 | 1,324 | 260 | Day | No wind, light precinitation |
| 29 | 8 | 1 -Nov | 11:50:08 | 11:58:53 | 0:08:45 | 1,024 | 145 | Day | No wind, light preciptation |
| 30 | 9 | 1 -Nov | 11:58. 53 | 12:11:44 | 0:12:51 | 1,503 | ${ }^{230}$ | Day | No wind, light precinitation |
| 31 | 10 | !-Nov | 12:11:44 | 12:17:30 | 0:05:46 | 675 | 150 | Day | No wind, light precipitation |
| 32 | 11 | 1 -Nov | $12: 1730$ | 12.30:21 | 0:12:51 | 1,503 | 230 | Day | No wind, light precipination |
| 33 | 12 | 1 -Nov | $12: 3021$ | 12:35:58 | 0:05.37 | 657 | 110 | Dav | No wind, light precipilation |
| 34 | 13 | ${ }^{1}$ - Now | 12.35 .58 | ${ }^{12: 45.01}$ | 0:095. 03 | 1,059 856 | $\begin{aligned} & 130 \\ & 130 \end{aligned}$ |  | No wind, light precipitation No wind, light rrecipitation |
| $\begin{aligned} & 35 \\ & 36 \end{aligned}$ | 14 15 | ${ }^{\text {1-Now }}$ | 12:45:01 $12: 52: 20$ | $12: 52: 20$ $12: 59: 27$ | 0:07:19 | 856 833 | $\begin{aligned} & 130 \\ & 270 \end{aligned}$ | ${ }_{\text {Day }}^{\text {Day }}$ | No wind, light precinitition No wind, light precipitation |
| 37 | 16 | 1 -Nov | 12:59.27 | 13:07:40 | 0008.13 | 961 | 130 | Day | No wind, light precipitation |
| 38 | 17 | 1 -Nov | 13:0740 | 13:13:4] | 0.0601 | 704 | 195 | Dav | No wind, light precipitation |
| 39 | 18 | 1 -Nov | 13:13:41 | 13:21.06 | 0.0725 | 868 | 80 | Day | No wind. licht precinitaion |
| 40 | 19 | ${ }^{1}$-Nov | 13:21:06 | ${ }^{13: 39: 05}$ | 0:17.59 | 2,104 | ${ }^{200}$ | Day | No wind, lipht prccipitution |
| 41 42 | 20 21 | ${ }_{\text {l }}^{\text {1-Now }}$ | 13:39:05 | $13: 42: 16$ $13: 48: 00$ | 0:03:1! | 372 671 | 80 150 | ${ }_{\text {Day }}^{\text {Day }}$ | No wind, light precipitation No wind, light precipitation |
| 43 | 1 | 1 -Nov | 22:08:20 | 22:22:14 | 0:13:54 | 1,626 | 180 | Nipht | No wind. light precipitation |
| 44 | 2 | 1 -Nov | 22:22:14 | 22:31.48 | 0:0934 | 1,119 | 110 | Nipht | No wind, heavy precipitation |
| 45 | 3 | 1 - Nov | 22:31:48 | 22:41:13 | 0:0925 | 1,102 | 240 | Night | No wind, havay precipitation |
| 46 | 4 | ${ }_{1}^{1-N o v}$ | 22:41:13 | 22:50:50 | $0: 09 \cdot 37$ | 1,125 | 190 | Night | No wird, light precipitation |
| 47 | 5 | 1 -Nor | 22:50:50 | 23:03:47 | 0:12:57 | 1.515 | 225 | Night | No wir.d, heavy precipitation |
| 48 | 6 | 1 -Nov | 23:03:47 | 23:13:37 | 00:99:50 | 1,151 | 130 | Nisht | No wirc. light precipituion |
| 49 | 7 | 1 -Nov | 23:13:37 | 23.24:39 | 0:11:02 | 1,291 | 250 | Nicht | No wird, , beavy precipitution |
| 50 | 8 | 1 -Nov | 23:24:39 | 23:32:42 | 0:08:03 | 942 | 120 | Nipht | No wird, heavy preseipitation |
| 51 | 9 | 2 -Nov | 23:32:42 | 23:45:48 | 0:13:06 | 1,533 | 270 | Night | No wind, heavy preceipitation |
| 52 | 10 | 2 -Nov | 23:45:48 | 23:50.44 | 0:04:56 | 577 | 140 | Nicht | No wind, ligit precipitation |
| 53 | 11 | 2 -Nov | 23.50:44 | 0:02:43 | 0:11:59 | 1,402 | 230 | Nipht | No wind, light precipitation |
| 54 | 12 | 2 2-Nov | 0.02:43 | 00.0925 | 0:06:42 | $\begin{array}{r}784 \\ \hline 145\end{array}$ | 120 | $\xrightarrow{\text { Nieht }}$ | No wind, light precipitation |
| 55 56 | 13 14 | ${ }^{2} 2$-Nov | 0:09:25 | 0:19:12 | 0.09:47 | 1,145 846 | 240 140 | $\xrightarrow{\text { Nipht }}$ Night | No wind, heavy precinitation No wind, moderate precipitation |
| 57 | 15 | 2 -Nov | 0.26:26 | $0.34 \cdot 12$ | 0007:46 | 909 | 270 | Nipht | Light wind, heary precipitation |
| 58 | 16 | 2 -Nov | 0:34:12 | $0.41: 55$ | 007:43 | 903 | 125 | Night | Light wid, heavy precipilution |
| 59 | 17 | 2 -Nov | 0.41:55 | 0.48 .18 | 0006:23 | 747 | 100 | Night | Light wind, moderate precipitation |
| 60 | 18 | 2 -Nov | 0:48:18 | 0.54:29 | 006611 | ${ }^{723}$ | 90 | Night | Light wind, moderate precipitation |
| 61 | 19 | ${ }^{2}$-Nov | 0.54:29 | 1.06:39 | 0.12:10 | 1,424 | 210 | Nipht | No wind, moderale errecipitation |
| 62 | 20 | ${ }^{2}$ 2-Nov | 1:06:39 | 1:10:02 | 0.03:23 | 396 581 | 90 140 | Night | No wind, moderate precipitation |
| 63 | 21 | $2-\mathrm{Nov}$ | 110:02 | 1:15:00 | 00458 | 581 | 140 | Night | No wind, modecrate precipitation |
|  |  |  |  | Nightr' | 3:48:00 | 26,676 |  |  |  |
|  |  |  |  | Day ${ }^{1}$ | 3.1615 | 22,961 |  |  |  |
|  |  |  |  | Night ${ }^{\text { }}$ | 3:06:40 | $\begin{aligned} & 21,840 \\ & 71477 \end{aligned}$ |  |  |  |

Appendix 13. Summary of target strength, sigma, and estimated fish length, by depth strata, of single targets tracked using a $200 \mathrm{kHz}, 6^{\circ}$ transducer aimed downward during night time, at Coquitlam Reservoir, May 2005.

| Strata |  | Target Strength (dB) |  | Sigma ( dB re $1 \mathrm{~m}^{-1}$ ) |  | Expected <br> Mean Target <br> Size (mm) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Depth <br> (m) | Number of Targets |  |  |  |  |  |
|  |  | Mean | SD | Mean | SD |  |
| 2-5 | 15 | -56.7 | 4.6 | $3.65 \mathrm{E}-06$ | 4.18E-06 | 48.4 |
| 5-10 | 121 | -51.5 | 8.2 | $3.62 \mathrm{E}-05$ | $6.77 \mathrm{E}-05$ | 121.5 |
| 10-15 | 412 | -49.2 | 9.0 | $6.01 \mathrm{E}-05$ | $9.99 \mathrm{E}-05$ | 163.8 |
| 15-20 | 798 | -48.5 | 9.1 | $7.23 \mathrm{E}-05$ | $1.16 \mathrm{E}-04$ | 179.6 |
| 20-25 | 1295 | -50.1 | 8.7 | $4.88 \mathrm{E}-05$ | $8.17 \mathrm{E}-05$ | 146.1 |
| 25-30 | 797 | -53.0 | 8.6 | $3.38 \mathrm{E}-05$ | $6.86 \mathrm{E}-05$ | 110.9 |
| 30-35 | 480 | -56.5 | 6.1 | $1.09 \mathrm{E}-05$ | $3.43 \mathrm{E}-05$ | 60.9 |
| 35-40 | 328 | -57.7 | 5.3 | $7.55 \mathrm{E}-06$ | $2.93 \mathrm{E}-05$ | 49.7 |
| 40-45 | 318 | -57.8 | 3.5 | 2.42E-06 | $3.18 \mathrm{E}-06$ | 40.8 |
| 45-50 | 242 | -57.5 | 3.1 | $2.38 \mathrm{E}-06$ | $2.57 \mathrm{E}-06$ | 41.3 |
| 50-55 | 99 | -56.0 | 5.1 | $6.84 \mathrm{E}-06$ | $1.66 \mathrm{E}-05$ | 56.8 |
| 55-60 | 72 | -57.6 | 3.2 | $2.31 \mathrm{E}-06$ | $2.25 \mathrm{E}-06$ | 40.8 |
| 60-65 | 78 | -56.2 | 3.2 | $3.36 \mathrm{E}-06$ | $4.39 \mathrm{E}-06$ | 48.3 |
| 65-70 | 64 | -56.1 | 3.3 | $3.28 \mathrm{E}-06$ | $2.76 \mathrm{E}-06$ | 48.6 |
| 70-75 | 52 | -56.3 | 3.5 | 3.19E-06 | $2.53 \mathrm{E}-06$ | 48.1 |
| 75-80 | 41 | -56.1 | 2.7 | $2.97 \mathrm{E}-06$ | $1.97 \mathrm{E}-06$ | 47.5 |
| 80-85 | 33 | -55.5 | 2.3 | $3.23 \mathrm{E}-06$ | $1.63 \mathrm{E}-06$ | 50.3 |
| Grand Total | 5,245 | -52.6 | 53.3 | $3.58 \mathrm{E}-05$ | 4.78E-09 | 113.2 |

Appendix 14. Summary of target strength, sigma, and estimated fish length, by depth strata, of single targets tracked using a $420 \mathrm{kHz}, 6^{\circ}$ transducer aimed sideward during night time, at Coquitlam Reservoir, May 2005.

| Strata <br> Range <br> (m) | Number of Targets | Target Strength (dB) |  | Sigma (dB re $1 \mathrm{~m}^{-1}$ ) |  | Expected <br> Mean Target <br> Size (mm) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  |  |
|  |  | Mean | SD | Mean | SD |  |
| 0-5 | 22 | -57.8 | 4.8 | 4.39E-06 | 1.12E-05 | 45.4 |
| 5-10 | 414 | -57.5 | 4.6 | 3.91E-06 | 8.22E-06 | 45.9 |
| 10-15 | 673 | -57.2 | 4.6 | $4.59 \mathrm{E}-06$ | $1.56 \mathrm{E}-05$ | 47.6 |
| 15-20 | 660 | -56.5 | 4.2 | 4.36E-06 | 9.23E-06 | 49.8 |
| 20-25 | 239 | -55.2 | 4.9 | $8.40 \mathrm{E}-06$ | $2.83 \mathrm{E}-05$ | 61.7 |
| Grand Total | 2008 | -56.8 | 57.0 | 4.83E-06 | $1.98 \mathrm{E}-10$ | 49.6 |

Appendix 15 Summary of target strength, sigma, and estimated fish length, by depth strata, of single targets tracked using a $200 \mathrm{kHz}, 6^{\circ}$ transducer aimed downward during night time, at Coquitlam Reservoir, November 2005.

| Strata |  | Target Strength (dB) |  | Sigma (dB re $1 \mathrm{~m}^{-1}$ ) |  | Expected <br> Mean Target Size (mm) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Depth <br> (m) | Number of Targets |  |  |  |  |  |
|  |  | Mean | SD | Mean | SD |  |
| 2-5 | 76 | -52.6 | 6.3 | $3.44 \mathrm{E}-05$ | $1.25 \mathrm{E}-04$ | 99.0 |
| 5-10 | 154 | -52.5 | 6.6 | $2.04 \mathrm{E}-05$ | 5.41E-05 | 94.3 |
| 10-15 | 529 | -45.7 | 8.9 | $1.05 \mathrm{E}-04$ | $1.46 \mathrm{E}-04$ | 231.2 |
| 15-20 | 1406 | -43.0 | 8.1 | $1.33 \mathrm{E}-04$ | $1.46 \mathrm{E}-04$ | 281.6 |
| 20-25 | 1302 | -44.5 | 8.4 | $1.11 \mathrm{E}-04$ | $1.36 \mathrm{E}-04$ | 248.6 |
| 25-30 | 402 | -48.1 | 9.0 | $7.27 \mathrm{E}-05$ | $1.05 \mathrm{E}-04$ | 184.3 |
| 30-35 | 129 | -52.1 | 7.0 | $2.50 \mathrm{E}-05$ | 5.51E-05 | 103.1 |
| 35-40 | 84 | -54.9 | 5.0 | $6.15 \mathrm{E}-06$ | 7.43E-06 | 61.7 |
| 40-45 | 45 | -56.8 | 3.8 | $3.23 \mathrm{E}-06$ | 3.88E-06 | 46.1 |
| 45-50 | 51 | -54.8 | 7.0 | $2.84 \mathrm{E}-05$ | $9.16 \mathrm{E}-05$ | 85.1 |
| 50-55 | 65 | -53.6 | 5.2 | $9.94 \mathrm{E}-06$ | $1.67 \mathrm{E}-05$ | 73.6 |
| 55-60 | 62 | -54.1 | 4.5 | $6.40 \mathrm{E}-06$ | 6.18E-06 | 65.3 |
| 60-65 | 30 | -53.0 | 5.8 | $1.18 \mathrm{E}-05$ | $1.44 \mathrm{E}-05$ | 81.5 |
| 65-70 | 32 | -54.9 | 3.3 | $4.27 \mathrm{E}-06$ | 3.70E-06 | 55.6 |
| 70-75 | 38 | -53.0 | 4.4 | $7.96 \mathrm{E}-06$ | $7.17 \mathrm{E}-06$ | 73.4 |
| 75-80 | 25 | -53.4 | 2.3 | 5.27E-06 | 3.09E-06 | 64.2 |
| Grand Total | 4,430 | -46.0 | 46.9 | $9.80 \mathrm{E}-05$ | 8.62E-09 | 221.3 |

Appendix 16 Summary of target strength, sigma, and estimated fish length, by depth strata, of single targets tracked using a $420 \mathrm{kHz}, 6^{\circ}$ transducer aimed sideward during night time, at Coquitlam Reservoir, November 2005.

| Strata |  |  |  |  |  | Expected |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Range | Number of Targets | Target Strength (dB) |  | Sigma (dB re $1 \mathrm{~m}^{-1}$ ) |  | Mean Target |
| (m) |  | Mean | SD | Mean | SD | Size (mm) |
| 0-5 | 35 | -57.6 | 4.3 | $2.84 \mathrm{E}-06$ | 3.10E-06 | 43.3 |
| 5-10 | 97 | -57.6 | 4.2 | 3.12E-06 | 5.16E-06 | 43.6 |
| 10-15 | 41 | -55.3 | 4.5 | $5.28 \mathrm{E}-06$ | 7.52E-06 | 57.1 |
| 15-20 | 15 | -54.1 | 6.4 | $1.66 \mathrm{E}-05$ | 4.17E-05 | 79.4 |
| 20-25 | 8 | -55.6 | 5.7 | 7.76E-06 | $1.56 \mathrm{E}-05$ | 60.9 |
| Grand Total | 196 | -56.6 | 57.09 | 5.16E-06 | $1.51 \mathrm{E}-10$ | 49.8 |

Appendix 17. Summary of target strength, sigma, and estimated fish length, by depth strata, of single targets tracked using a $200 \mathrm{kHz}, 6^{\circ}$ transducer aimed downward during day time, at Coquitlam Reservoir, November 2005.

| Strata |  | Target Strength (dB) |  | Sigma (dB re $1 \mathrm{~m}^{-1}$ ) |  | Expected |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Depth <br> (m) | Number of Targets |  |  | Mean Target |
|  |  | Mean | SD |  |  | Mean | SD | Size (mm) |
| 2-5 | 8 | -54.04 | 3.3737 | 4.99E-06 | 3.30E-06 | 61.2 |
| 5-10 | 9 | -56.72 | 4.6232 | $3.28 \mathrm{E}-06$ | $2.89 \mathrm{E}-06$ | 47.6 |
| 10-15 | 58 | -50.45 | 9.4502 | $6.25 \mathrm{E}-05$ | $1.20 \mathrm{E}-04$ | 155.0 |
| 15-20 | 71 | -54.30 | 5.9803 | $1.20 \mathrm{E}-05$ | $2.69 \mathrm{E}-05$ | 73.6 |
| 20-25 | 146 | -51.31 | 8.134 | $3.98 \mathrm{E}-05$ | $7.75 \mathrm{E}-05$ | 125.4 |
| 25-30 | 103 | -53.67 | 5.4675 | $1.32 \mathrm{E}-05$ | $3.25 \mathrm{E}-05$ | 76.5 |
| 30-35 | 67 | -55.76 | 3.7678 | $4.08 \mathrm{E}-06$ | 5.54E-06 | 52.1 |
| 35-40 | 53 | -56.29 | 4.0721 | $3.74 \mathrm{E}-06$ | 4.32E-06 | 49.7 |
| 40-45 | 39 | -57.84 | 4.5805 | 4.21E-06 | $9.56 \mathrm{E}-06$ | 44.9 |
| 45-50 | 58 | -53.23 | 4.0157 | 6.84E-06 | $5.79 \mathrm{E}-06$ | 69.6 |
| 50-55 | 92 | -55.71 | 3.533 | $3.67 \mathrm{E}-06$ | $2.96 \mathrm{E}-06$ | 51.4 |
| 55-60 | 49 | -56.73 | 2.9987 | $2.67 \mathrm{E}-06$ | $1.84 \mathrm{E}-06$ | 44.7 |
| 60-65 | 57 | -52.68 | 5.9818 | $1.78 \mathrm{E}-05$ | $3.73 \mathrm{E}-05$ | 89.0 |
| 65-70 | 55 | -55.60 | 2.8671 | $3.44 \mathrm{E}-06$ | $2.58 \mathrm{E}-06$ | 50.7 |
| 70-75 | 31 | -52.93 | 4.2555 | $7.50 \mathrm{E}-06$ | $6.03 \mathrm{E}-06$ | 72.7 |
| 75-80 | 33 | -51.83 | 4.1701 | $9.81 \mathrm{E}-06$ | 8.85E-06 | 82.6 |
| Grand Total | 929 | -54.0 | 54.3 | $1.62 \mathrm{E}-05$ | $2.12881 \mathrm{E}-09$ | 78.6 |

Appendix 18. Summary of target strength, sigma, and estimated fish length, by depth strata, of single targets tracked using a $420 \mathrm{kHz}, 6^{\circ}$ transducer aimed sideward during day time, at Coquitlam Reservoir, November 2005.

| Strata |  | Target Strength (dB) |  |  |  | Expected <br> Mean Target |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Range (m) | Number of Targets |  |  | Sigma (dB |  |  |
|  |  | Mean | SD | Mean | SD | Size (mm) |
| 0-5 | 6 | -54.2 | 5.3 | 7.96E-06 | 1.18E-05 | 67.6 |
| 5-10 | 161 | -57.1 | 4.2 | 4.64E-06 | $1.66 \mathrm{E}-05$ | 47.3 |
| 10-15 | 360 | -55.9 | 5.4 | $9.14 \mathrm{E}-06$ | $3.34 \mathrm{E}-05$ | 60.0 |
| 15-20 | 421 | -55.6 | 5.7 | $1.31 \mathrm{E}-05$ | 5.25E-05 | 65.3 |
| 20-25 | 215 | -55.5 | 5.4 | $1.14 \mathrm{E}-05$ | $4.27 \mathrm{E}-05$ | 64.0 |
| Grand Total | 1163 | -55.9 | 56.2 | $1.04 \mathrm{E}-05$ | 1.62E-09 | 61.0 |

Appendix 19. Fish densities (No. - hestare ${ }^{-1}$ ), by depth strata, for hydroacoustic survecys at Coqutilam Reservoir in Mav, 2005.

| Cruise 1 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Denh (m) | I | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 | 17 | 18 | 19 | 20 | 21 | Mean |  | Variance | Samplo <br> Size | $\begin{aligned} & \text { Stratum } \\ & \text { Area (hat) } \end{aligned}$ |
| 0.5 | 119 | 255 | 28 | 268 | 467 | 184 | 97 | 142 | 26 | 112 | 216 | 92 | 144 | 85 | 114 | 11 | 69 | 42 | 45 | 87 | 55 | 127 | 106 | 11,284 | 21 | 1202 |
| 5-10 | 69 | 134 | 95 | 119 | 9 | ${ }_{96}$ | 58 | 14 | 3 | 6 | 107 | 1 | 36 | 86 | 5 | 6 | 51 | 8 | 35 | 0 | 1 | 45 | 45 | 2.048 | 21 | 1140 |
| $10-15$ | 4 | 29 | 45 | 113 | 11 | 50 | 33 | 31 | 119 | 257 | 153 | 85 | 30 | 43 | 49 | 69 | 19 | 59 | 1 | 0 | 0 | 58 | ${ }^{6} 3$ | 4.030 | 21 | 1030 |
| 15-20 | 51 | 38 | 17 | 82 | 48 | 81 | 49 | 36 | 103 | 87 | 30 | 2. | 59 | 57 | 41 | 24 | 2 | 1 | 0 |  |  | 4 | 30 | 884 | 19 | 932 |
| $20-25$ | 43 | 4 | 28 | 51 | 13 | 79 | 59 | 71 | 123 | 48 | 120 | 16 | 107 | 122 | 61 | 19 | 0 | 15 | 0 |  |  | 52 | 42 | 1,803 | 19 | 869 |
| 25-30 | 2 | 0 | 14 | 4 | 3 | 0 | 3 | 36 | 19 | 50 | 17 | 32 | ${ }^{4}$ | 46 | 2 | 24 | 7 | 2 | 0 |  |  | 14 | 16 | 265 | 19 | ${ }_{8}^{834}$ |
| 30-35 | 1 | 0 | 2 | 2 | 76 | 3 | 1 | 18 | 1 | 27 | 12 | 0 | 0 | 2 | 0 | 2 | 0 | 0 |  |  |  | 8 | 19 | 344 | 18 | 802 |
| 35.40 | 4 | 0 | 1 | 1 | 1 | 2 | 1 | 0 | 18 | 1 | 0 | 1 | 27 | 0 | 0 | 0 | 0 | ${ }^{0}$ |  |  |  | 3 | 7 | 53 | 18 | ${ }_{7}^{767}$ |
| 40.45 | 4 | 0 | 3 | 6 | 7 | 2 | $s$ | 1 | 0 | 2 | 1 | 2 | 0 | 0 | 0 | 0 | 7 | 0 |  |  |  | 2 | 3 | 7 | 18 | 735 |
| 45-50 | ! | 0 | 7 | 21 | 3 | 4 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  | 2 | 5 | 28 | 17 | 705 |
| 50-55 | 0 | 0 | 2 | 2 | 0 | 4 | 2 | 0 | 11 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  | 1 | 3 | 7 | 17 | ${ }^{672}$ |
| 55-60 |  | 0 | 3 | 2 | 0 | 0 | 1 | 0 | 1 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  | 1 | ! | 1 | 16 | ${ }^{633}$ |
| 60.65 |  | 0 | 4 | 1 | 0 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  | 0 | ! | 1 | 16 | 584 |
| 65.70 |  | 0 | 0 | 1 | : | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  | 0 | 1 | 0 | 15 | 545 519 |
| 70.75 |  |  | 0 | 0 | 0 | 1 | - | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  | 0 | 0 | 0 | 14 | 519 |
| 75-80 |  |  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  | ${ }^{0}$ | ${ }^{0}$ | 0 | 14 | 493 |
| $80-85$ |  |  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  | 0 | 0 | 0 | 14 | 464 |
| 85.90 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | $\bigcirc$ | 436 409 |
| 90.-95 $95-100$ |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 0 | 384 |
| $100-105$ |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 0 | 359 |
| 105-110 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 97 | 336 |


| Ciuse 2 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Deprb (m) |  | 2 | 3 | 4 | 5 |  | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 | 17 | 18 | 19 | 20 | $2!$ | Mcan | sidev | Variance | Size Area tha |  |
|  | 31 | 155 | 143 | 73 | 190 | 157 | 318 |  | 198 | 181 | 189 | 154 | 2 | 58 | 149 | 9 | 74 | 67 | 1 | 14 | 13 | 118 | 95 | 8,949 | 21 | ${ }^{1202}$ |
| 5-10 | 23 | 59 | 27 | 56 | 44 | 7 | 318 | 31 | 246 | 70 | 132 | 439 | 380 | 35 | 50 | 113 | 21 | 3 | 11 | 18 | 4 | 99 | 131 | 17,124 | 21 | 1140 |
| 10.15 | so | 140 | 63 | 52 | 162 | 138 | 45 | 58 | 108 | 245 | 34) | 109 | 172 | 126 | 92 | 38 | 2 | 2 | 3 | 2 | 0 | 93 | 88 | 7,675 | 21 | 1030 |
| 15-20 | 79 | 86 | 22 | 44 | 53 | 49 | 108 | 113 | 47 | 28 | 168 | 125 | 32 | 124 | 13 | :2 | 0 | 0 | 7 |  |  | 58 | 50 | 2.470 | 19 | 932 |
| 20-25 | 13 | 27 | 4 | 0 | 22 | 90 | 28 | 63 | 83 | 25 | 51 | 129 | 63 | : | 0 | 0 | 1 | 1 | 1 |  |  | 32 | 38 | 1,435 | 19 | 869 |
| 25-30 | 2 | 0 | 0 | 0 | 2 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 10 | 0 | $\bigcirc$ | $\bigcirc$ | 0 | 0 | 0 |  |  | 1 | 2 | 5 | 19 | 834 |
| 30-35 | 6 | 0 |  | 0 | 0 | 0 | 0 | 1 | 1 | 0 | 0 | 0 | 2 | 0 | 0 | 0 | 1 | 0 |  |  |  | 1 | $\stackrel{2}{2}$ | 5 | 18 | 802 |
| 35-40 | 4 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 1 | 3 | 0 | 0 | 0 | 0 | 0 |  |  |  | 1 | 1 | 1 | 18 | ${ }^{767}$ |
| 40-45 | 8 | 0 | 1 | 0 | 5 | 0 | 0 | 1 | 1 | 4 | 0 | 0 | 13 | 0 | 0 | 0 | 0 | 2 |  |  |  | $\stackrel{2}{2}$ | ${ }^{4}$ | 12 | 18 | ${ }_{7} 735$ |
| 45-50 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 2 | 0 | 0 | 9 | 0 | 0 | 0 | 0 |  |  |  |  | 1 | 2 | 5 | 17 | 705 |
| 50.55 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 |  | 0 | 0 |  |  |  |  | 0 | 0 | 0 | 17 |  |
| 55-60 |  | 0 | 0 |  | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 1 | 1 | 0 | 0 | 0 | 0 |  |  |  |  | ${ }^{0}$ | 0 | 0 | 16 | ${ }_{593} 6$ |
| 60-65 |  | 0 | 0 | 0 | 0 | 3 | 0 | 2 | 0 | 0 | 0 | 0 | 7 | 0 | 0 | 0 | 0 |  |  |  |  | 1 | 2 |  | 16 |  |
| 65-70 |  | , | 0 | 0 | 0 | 0 | 0 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  | 0 | 0 | 0 | 15 | 545 |
| 70.75 |  |  | 0 | 0 | 1 | 0 | 3 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  | 0 | 1 | , | 14 | 519 |
| 75-80 |  |  | 1 | 0 | 0 | 0 | 0 | ; | 0 | 0 | 0 | 0 | 2 | 0 | 2 | 0 |  |  |  |  |  | 0 | 1 | 0 | 14 | 493 |
| 80.85 |  |  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  | 14 |  |
| 85.90 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | $\bigcirc$ | 436 409 |
| 90-95 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 0 |  |
| 25-100 $100 \cdot 105$ |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 0 | 359 |
| 105-110 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 0 | 336 |
| Total | 219 | 466 | 270 | 225 | 477 | 443 | 821 | 578 | 685 | 556 | 882 | 958 | 695 | 345 | 306 | 172 | 99 | 76 | 23 | 34 | 18 | 407 | 416 | 37,686 |  |  |


| Mean Densisty (All Cruises) |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| $\frac{\text { Depth (m) }}{0.5}$ |  |  | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 | 17 | 18 | 19 | 20 | 21 | Mcan | Stev | Varrance | Sample Stratum Size Area (ha) |  |
|  | 75 | 205 | 85 |  | 328 | 171 |  |  | 112 |  | 203 |  | 73 | 71 | 132 | 10 | 72 | 54 | 23 |  | 34 | 122 | 80 | 6,479 | 21 | 1202 |
| 5-10 | 46 | 97 | 61 | 87 | 26 | 51 | 188 | 22 | 125 | 38 | 119 | 220 | 208 | 61 | 27 | 59 | 36 | 5 | 23 | 9 | 3 | 72 | 66 | 4,295 | ${ }^{21}$ | 1140 |
| 10-15 | 27 | 84 | 54 | 83 | 87 | 94 | 39 | 44 | 113 | 256 | 247 | 97 | 101 | 84 | 70 | 54 | 10 | 30 | 2 | 1 | 0 | 75 | 68 | 4,686 | 21 | 1030 |
| 15-20 | 65 | 62 | 20 | 63 | 50 | 65 | 78 | 75 | 75 | 57 | 99 | 74 | 46 | 90 | 27 | 18 | 1 | 1 | 4 |  |  | 51 | 31 | 945 | 19 | 932 |
| 20.25 | 28 | 15 | 16 | 26 | 17 | 84 | 44 | 67 | 103 | 36 | 85 | 72 | 85 | 61 | 30 | 10 | 0 | 8 | 1 |  |  | 42 | 33 | 1,083 | 19 | 869 |
| 25-30 | 2 | 0 | 7 | 2 | 2 | 0 | , | 19 | 9 | 25 | 9 | 16 | 7 | 23 | 1 | 12 | 4 | 1 | 0 |  |  | 7 | 8 | ${ }^{65}$ | 19 | 834 |
| 30-35 | 3 | 0 | 5 | 1 | 38 | 1 | 1 | 10 | 1 | 14 | 6 | 0 | 1 | 1 | 0 | 1 | 0 |  |  |  |  | 5 | 9 | 84 | 18 | 802 |
| 35-40 | 4 | 0 | 1 | 0 | 0 | 1 | 1 | 0 | 9 | 0 | 0 | I | 15 | 0 | 0 | 0 | 0 | $\bigcirc$ |  |  |  |  | 4 | $\stackrel{16}{5}$ | 18 | 767 735 |
| 40.45 | 6 | 0 | 2 | 3 | 0 | 1 | 3 | 1 | 0 | 3 | 0 | , | 6 | 0 | 0 | 0 | 4 | 1 |  |  |  |  | 2 | 5 | 18 | 735 |
| 45-50) | 1 | 0 | 4 | 11 | 1 | 2 | 1 | 0 | 0 | 1 | 0 | 0 | 5 | 0 | 0 | 0 | 0 |  |  |  |  | 2 | 3 |  | 17 |  |
| 50.55 | 0 | 0 | 1 | 1 | 0 | 2 | 1 | , | 6 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  | 1 | 1 | 2 | 17 | 672 |
| 55-60 |  | 0 | 1 | 1 | 0 | 0 | , | I | 0 | 0 | 1 | 1 | 1 | 0 | 0 | 0 | 0 |  |  |  |  | 1 | 1 | 0 | 16 | 633 584 |
| 60.65 |  | 0 | 2 | 0 | 0 | 2 | 0 | 1 | 0 | 0 | 0 | 0 | 4 | 0 | 0 | 0 | 0 |  |  |  |  | 1 | 1 | $!$ | 16 | 584 545 |
| 65.70 |  | 0 | 0 | 0 | 1 | 1 | , | 1 | 0 | 0 | 0 |  | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  | 15 | 545 519 |
| 70.75 |  |  | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  | 0 | 0 | 0 | 14 | 519 |
| 75-80 |  |  | 1 | 0 | 0 |  | 0 | 1 | 0 | 0 | 0 | 0 | 1 | 0 | 1 | $\bigcirc$ |  |  |  |  |  | 0 | 0 | 0 | 14 | 493 |
| 80-85 |  |  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  | 0 | 0 | 0 | 14 |  |
| 85-9\% |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 0 |  |
| 90-95 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 0 | $\stackrel{384}{ }$ |
| 95-100 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 0 | 359 |
| 105-110 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 97 | 336 |
| Total | 258 | 463 | 26) | 449 | 559 | 476 | 566 | 464 | 554 | 579 | 770 | 606 | 551 | 392 | 289 | 163 | 127 | 101 | 52 | 60 | 37 | 382 | 308 | 12,669 | 297 |  |


| Night Cruise |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | Sample | Stratum |
| Depth (m) | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 | 17 | 18 | 19 | 20 | 21 | Mean | Stdev | Variance | Size | Area (ha) |
| 0-5 | 19 | 10 | 13 | 19 | 21 | 8 | 12 | 47 | 233 | 181 | 18 | 36 | 41 | 19 | 78 | 32 | 491 | 325 | 322 | 25 | 82 | 97 | 135 | 18,127 | 21 | 1202 |
| 5-10 | 40 | 4 | 15 | 0 | 19 | 2 | 5 | 10 | 2 | 1 | 1 | 21 | 10 | 0 | 8 | 10 | 25 | 5 | 1 | 6 | 6 | 9 | 10 | 100 | 21 | 1140 |
| 10-15 | 16 | 22 | 11 | 1 | 5 | 6 | 6 | 0 | 5 | 0 | 2 | 0 | 1 | 3 | 0 | 20 | 10 | 31 | 1 | 0 | 0 |  | 9 | 76 | 21 | 1030 |
| 15-20 | 23 | 35 | 50 | 14 | 83 | 21 | 75 | 32 | 28 | 59 | 21 | 13 | 29 | 24 | 4 | 0 | 6 | 40 | 10 |  |  | 30 | 23 | 534 | 19 | 932 |
| 20-25 | 12 | 45 | 75 | 53 | 46 | 18 | 117 | 49 | 68 | 44 | 26 | 25 | 24 | 19 | 26 | 12 | 19 | 6 | 0 |  |  | 36 | 28 | 802 | 19 | 869 |
| 25-30 | 2 | 0 | 25 | 14 | 10 | 17 | 23 | 11 | 16 | 14 | 6 | 10 | 4 | 1 | 9 | 0 | 1 | 5 | 0 |  |  | 9 | 8 | 60 | 19 | 834 |
| 30-35 | 1 | 0 | 84 | 0 | 6 | 0 | 1 | 41 | 11 | 0 | 0 | 9 | 1 | 0 | 21 | 0 | 2 | 26 |  |  |  | 11 | 21 | 459 | 18 | 802 |
| 35-40 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 1 | 1 | 0 | 0 | 0 | 0 |  |  |  | 0 | 0 | 0 | 18 | 767 |
| 40-45 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 0 | 0 | 0 | 2 | 3 | 0 | 0 | 0 | 0 |  |  |  | , | 1 | 1 | 18 | 735 |
| 45-50 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  | 0 | 0 | 0 | 17 | 705 |
| 50-55 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  | 0 | 0 | 0 | 17 | 672 |
| 55-60 |  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 0 | 3 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  | 0 | 1 | 1 | 16 | 633 |
| 60-65 |  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 |  |  |  |  | 0 | 0 | 0 | 16 | 584 |
| 65-70 |  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  | 0 | 0 | 0 | 15 | 545 |
| 70-75 |  |  | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  | 0 | 0 | 0 | 14 | 519 |
| 75-80 |  |  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  |  | 0 | 0 | 14 | 493 |
| 80-85 |  |  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  | 0 | 0 | 0 | 14 | 464 |
| 85-90 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 0 | 436 |
| 90-95 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 0 | 409 |
| 95-100 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 0 | 384 |
| 100-105 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 0 | 359 |
| 105-110 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 0 | 336 |
| Total | 94 | 106 | 272 | 100 | 191 | 75 | 239 | 190 | 370 | 299 | 80 | 114 | 112 | 71 | 146 | 74 | 553 | 438 | 334 | 31 | 89 | 200 | 237 | 20,160 | 297 |  |


| Day Cruise |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Depth (m) | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 | 17 | 18 | 19 | 20 | 21 | Mean | Stdev | Variance | $\begin{array}{r} \hline \text { Sample } \\ \text { Size } \\ \hline \end{array}$ | Stratum Area (ha) |
| 0.5 | - | 12 | 8 | 24 | 23 | 213 | 107 | 34 | 38 | 3 | 23 | 55 | 292 | 18 | 346 | 78 | 98 | 82 | 218 | 23 | 201 | 95 | 103 | 10,576 | 20 | 1202 |
| 5-10 | - | 0 | 1 | 0 | 0 | 0 | 7 | 0 | 2 | 0 | 0 | 9 | 0 | 2 | 0 | 0 | 0 | 0 | 5 | 0 | 0 | 1 | 3 | 7 | 20 | 1140 |
| 10-15 | - | 0 | 0 | 0 | 28 | 0 | 4 | 1 | 0 | 60 | 32 | 0 | 0 | 5 | 0 | 0 | 1 | 0 | 48 | 0 | 0 | 9 | 18 | 323 | 20 | 1030 |
| 15-20 | - | 0 | 0 | 0 | 1 | 0 | 21 | 2 | 39 | 19 | 37 | 0 | 0 | 1 | 0 | 0 | 2 | 0 | 0 |  |  | 7 | 13 | 169 | 18 | 932 |
| 20-25 | . | 0 | 0 | 0 | 1 | 8 | 3 | 7 | 0 | 0 | 12 | 0 | 4 | 0 | 0 | 0 | 74 | 0 | 0 |  |  | 6 | 17 | 296 | 18 | 869 |
| 25-30 | - | 0 | 0 | 0 | 0 | 8 | 27 | 27 | 0 | 1 | 29 | 0 | 1 | 54 | 0 | 44 | 0 | 0 | 0 |  |  | 11 | 17 | 305 | 18 | 834 |
| 30-35 | - | 0 | 0 | 0 | 1 | 1 | 1 | 1 | 0 | 4 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 |  |  |  | 1 | 1 | 1 | 17 | 802 |
| 35-40 | - | 0 | 0 | 0 | 0 | 1 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  | 0 | 0 | 0 | 17 | 767 |
| 40-45 | - | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  | 0 | 0 | 0 | 17 | 735 |
| 45-50 | - | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  | 0 | 0 | 0 | 16 | 705 |
| 50-55 | - | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 1 |  |  |  |  | 0 | 0 | 0 | 16 | 672 |
| 55-60 |  | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  | 0 | 0 | 0 | 16 | 633 |
| 60-65 |  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  | 0 | 0 | 0 | 15 | 584 |
| 65-70 |  | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  | 0 | 0 | 0 | 15 | 545 |
| 70.75 |  |  | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  | 0 | 0 | 0 | 14 | 519 |
| 75-80 |  |  | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  | 0 | 0 | 0 | 14 | 493 |
| 80-85 |  |  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  | 0 | 0 | 0 | 14 | 464 |
| 85-90 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 0 | 436 |
| 90-95 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 0 | 409 |
| 95-100 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 0 | 384 |
| 100-105 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 0 | 359 |
| 105-110 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 0 | 336 |
| Total | 0 | 12 | 9 | 24 | 56 | 236 | 176 | 73 | 80 | 87 | 135 | 63 | 298 | 80 | 346 | 123 | 176 | 82 | 271 | 23 | 201 | 130 | 174 | 11,677 | 285 | - |



| Gear |  |  |  |  |  |  |  |  |  |  | Cutch" |  |  |  |  |  |  |  |  |  |  | CPUE ${ }^{\text {b }}$ |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Sur Daic | Stan Sel (hhimniss) | Completely Oet Star haul in(hhmmss) (htmmes) |  |  |  | $\begin{gathered} \text { Depph } \\ (m) \end{gathered}$ | Northing | Wetins | Rexion | Area | Stie 4 Sut(No.) | CT | ко | co | Su | NSC | PCC | RSS | sc | TSSB | Towl | ст | ко | co | Su | NSC | PCC | RSS | sc | TSS3 | Toual |
| 15May0s | 20:3:00 | 20.53:00 | 13:15:00 | 14.00:90 | 16:33:00 | 15 | 49\%23.595 | $1222^{247.006}$ | South | Limmecic | 8 : | 4 | 16 | 0 | 0 | 6 | 5 | 0 | 0 | $\bigcirc$ | 31 | 0.24 | 0.95 | 0.00 | 0.00 | 0.36 | 0.36 | $\cdots$ | 0.00 | 0.00 | ${ }^{1.94}$ |
| 15 mayns | 19:00:3\% | 19:27:00 | 0930:00 | 10:00:00 | 14:34.30 | 5 | $4{ }^{2} 22.148$ | $122^{247} 265$ | South | Litural | 41 | 5 | 16 | 0 | 0 | 7 | 0 | 0 | $\bigcirc$ | $\bigcirc$ | 28 | 0.34 | 1.15 | 0.00 | 0.00 | 0,48 | 0.00 0.34 | 0,00 | 0.00 0.00 | 0,000 | 1.192 <br> 1.93 <br> 1.98 |
| 15 May ${ }^{\text {as }}$ | 20:02:00 | 20.25:00 | 10:30:90 | 11:00:00 | 14:31:30 | 5 | $49^{2222074}$ | ${ }^{1222^{\text {a }} 4.7 .726}$ | South | Limnecic | 31 | ${ }^{3}$ | 17 | 0 | 0 | 3 | 5 | 0 | 0 | - | 28 38 | 0.21 | 1.92 | 0.00 | 0.00 | 0.36 | 0.00 | 0.000 | 0.00 | 0.00 | \% |
| 15May ${ }^{\text {a }}$ | 19:32:80 | 19:50:00 | 12:00:00 | 12:45:00 | 16:4130 | 10 | $4{ }^{4923.323}$ | 122"48802 | South | Limmetic | 71 | , | 32 | 0 | 0 | 6 | 2 | 0 | - | 0 | 16 | 0.21 | 0.13 | 0.09 | 0.00 | 0.25 | 0.08 | 0.00 | 0.00 | 0.00 | 0.67 |
| $16, M^{2 y} 05$ | 09:36:00 | 10:00:00 | 093:30.03 103000 | 10.0000 11.0000 | 24.00:00 $24.22: 30$ | ${ }_{15}^{5}$ |  | $122^{4}+7.7261$ $122^{4}+7.726$ | South South | ${ }_{\text {Litural }}$ | 4 3 | 5 | 16 | - | 0 | 2 | ${ }_{0}$ | 0 | 0 | 0 | 22 | 0.16 | 0.66 | 0.00 | 0.00 | 0.08 | 0.00 | 0.00 | 0.00 | 0.00 | 0.90 |
| ${ }^{\text {commayas }}$ | 12:00:00 | 12:48:30 | 1130000 | 1200:00 | 23:21:00 | 15 | 4, ${ }^{2 \times 2,3,323}$ | 122448.202 | South | Limmetic | 72 | 3 | 16 | 0 | 1 | 4 | 2 | 3 | 0 | 0 | 29 | 0.13 | 0.69 | 0.00 | 0.04 | 0.17 | 0.09 | 0.13 | 0.00 | 0.00 | $1: 24$ |
| 16 Mmyas | 13:15:00 | 14:00:00 | 12:2000 | 13.1500 | 23:9:60 | 25 | $49^{\circ 23.595}$ | $122^{\text {di4, }}$, 606 | South | Limnetic | $8{ }^{2}$ | 2 | ${ }^{6}$ | 0 | 5 | 14 | 12 | 0 | 0 | 0 | ${ }^{39}$ | 0.99 | 0.26 | 0.00 | 0.22 | 0.60 | 0.52 | 0.00 000 | 0.00 0.000 | 0.00 0.00 |  |
| 17May95 | 19:00:00 | 1935:50 | 00,30,00 | 10000:00 | 14:273:30 | 20 | $4{ }^{49^{2} 24.559}$ | $1222^{147} 7388$ | Centual | Limmetic |  | $\bigcirc$ | ${ }^{2}$ | $\bigcirc$ | $\bigcirc$ | ${ }_{7}$ | 0 | $\bigcirc$ | $\bigcirc$ | 0 | 23 | ${ }_{0}^{0.00}$ | 0.14 0.86 | 0.00 0.00 | 0.00 0.07 | ${ }_{0}^{0.00}$ | 0.007 0.07 | 0.00 | 0.00 | 0.00 | ${ }_{1.52}^{1.48}$ |
| 17 May as | 20:00:00 | 20:25:00 | 10.5000 | 11:4500 | 15:05:00 | 20 | $49^{220.803}$ | $122^{246.225}$ | North | Limmetic | 16 | 1 | 13 | 0 | ? |  |  |  |  |  | 25 | 150 | 786 | co | 0.32 | 297 | 1.39 | 0.13 | 0.00 | 0.00 |  |
|  |  |  |  | 1 weal | 1871 |  |  |  |  |  | ${ }_{\text {Mcan }}^{\text {Toual }}$ | ${ }_{3}^{28}$ | $\stackrel{137}{14}$ |  |  |  | ${ }_{3}^{27}$ | ${ }_{0}$ | 0 |  | 26 | 0.15 | 0.79 | 0.00 | 0.03 | 0.30 | 0.14 | 0.01 | 0.00 | 0.00 | 1.42 |
|  |  |  |  | Median | 16.78 | 15 |  |  |  |  | Mclian | 3 | 16 | $\cdots$ | \% | ${ }_{6}^{6}$ | ${ }^{2}$ | 0 | 10 | 10 | 28 10 | 0.15 10 | ${ }_{0}^{0.77} 10$ | 0.00 10 | 0.00 10 | 0.30 10 | ${ }_{0}^{0.07}$ | ${ }^{0.00} 10$ | ${ }^{0.00} 10$ | ${ }_{10}$ |  |





Stendiardived to tish pex hour

Anpendix 23. Sumpary of falcè using Eillusting sit Cuquidum Reservoii, Novermbor 2005.



| Suri Dute | Sear |  |  |  |  |  | ocution |  |  |  |  | Catch ${ }^{\text {a }}$ |  |  |  |  |  |  |  |  |  | crue ${ }^{\text {a }}$ |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Completaly Out Start haul in (hlt:mancss) (hh:massc) |  |  |  | $\underset{\substack{\text { Deph } \\(\mathrm{m})}}{ }$ | Nootting | Wesins | Resiun | Aren | Ste ${ }^{3}$ | Stet ${ }_{\text {No, }}$ ) | Cr | ко | co | su | NSC | PCC | RSS | sc | Total | ст | ко | co | su | nsc | PCC | RSs | sc | Towal |
| 31Oceves | 10:55:00 | 17:15:00 | 12:00.00 | 12:20:00 | 1905:00 | 2 | 49'22.148 | $122^{4472.261}$ | Souri | Lilurut | 4 | 1 | 0 | 0 | - | 0 | 2 | 0 | 0 | 0 | 2 | 0.00 | 0.00 | 0.00 | 0.00 | 0.03 | 0.00 | 0.90 | 0.00 | 0.05 0.05 |
| 31OCutbs | 16:30.00 | 16:45:00 | 13:50.00 | 14:00:00 | 21:17:30 | 2 | 4422.595 | $122^{4} 77.606$ | Soutl | Littoral | 8 | 2 | 0 | 0 | 0 | 0 | 2 | 0 | 0 | 1 | 3 | 0.00 | 0.00 | 0.00 | 0.00 | 0.03 | 0.010 | 0.00 | -002 | 0.05 0.26 |
| $31 / \mathrm{ctos}$ | 16:05:00 | 16:20:00 | 15:3000 | 15.45:00 | 23:25:00 | 2 | $4)^{23} 3.323$ | 122448.202 | Seuth | Lituoral | 7 | 3 | 0 | ${ }^{0}$ | 0 | 0 | $\bigcirc$ | 0 | 17 |  | ${ }^{18}$ | 0.00 | 0.00 | 0.00 | 0.08 | 0.80 | 0.00 | 0.06 | 0.04 | 0.26 |
| 2 Now 0 s | 00:20:00 | 00:25:30 | 15:50:00 | 16:00:00 | 1:323:30 | 2 | 44252.470 | $122^{4} 45.340$ | Centrul | Litural | 17 | 4 | 0 | 0 | 0 | 0 | 0 | 0 | 3 | 2 | 5 | 0.00 0.90 | 0.00 <br> 0.00 | 0.00 0.00 | 0.00 0.00 | 0.00 0.05 | 0.000 | 0.15 | 0.00 | 0.19 |
| 2 Now 0 S | 01:50:30 | 02:00:03 | 15:4000 | 15:45:08 | 13:47:30 | 2 | $45^{225} 5.520$ | ${ }^{122} 1244.4 .320$ | Central | Literal | 18 | 5 | ${ }_{0}^{0}$ | 0 | 0 | 0 | 3 | 0 | 1 | 0 | 4 | 0.00 | 0.00 | 0.00 | 0.00 | 0.08 | 0.00 | 3 | 0.00 | 0.11 |
| 2 Novas | 03:00:30 | 03:19:60 | 15:23:00 | 15:30:00 | 12:20:00 | 2 | $49^{22} 26.610$ |  |  | Litoral | 15 |  | 0 |  |  |  |  | 0 |  | 4 |  | 0.00 | 0.00 | 0.00 | 0.00 | 0.21 | 0.00 | 0.48 | 0.07 | 0.77 |
|  |  |  |  | Teai | 98.7 1545 | 2 |  |  |  | . | . | Mean | 0 | - | 0 | 0 | 2 | 0 | $s$ | 1 | 7 | 0.00 | 0.00 | 0.00 | 0.00 | 0.04 | 0.00 | 008 008 | 0.01 | 0.13 0.11 |
|  |  |  |  | Mcdian | 14.67 | 2 |  |  | - |  |  | ${ }_{\text {Mctian }}$ | $\bigcirc$ | $\stackrel{0}{6}$ | $\bigcirc$ | $\stackrel{8}{6}$ | $\frac{2}{1}$ | ${ }_{6}$ | ${ }_{6}$ | ${ }_{6}$ | $s$ | ${ }^{0.009}$ | ${ }^{0.00}$ | 0.00 | 0.00 | ${ }_{0}^{0.04}$ | ${ }^{600}$ | ${ }_{6}^{0.08}$ | ${ }_{6}$ | ${ }^{0.11}$ |




[^0]Appendix 26. Summary of catch, catch-per-unit-effort, and biomass of fish collected using gillnets at Coquitlam Reservoir, May 2005.

|  | Catch (No.) |  |  |  | Proportion of Catch (\%) |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | North | Central | South | Total | North | Central | South | Total |
| Cutthroat trout | 1 | 0 | 27 | 28 | 0\% | 0\% | 11\% | 11\% |
| Kokanee | 13 | 2 | 122 | 137 | 5\% | 1\% | 47\% | 53\% |
| Coho | 0 | 0 | 0 | 0 | 0\% | 0\% | 0\% | 0\% |
| Large scale sucker | 1 | 0 | 6 | 7 | 0\% | 0\% | 2\% | 3\% |
| Northern pikeminnow | 7 | 0 | 48 | 55 | 3\% | 0\% | 19\% | 21\% |
| Peamouth chub | 1 | 0 | 26 | 27 | 0\% | 0\% | 10\% | 11\% |
| Redside shiner | 0 | 0 | 3 | 3 | 0\% | 0\% | 1\% | 1\% |
| Sculpin | 0 | 0 | 0 | 0 | 0\% | $0 \%$ | 0\% | 0\% |
| Three-spine stickleback | 0 | 0 | 0 | 0 | 0\% | 0\% | 0\% | 0\% |
| Total | 23 | 2 | 232 | 257 | 9\% | 1\% | 90\% | 100\% |


|  | CPUE (No. $/ 90 \mathrm{~m}^{2} \bullet$ hr) |  |  |  | Proportion of CPUE (\%) |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | North | Central | South | Total | North | Central | South | Total |
| Cutthroat trout | 0.07 | 0.00 | 0.18 | 0.25 | 2\% | 0\% | 6\% | 8\% |
| Kokanee | 0.86 | 0.14 | 0.86 | 1.86 | 27\% | 4\% | 27\% | 58\% |
| Coho | 0.00 | 0.00 | 0.00 | 0.00 | 0\% | 0\% | 0\% | 0\% |
| Large scale sucker | 0.07 | 0.00 | 0.03 | 0.10 | 2\% | 0\% | 1\% | 3\% |
| Northern pikeminnow | 0.46 | 0.00 | 0.31 | 0.78 | 14\% | 0\% | 10\% | 24\% |
| Peamouth chub | 0.07 | 0.00 | 0.17 | 0.23 | 2\% | 0\% | 5\% | 7\% |
| Redside shiner | 0.00 | 0.00 | 0.02 | 0.02 | 0\% | 0\% | 0\% | 0\% |
| Sculpin | 0.00 | 0.00 | 0.00 | 0.00 | 0\% | 0\% | 0\% | 0\% |
| Three-spine stickleback | 0.00 | 0.00 | 0.00 | 0.00 | 0\% | 0\% | 0\% | 0\% |
| Total | 1.52 | 0.14 | 1.56 | 3.23 | 47\% | 4\% | 48\% | 100\% |


|  | Biomass (kg) |  |  |  | Proportion of Biomass (\%) |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | North | Central | South | Total | North | Central | South | Total |
| Cutthroat trout | 0.183 | 0.000 | 6.231 | 6.414 | 0\% | 0\% | 15\% | 15\% |
| Kokanee | 1.141 | 0.114 | 9.698 | 10.953 | 3\% | 0\% | 23\% | 26\% |
| Coho | 0.000 | 0.000 | 0.000 | 0.000 | 0\% | 0\% | 0\% | 0\% |
| Large scale sucker | 0.054 | 0.000 | 2.025 | 2.079 | 0\% | 0\% | 5\% | 5\% |
| Northern pikeminnow | 2.971 | 0.000 | 17.314 | 20.285 | 7\% | 0\% | 42\% | 49\% |
| Peamouth chub | 0.068 | 0.000 | 1.819 | 1.888 | 0\% | $0 \%$ | 4\% | 5\% |
| Redside shiner | 0.000 | 0.000 | 0.050 | 0.050 | 0\% | 0\% | 0\% | 0\% |
| Sculpin | 0.000 | 0.000 | 0.000 | 0.000 | 0\% | $0 \%$ | 0\% | 0\% |
| Three-spine stickleback | 0.000 | 0.000 | 0.000 | 0.000 | 0\% | $0 \%$ | 0\% | 0\% |
| Total | 4.418 | 0.114 | 37.137 | 41.669 | 11\% | 0\% | 89\% | 100\% |

Appendix 27. Summary of catch, catch-per-unit-effort, and biomass of fish collected using minnow traps at Coquitlam Reservoir, May 2005.

|  | Catch (No.) |  |  |  | Proportion of Catch (\%) |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | North | Central | South | Total | North | Central | South | Total |
| Cutthroat trout | 0 | 1 | 4 | 5 | 0\% | 1\% | 4\% | 4\% |
| Kokanee | 0 | 0 | 0 | 0 | 0\% | 0\% | 0\% | 0\% |
| Coho | 0 | 0 | 0 | 0 | 0\% | 0\% | 0\% | 0\% |
| Large scale sucker | 1 | 0 | 0 | 1 | 1\% | 0\% | 0\% | 1\% |
| Northern pikeminnow | 8 | 5 | 25 | 38 | 7\% | 4\% | 22\% | 34\% |
| Peamouth chub | 0 | 0 | 1 | 1 | 0\% | 0\% | 1\% | 1\% |
| Redside shiner | 1 | 0 | 29 | 30 | 1\% | 0\% | 26\% | 27\% |
| Sculpin | 0 | 1 | 6 | 7 | 0\% | 1\% | 5\% | 6\% |
| Three-spine stickleback | 1 | 5 | 24 | 30 | 1\% | 4\% | 21\% | 27\% |
| Total | 11 | 12 | 89 | 112 | 10\% | 11\% | $79 \%$ | 100\% |


|  | CPUE (No. / trap •hr) |  |  |  | Proportion of CPUE (\%) |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | North | Central | South | Total | North | Central | South | Total |
| Cutthroat trout | 0.00 | 0.01 | 0.01 | 0.02 | 0\% | 2\% | 1\% | 3\% |
| Kokanee | 0.00 | 0.00 | 0.00 | 0.00 | 0\% | 0\% | 0\% | 0\% |
| Coho | 0.00 | 0.00 | 0.00 | 0.00 | 0\% | 0\% | 0\% | 0\% |
| Large scale sucker | 0.03 | 0.00 | 0.00 | 0.03 | 5\% | 0\% | 0\% | 5\% |
| Northern pikeminnow | 0.27 | 0.06 | 0.06 | 0.38 | 40\% | 9\% | 9\% | 58\% |
| Peamouth chub | 0.00 | 0.00 | 0.00 | 0.00 | 0\% | 0\% | 0\% | 0\% |
| Redside shiner | 0.03 | 0.00 | 0.05 | 0.09 | 5\% | 0\% | 8\% | 13\% |
| Sculpin | 0.00 | 0.01 | 0.01 | 0.02 | 0\% | 2\% | 1\% | 3\% |
| Three-spine stickleback | 0.03 | 0.06 | 0.03 | 0.12 | 5\% | 9\% | 4\% | 18\% |
| Total | 0.36 | 0.14 | 0.15 | 0.66 | 56\% | 21\% | 23\% | 100\% |


|  | Biomass (kg) |  |  |  | Proportion of Biomass (\%) |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | North | Central | South | Total | North | Central | South | Total |
| Cutthroat trout | 0.000 | 0.010 | 0.055 | 0.065 | 0\% | 2\% | 12\% | 14\% |
| Kokanee | 0.000 | 0.000 | 0.000 | 0.000 | 0\% | 0\% | 0\% | 0\% |
| Coho | 0.003 | 0.000 | 0.000 | 0.003 | 1\% | 0\% | 0\% | 1\% |
| Large scale sucker | 0.003 | 0.000 | 0.000 | 0.003 | 1\% | 0\% | 0\% | 1\% |
| Northern pikeminnow | 0.036 | 0.041 | 0.133 | 0.211 | 8\% | 9\% | 29\% | 46\% |
| Peamouth chub | 0.000 | 0.000 | 0.001 | 0.001 | 0\% | 0\% | 0\% | 0\% |
| Redside shiner | 0.002 | 0.000 | 0.067 | 0.069 | 0\% | 0\% | 15\% | 15\% |
| Sculpin | 0.000 | 0.007 | 0.064 | 0.071 | 0\% | 1\% | 14\% | 16\% |
| Three-spine stickleback | 0.001 | 0.004 | 0.027 | 0.032 | 0\% | 1\% | 6\% | 7\% |
| Total | 0.044 | 0.063 | 0.346 | 0.453 | 10\% | 14\% | 76\% | 100\% |

Appendix 28. Summary of catch, catch-per-unit-effort, and biomass of fish collected using gillnets at Coquitlam Reservoir, November 2005.

|  | Catch (No.) |  |  |  | Proportion of Catch (\%) |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | North | Central | South | Total | North | Central | South | Total |
| Cutthroat trout | 1 | 15 | 9 | 25 | 0\% | 7\% | 4\% | 12\% |
| Kokanee | 4 | 45 | 44 | 93 | 2\% | 21\% | 20\% | 43\% |
| Coho | 1 | 0 | 0 | 1 | 0\% | 0\% | 0\% | 0\% |
| Large scale sucker | 0 | 11 | 1 | 12 | 0\% | 5\% | 0\% | 6\% |
| Northern pikeminnow | 0 | 24 | 6 | 30 | 0\% | 11\% | 3\% | 14\% |
| Peamouth chub | 8 | 40 | 4 | 52 | 4\% | 19\% | 2\% | 24\% |
| Redside shiner | 0 | 0 | 0 | 0 | 0\% | 0\% | 0\% | 0\% |
| Sculpin | 0 | 3 | 0 | 3 | 0\% | 1\% | 0\% | 1\% |
| Three-spine stickleback | 0 | 0 | 0 | 0 | 0\% | 0\% | 0\% | 0\% |
| Total | 14 | 138 | 64 | 216 | 6\% | 64\% | 30\% | 100\% |


|  | CPUE (No. $/ 90 \mathrm{~m}^{2} \bullet$ hr) |  |  |  | Proportion of CPUE (\%) |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | North | Central | South | Total | North | Central | South | Total |
| Cuthroat trout | 0.04 | 0.37 | 0.20 | 0.60 | 1\% | 9\% | 5\% | 15\% |
| Kokanee | 0.45 | 0.97 | 0.60 | 2.02 | 11\% | 24\% | 15\% | 50\% |
| Coho | 0.04 | 0.00 | 0.00 | 0.04 | 1\% | 0\% | 0\% | 1\% |
| Large scale sucker | 0.00 | 0.09 | 0.05 | 0.14 | 0\% | 2\% | 1\% | 3\% |
| Northern pikeminnow | 0.00 | 0.24 | 0.13 | 0.38 | 0\% | 6\% | 3\% | 9\% |
| Peamouth chub | 0.31 | 0.35 | 0.16 | 0.82 | 8\% | 9\% | 4\% | 20\% |
| Redside shiner | 0.00 | 0.00 | 0.00 | 0.00 | 0\% | 0\% | 0\% | 0\% |
| Sculpin | 0.00 | 0.04 | 0.00 | 0.04 | 0\% | 1\% | 0\% | 1\% |
| Three-spine stickleback | 0.00 | 0.00 | 0.00 | 0.00 | 0\% | 0\% | 0\% | 0\% |
| Total | 0.84 | 2.06 | 1.13 | 4.04 | 21\% | 51\% | 28\% | 100\% |


|  | Biomass (kg) |  |  |  | Proportion of Biomass (\%) |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | North | Central | South | Total | North | Central | South | Total |
| Cutthroat trout | 0.115 | 3.488 | 4.980 | 8.582 | 0\% | 10\% | 14\% | 25\% |
| Kokanee | 0.375 | 6.243 | 6.003 | 12.621 | 1\% | 18\% | 17\% | 36\% |
| Coho | 0.016 | 0.000 | 0.000 | 0.016 | 0\% | 0\% | 0\% | 0\% |
| Large scale sucker | 0.000 | 2.006 | 0.441 | 2.447 | 0\% | 6\% | 1\% | 7\% |
| Northern pikeminnow | 0.000 | 5.058 | 1.581 | 6.638 | 0\% | 14\% | 5\% | 19\% |
| Peamouth chub | 0.581 | 3.486 | 0.495 | 4.562 | 2\% | 10\% | 1\% | 13\% |
| Redside shiner | 0.000 | 0.000 | 0.000 | 0.000 | 0\% | 0\% | 0\% | 0\% |
| Sculpin | 0.000 | 0.139 | 0.000 | 0.139 | 0\% | 0\% | 0\% | 0\% |
| Three-spine stickleback | 0.000 | 0.000 | 0.000 | 0.000 | 0\% | 0\% | $0 \%$ | 0\% |
| Total | 1.086 | 20.419 | 13.499 | 35.005 | 3\% | 58\% | 39\% | 100\% |

Appendix 29. Summary of catch, catch-per-unit-effort, and biomass of fish collected using minnow traps at Coquitlam Reservoir, November 2005.

|  | Catch (No.) |  |  |  | Proportion of Catch (\%) |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | North | Central | South | Total | North | Central | South | Total |
| Cutthroat trout | 0 | 0 | 0 | 0 | 0\% | 0\% | 0\% | 0\% |
| Kokanee | 0 | 0 | 0 | 0 | 0\% | 0\% | 0\% | 0\% |
| Coho | 0 | 0 | 0 | 0 | 0\% | 0\% | 0\% | 0\% |
| Large scale sucker | 0 | 0 | 0 | 0 | 0\% | 0\% | 0\% | 0\% |
| Northern pikeminnow | 3 | 2 | 4 | 9 | 8\% | 5\% | 10\% | 23\% |
| Peamouth chub | 0 | 0 | 0 | 0 | 0\% | 0\% | 0\% | 0\% |
| Redside shiner | 1 | 8 | 17 | 26 | 3\% | 21\% | 44\% | 67\% |
| Sculpin | 0 | 3 | 1 | 4 | 0\% | 8\% | 3\% | 10\% |
| Three-spine stickleback | 0 | 0 | 0 | 0 | 0\% | 0\% | 0\% | 0\% |
| Total | 4 | 13 | 22 | 39 | 10\% | 33\% | 56\% | 100\% |


|  | CPUE (No. / trap •hr) |  |  |  | Proportion of CPUE (\%) |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | North | Central | South | Total | North | Central | South | Total |
| Cuthroat trout | 0.00 | 0.00 | 0.00 | 0.00 | 0\% | 0\% | 0\% | 0\% |
| Kokanee | 0.00 | 0.00 | 0.00 | 0.00 | 0\% | 0\% | 0\% | 0\% |
| Coho | 0.00 | 0.00 | 0.00 | 0.00 | 0\% | 0\% | 0\% | 0\% |
| Large scale sucker | 0.00 | 0.00 | 0.00 | 0.00 | 0\% | 0\% | 0\% | 0\% |
| Northern pikeminnow | 0.08 | 0.02 | 0.03 | 0.13 | 21\% | 6\% | 8\% | 35\% |
| Peamouth chub | 0.00 | 0.00 | 0.00 | 0.00 | 0\% | 0\% | 0\% | 0\% |
| Redside shiner | 0.03 | 0.10 | 0.08 | 0.21 | 7\% | 28\% | 21\% | 56\% |
| Sculpin | 0.00 | 0.02 | 0.01 | 0.03 | 0\% | 6\% | 3\% | 8\% |
| Three-spine stickleback | 0.00 | 0.00 | 0.00 | 0.00 | 0\% | 0\% | 0\% | 0\% |
| Total | 0.11 | 0.15 | 0.12 | 0.38 | 29\% | 40\% | 32\% | 100\% |


|  | Biomass (kg) |  |  |  | Proportion of Biomass (\%) |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | North | Central | South | Total | North | Central | South | Total |
| Cutthroat trout | 0.000 | 0.000 | 0.000 | 0.000 | 0\% | 0\% | 0\% | 0\% |
| Kokanee | 0.000 | 0.000 | 0.000 | 0.000 | 0\% | 0\% | 0\% | 0\% |
| Coho | 0.000 | 0.000 | 0.000 | 0.000 | 0\% | 0\% | 0\% | 0\% |
| Large scale sucker | 0.000 | 0.000 | 0.000 | 0.000 | 0\% | 0\% | 0\% | 0\% |
| Northern pikeminnow | 0.033 | 0.030 | 0.075 | 0.138 | 13\% | 12\% | 30\% | 55\% |
| Peamouth chub | 0.000 | 0.000 | 0.000 | 0.000 | 0\% | 0\% | 0\% | 0\% |
| Redside shiner | 0.006 | 0.025 | 0.052 | 0.083 | 2\% | 10\% | 21\% | 33\% |
| Sculpin | 0.000 | 0.029 | 0.000 | 0.029 | 0\% | 12\% | 0\% | 12\% |
| Three-spine stickleback | 0.000 | 0.000 | 0.000 | 0.000 | 0\% | 0\% | 0\% | 0\% |
| Total | 0.039 | 0.085 | 0.127 | 0.250 | 15\% | 34\% | 51\% | 100\% |







${ }^{2}$ Cear abbreviaioions: Gillnct (GN). Minnow Trap (MT), and Mid-Water Traw (TRWL)

Scx abbreviations: Malc (M), Female ( F ) , and Unknown (C)
Lifc Sage abbreviaions: Immature (1), Maturing (2), Maturc (3). Spauning (4), and Spent (5)
Capture type abbreviations: Enlangled by snoul (E), Gilled (G). Wedged (W), and Snagged by fin (S)
The distance from thic focus to the $7^{\text {th }}$ c circulus, and the distance from the focus to the first anaulus are provided with measurcments in mm at 100 power.
Relative slomach fultncss rating : Empty ( 0 ), $1-25 \%$ full ( 1 ) $26-50 \%$ full (2), $51-75 \%$ full ( 3 ). $76-100 \%$ full.






Gear abbrvivations: Gillnce (GN), Minnow Trap (MT), and Mid-Water Trawl (TRWL)

Scs abbreviations: Male (M). Fcmaie (F) and Unknown (U)
Life Stage abbrcuiations: Inwature (1) Maturing (2) Mature (3) Spawning (4) and Spent (5).
Thic distance from the focus 10 the $7^{\text {th }}$ circulus, and the distance from the focus to the first annuiss are provided with measurements in inn at 100 power
Rclative stomach fullicss rating: Emply (0). $1-25 \%$ fall ( 1 ). $26-50 \%$ futl ( 2 ) $.51-75 \%$ full (3), $76-100 \%$ fiul.
Proporional mass of gonad-to-somatic tissue

Appendix 32. Summary of biological characteristics of fishes, by sex, collected using gillnetting at Coquitlam Reservoir, May 2005.

${ }^{\text {a }}$ Includes fish classified as sex 'unknown'.

## Appendix 33. Summary of biological characteristics

of fishes collected using minnow trapping
at Coquitlam Reservoir, May 2005.

|  |  | CT | NCC | PCC | RSS | SC | TSSB |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Sample size (\#) |  | 5 | 30 | 1 | 29 | 7 | 29 |
| Length (mm) | Mean | 112.2 | 83.1 | 47.0 | 57.9 | 96.1 | 56.4 |
|  | StdDev | 22.2 | 13.7 | - | 18.5 | 21.9 | 9.1 |
|  | Min | 85.0 | 49.0 | 47.0 | 31.0 | 59.0 | 39.0 |
|  | Max | 135.0 | 118.0 | 47.0 | 103.0 | 125.0 | 69.0 |
|  |  |  |  |  |  |  |  |
| Mass (g) | Mean | 13.5 | 6.3 | 1.3 | 2.8 | 10.6 | 1.6 |
|  | StdDev | 7.2 | 2.8 | - | 3.4 | 6.6 | 0.6 |
|  | Min | 5.6 | 1.0 | 1.3 | 0.2 | 2.4 | 0.5 |
|  | Max | 21.9 | 15.4 | 1.3 | 12.5 | 21.6 | 2.6 |
|  |  |  |  |  |  |  |  |
| K-factor | Mean | 0.88 | 1.03 | 1.25 | 1.04 | 1.06 | 0.83 |
|  | StdDev | 0.04 | 0.14 | - | 0.16 | 0.11 | 0.14 |
|  | Min | 0.82 | 0.85 | 1.25 | 0.67 | 0.92 | 0.55 |
|  | Max | 0.91 | 1.52 | 1.25 | 1.43 | 1.23 | 1.29 |

Appendix 34. Summary of biological characteristics of fishes, by sex, collected using gillnetting at Coquitlam Reservoir, November 2005.

|  |  | CT |  |  | KO |  |  | CO |  |  |  | SU |  |  |  | NSC |  |  | PCC |  |  | SC |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | M | F | All | M | F | All | M | F |  | All | M | F |  | All | M | F | All | M | F | All | M | F |  | All |
| Sample size (\#) |  | 8 | 13 | 25 | 52 | 41 | 93 |  |  | - | 1 | 3 |  | - | 12 | 16 | 4 | 30 | 6 | 15 | 52 | 2 |  |  | 3 |
| Length (mm) | Mean | 286.1 | 324.2 | 301.0 | 224.7 | 223.3 | 224.1 |  |  | - | 203.0 | 326.7 |  | - | 253.3 | 234.2 | 365.0 | 248.1 | 175.7 | 207.9 | 189.1 | 166.0 |  | - | 152.7 |
|  | StdDev | 66.5 | 87.7 | 78.2 | 18.3 | 21.8 | 19.8 |  |  | - |  | 18.3 |  | - | 75.9 | 93.3 | 122.9 | 100.6 | 23.0 | 23.8 | 27.2 | 7.1 |  | - | 23.6 |
|  | Min | 215.0 | 226.0 | 193.0 | 145.0 | 128.0 | 128.0 |  |  | - | 203.0 | 306.0 |  | - | 167.0 | 121.0 | 211.0 | 118.0 | 153.0 | 161.0 | 106.0 | 161.0 |  |  | 126.0 |
|  | Max | 402.0 | 491.0 | 491.0 | 251.0 | 249.0 | 251.0 |  |  | - | 203.0 | 341.0 |  | - | 341.0 | 395.0 | 486.0 | 486.0 | 215.0 | 235.0 | 235.0 | 171.0 |  | - | 171.0 |
| Mass (g) | Mean | 286.0 | 431.2 | 343.3 | 138.0 | 132.8 | 135.7 |  |  | - | 16.3 | 447.3 |  | - | 203.9 | 184.8 | 502.4 | 221.3 | 65.3 | 115.2 | 87.7 | 57.5 |  | - | 46.4 |
|  | StdDev | 252.3 | 376.3 | 317.1 | 29.0 | 31.7 | 30.1 |  |  | - |  | 122.6 |  | - | 255.7 | 224.3 | 620.6 | 291.0 | 28.1 | 34.3 | 35.9 | 5.0 |  |  | 19.5 |
|  | Min | 94.3 | 109.2 | 70.3 | 45.3 | 26.9 | 26.9 |  |  | - | 16.3 | 328.0 |  | - | 51.3 | 17.6 | 99.7 | 16.4 | 42.0 | 51.2 | 13.9 | 53.9 |  | - | 24.2 |
|  | Max | 797.0 | 1223.0 | 1223.0 | 181.8 | 179.4 | 181.8 |  |  | - | 16.3 | 573.0 |  | - | 573.0 | 772.0 | 1410.0 | 1410.0 | 114.0 | 168.1 | 168.1 | 61.0 |  | - | 61.0 |
| K-factor | Mean | 1.00 | 1.03 | 1.00 | 1.20 | 1.17 | 1.19 |  |  | - | 0.19 | 1.27 |  | - | 1.09 | 1.08 | 1.01 | 1.08 | 1.15 | 1.25 | 1.22 | 1.26 |  | - |  |
|  | StdDev | 0.17 | 0.06 | 0.11 | 0.10 | 0.11 | 0.10 |  |  | - |  | 0.24 |  | - | 0.98 | 0.29 | 0.67 | 0.30 | 0.07 | 0.10 | 0.09 | 0.05 |  | - | 0.04 |
|  | Min | 0.64 | 0.94 | 0.64 | 1.00 | 0.95 | 0.95 |  |  | - | 0.19 | 1.11 |  | - | 0.24 | 0.12 | 0.09 | 0.09 | 1.03 | 1.04 | 1.03 | 1.22 |  | - |  |
|  | Max | 1.23 | 1.14 | 1.23 | 1.49 | 1.44 | 1.49 |  |  | - | 0.19 | 1.55 |  | - | 1.55 | 1.48 | 1.68 | 1.68 | 1.23 | 1.44 | 1.49 | 1.29 |  | $\cdot$ | 1.29 |

a Includes fish classified as sex 'unknown'.

```
Appendix 35. Summary of biological characteristics
    of fishes collected using minnow trapping
```

    at Coquitlam Reservoir, November 2005.
    |  |  | NSC | RSS | SC |
| :--- | ---: | ---: | ---: | ---: |
| Sample size (\#) |  | 9 | 26 | 3 |
| Length (mm) | Mean | 113.00 | 62.81 | 89.67 |
|  | StdDev | 13.96 | 11.67 | 23.18 |
|  | Min | 89.00 | 48.00 | 65.00 |
|  | Max | 132.00 | 98.00 | 111.00 |
| Mass (g) |  |  |  |  |
|  | Mean | 15.34 | 3.21 | 9.67 |
|  | StdDev | 6.44 | 2.24 | 6.51 |
|  | Min | 7.30 | 1.00 | 2.50 |
|  | Max | 25.60 | 12.10 | 15.20 |
| K-factor |  |  |  |  |
|  | Mean | 1.01 | 1.18 | - |
|  | StdDev | 0.15 | 0.28 | - |
|  | Min | 0.75 | 0.80 | - |
|  | Max | 1.18 | 1.95 | - |
|  |  |  |  |  |



Appendix 37 (a) May 2005


Appendix 37 (b) November 2005


| Depth (m) | Mean | Variance | Samplc Size | Stratum Arca <br> (ha) | Population Estimate | SE of Population Estimate | Lower 95\% Confidence Level | Upper $95 \%$ Confidence Level |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0.5 | 75 | 3,938 | 2 | 140 | 10,507 | 6,213 | - | 23,617 |
| 5-10 | 46 | 1,049 | 2 | 133 | 6,143 | 3,041 | - | 12,559 |
| 10-15 | 27 | 1,078 | 2 | 120 | 3.212 | 2.784 | - | 9,086 |
| 15-20 | 65 | 409 | 2 | 109 | 7.075 | 1,553 | 3,799 | 10,351 |
| 20-25 | 28 | 450 | 2 | 101 | 2,841 | 1,517 | - | 6,043 |
| 25-30 | 2 | 0 | 2 | 97 | 185 | 28 | 126 | 243 |
| 30-35 | 3 | 12 | 2 | 93 | 326 | 229 | - | 809 |
| 35-40 | 4 | 0 | 2 | 89 | 362 | 33 | 293 | 431 |
| 40-45 | 6 | 6 | 2 | 86 | 521 | 153 | 199 | 843 |
| 45-50 | 1 | 1 | 2 | 82 | 102 | 61 | - | 231 |
| 50-55 | 0 | 0 | 2 | 78 | 8 | 8 | - | 25 |

Appendix 38b. Estimated number of fish in the Central Basin at Coquitlam Reservoir, May 2005.

| Depth (m) | Mean | Variance | Sample Size | Stratum Area $\qquad$ | Population Estimate | SE of Population Estimate | Lower 95\% Confidence Level | Upper 95\% Confidence Level |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0.5 | 185 | 10,804 | 20 | 602 | 111,351 | 13,911 | 82,235 | 140,467 |
| 5-10 | 82 | 6,685 | 20 | 571 | 46,527 | 10,430 | 24.697 | 68,358 |
| 10-15 | 110 | 7,949 | 20 | 516 | 56,769 | 10,231 | 35,356 | 78,183 |
| 15-20 | 65 | 1,454 | 20 | 467 | 30,110 | 4,192 | 21,337 | 38,883 |
| 20-25 | 49 | 1,344 | 20 | 435 | 21,513 | 3,589 | 14,000 | 29,025 |
| 25-30 | 8 | 187 | 20 | 418 | 3,145 | 1,276 | 475 | 5,816 |
| 30-35 | 8 | 312 | 20 | 402 | 3,089 | 1,587 | - | 6,411 |
| 35-40 | 1 | 17 | 20 | 384 | 517 | 349 | - | 1,248 |
| 40-45 | 2 | 5 | 20 | 368 | 730 | 192 | 329 | 1,131 |
| 45-50 | 2 | 24 | 20 | 353 | 705 | 387 | - | 1,515 |
| 50-55 | 1 | 6 | 20 | 336 | 352 | 190 | - | 749 |
| 55-60 | 0 | 1 | 20 | 317 | 146 | 56 | 28 | 264 |
| 60-65 | 1 | 1 | 20 | 292 | 173 | 76 | 14 | 333 |
| 65-70 | 0 | 0 | 20 | 273 | 92 | 35 | 19 | 166 |
| 70-75 | 0 | 0 | 20 | 260 | 54 | 35 | - | 127 |
| 75-80 | 0 | 0 | 20 | 247 | 45 | 20 | 4 | 86 |
| 80-85 | 0 | 0 | 20 | 232 | 5 | 5 | - | 17 |
| 85-90 | 0 | - | 20 | 218 | - | - | - | - |
| 90-95 | 0 | - | 20 | 205 | - | - | - | - |
| 95-100 | 0 | - | 20 | 192 | - | - | - | - |
| 100-105 | 0 | - | 20 | 180 | - | $\bullet$ | - | - |
| 105-110 | 0 | - | 20 | 168 | - | - | - | $\cdot$ |

Appendix 38c. Estimated number of fish in the South Basin at Coquitlam Reservoir, May 2005.

| Depth (m) | Mcan | Variance | Sample Size | Stratum Area | Population Estimate | SE of Population Estimate | Lower 95\% Upper 95\% Confidence Confidence Level Level |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  | (ha) |  |  |  |  |
| 0-5 | 64 | 2.414 | 20 | 462 | 29,684 | 5,070 | 19,072 | 40,296 |
| 5-10 | 65 | 14,875 | 20 | 438 | 28,471 | 11,934 | 3,494 | 53,448 |
| 10-15 | 45 | 2,506 | 20 | 395 | 17,768 | 4,425 | 8,507 | 27,030 |
| 15-20 | 26 | 1,488 | 20 | 358 | 9,312 | 3,087 | 2.850 | 15,773 |
| 20-25 | 27 | 1,944 | 20 | 333 | 8,923 | 3,288 | 2,041 | 15,805 |
| 25-30 | 6 | 160 | 20 | 320 | 2,011 | 907 | 113 | 3,909 |
| 30-35 | 0 | 0 | 20 | 308 | 108 | 43 | 17 | 199 |
| 35-40 | 2 | 35 | 20 | 295 | 458 | 392 | - | 1,278 |
| 40-45 | 2 | 13 | 16 | 282 | 44 I | 258 | - | 982 |
| 45-50 | 1 | 5 | 16 | 271 | 157 | 156 | - | 484 |
| 50-55 | 0 | 0 | 16 | 258 | 6 | 4 | - | 14 |
| 55-60 | 0 | 0 | 14 | 243 | 46 | 28 | - | 106 |
| 60-65 | 1 | 4 | 14 | 224 | 122 | 122 | - | 376 |
| 65-70 | 0 | 0 | 14 | 209 | 2 | 2 | - | 6 |
| 70-75 | 0 | - | 14 | 199 | - | - | - | 5 |
| 75-80 | 0 | 1 | 11 | 189 | 63 | 44 | - | 155 |
| 80-85 | 0 | - | 11 | 178 | - | - | - | - |
| 85-90 | 0 | - | 11 | 167 | - | - | - | - |
| 90-95 | 0 | - | 11 | 157 | - | - | - | - |
| 95-100 | 0 | - | 11 | 147 | - | - | - | - |
| 100-105 | 0 | - | 11 | 138 | - | - | - | - |
| 105-110 | 0 | - | 11 | 129 | - | - | - | - |

Appendix 38d. Estimated number of fish at Coquitlam Reservoir, May 2005.

| Region | Population Estimate | SE of <br> Population <br> Estimatc | Lower $95 \%$ Confidence Level | Upper 95\% Confidence Level |
| :---: | :---: | :---: | :---: | :---: |
| North | 31,281 | 7,772 | 15.069 | 47,493 |
| Central | 275,323 | 21,022 | 228,485 | 322,161 |
| South | 97,573 | 14.461 | 65,353 | 129,792 |
| Whole Lake | 404,177 | 26,673 | 348,136 | 460,218 |

Appendix 39 a. Estimated number of fish in the North Basin at Coquitlam Reservoir, Novenber 2005.

| Depth (m) | Mean | Variance | Sample Size | Stratum Area <br> (ha) | Population Estimate | SE of <br> Population Estimate | Lower 95\% Confidence Level | Upper 95\% Confidence Level |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0-5 | 19 | - | 1 | 140 | 2,713 | - | - |  |
| 5-10 | 40 | - | 1 | 133 | 5,285 | - | - | - |
| 10-15 | 16 | - | 1 | 120 | 1,885 | - | - | - |
| 15-20 | 23 | - | 1 | 109 | 2,487 | - | - | - |
| 20.25 | 12 | - | 1 | 101 | 1,219 | - | - | - |
| 25-30 | 2 | - | 1 | 97 | 234 | - | - | - |
| 30-35 | 1 | - | 1 | 93 | 69 | - | - | - |
| 35-40 | 0 | $\cdot$ | 1 | 89 | - | * | - | - |
| 40.45 | 0 | - | 1 | 86 | - | - | - | - |
| $45 \cdot 50$ | 0 | - | 1 | 82 | 0 | - | - | - |
| 50-55 | 0 | - | 1 | 78 | 1 | - | - | - |

Appendix 39 b Estimated number of fish in the Central Basin at Coquitlam Reservoir, November2005.

| Depth (m) | Mean | Variance | Sample Size | Stratum Area (ha) | Population Estimate | SE of Population Estimate | Lower 95\% Confidence Leve! | Upper 95\% Confidence Level |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0-5 | 56 | 6,571 | 10 | 602 | 33,806 | 15,515 | 1,333 | 66,280 |
| 5-10 | 6 | 44 | 10 | 571 | 3,391 | 1,533 | 183 | 6,600 |
| 10-15 | 6 | 44 | 10 | 516 | 2,964 | 1,724 | - | 6,571 |
| 15-20 | 42 | 573 | 10 | 467 | 19,571 | 3,522 | 12,199 | 26,943 |
| 20-25 | 54 | 764 | 10 | 435 | 23,503 | 4,510 | 14,064 | 32,942 |
| 25-30 | 14 | 54 | 10 | 418 | 5,728 | 978 | 3,681 | 7,775 |
| 30-35 | 14 | 759 | 10 | 402 | 5,716 | 3,497 | - | 13,036 |
| 35-40 | 0 | 0 | 10 | 384 | 72 | 55 | - | 187 |
| 40-45 | 0 | 0 | 10 | 368 | 78 | 65 | - | 215 |
| 45-50 | 0 | 0 | 10 | 353 | 82 | 53 | - | 193 |
| 50-55 | 0 | 0 | 10 | 336 | 77 | 42 | - | 165 |
| 55-60 | 0 | 1 | 10 | 317 | 143 | 89 | - | 330 |
| 60-65 | 0 | 0 | 10 | 292 | 18 | 6 | 6 | 3 I |
| 65-70 | 0 | 0 | 10 | 273 | 9 | 6 | - | 22 |
| 70-75 | 0 | 0 | 10 | 260 | 39 | 31 | - | 104 |
| 75.80 | 0 | 0 | 10 | 247 | 10 | 7 | - | 24 |
| 80-85 | 0 | 0 | 10 | 232 | 6 | 5 | - | 16 |
| 85-90 | 0 | - | 10 | 218 | - | - | - | - |
| 90-95 | 0 | - | 10 | 205 | - | - | - | - |
| 95-100 | 0 | - | 10 | 192 | - | - | - | - |
| 100-105 | 0 | - | 10 | 180 | - | - | - | - |
| 105-110 | 0 | - | 10 | 168 | - | - | - | - |

Appendix 39 c . Estimated number of fish in the South Basin at Coquitlam Reservoir, November 2005.

| Depth (m) | Mean | Variance | Sample Size | Stratum Area <br> (ha) | Population Estimate | SE of <br> Population Estimate | Lower 95\% Confidence Level | Upper $95 \%$ Confidence Level |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0.5 | 145 | 28,613 | 10 | 462 | 66,991 | 33,289 | - | 136,664 |
| 5-10 | 9 | 63 | 10 | 438 | 4,005 | 1,729 | 387 | 7,624 |
| 10-15 | 7 | 115 | 10 | 395 | 2,633 | 1,603 | - | 5,988 |
| 15-20 | 13 | 196 | 10 | 358 | 4,517 | 2,186 | * | 9,092 |
| 20-25 | 13 | 117 | 10 | 333 | 4,355 | 1,847 | 488 | 8,221 |
| 25-30 | 3 | 15 | 10 | 320 | 959 | 509 | - | 2,024 |
| 30-35 | 6 | 95 | 10 | 308 | 1,803 | 1,124 | - | 4,156 |
| 35-40 | 0 | 0 | 10 | 295 | 49 | 34 | - | 121 |
| 40-45 | 1 | 1 | 8 | 282 | 161 | 132 | - | 438 |
| 45-50 | 0 | 0 | 8 | 271 | 3 | 2 | - | 8 |
| 50-55 | 0 | 0 | 8 | 258 | 0 | 0 | - | 1 |
| 55-60 | 0 | - | 7 | 243 | - | - | - | - |
| 60.65 | 0 | 0 | 7 | 224 | 20 | 23 | - | 68 |
| 65-70 | 0 | 0 | 7 | 209 | 0 | 0 | - | 1 |
| 70-75 | 0 | - | 7 | 199 | - | - | - | - |
| 75-80 | 0 | 0 | 6 | 189 | 2 | 2 | - | 6 |
| 80.85 | 0 | - | 6 | 178 | - | - | - | - |
| 85.90 | 0 | - | 6 | 167 | - | - | - | - |
| 90.95 | 0 | - | 6 | 157 | - | - | - | - |
| 95-100 | 0 | - | 6 | 147 | - | - | - | - |
| 100-105 | 0 | - | 6 | 138 | - | - | - | - |
| 105-110 | 0 | - | 6 | 129 | - | $\cdot$ | - | - |

Appendix 39 d Estimated number of fish at Coquitlam Reservoir, November 2005

| Region | Population Estimate | SE of Population Estimate | Lower 95\% Confidence Level | Upper 95\% Confidence Level |
| :---: | :---: | :---: | :---: | :---: |
| North | 13,893 | - | - | - |
| Central | 95,213 | 17,088 | 57,141 | 133,285 |
| South | 85,498 | 33,517 | 10,821 | 160,175 |
| Whole Lake | 194,604 | 37,622 | 115,560 | 273,648 |

Appendix 40. Distribution of population and biomass, by size, using hydroacoustics at Coquitlam Reservoir, May 2005.

| Length (mm) | \% tracked | Predicted weight (g) | Population Estimate |  |  | Biomass Estimate (kg) |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | Expected | Lower 95\% | Upper 95\% | Expected | Lower 95\% | Upper 95\% |
| 10 | 1.8\% | 0.01 | 7,411 | 6,384 | 8,439 | 0.1 | 0.1 | 0.1 |
| 20 | 22.6\% | 0.1 | 91,446 | 78,766 | 104,125 | 5.4 | 4.7 | 6.2 |
| 30 | 19.8\% | 0.2 | 80,078 | 68,974 | 91,181 | 15.4 | 13.3 | 17.5 |
| 40 | 12.7\% | 0.4 | 51,156 | 44,063 | 58,249 | 22.6 | 19.5 | 25.8 |
| 50 | 7.4\% | 0.8 | 30,092 | 25,919 | 34,264 | 25.4 | 21.9 | 29.0 |
| 60 | 4.2\% | 1.4 | 16,996 | 14,640 | 19,353 | 24.4 | 21.0 | 27.8 |
| 70 | 3.3\% | 2.2 | 13,263 | 11,424 | 15,102 | 29.7 | 25.6 | 33.9 |
| 80 | 2.3\% | 3.3 | 9,362 | 8,064 | 10,660 | 30.9 | 26.6 | 35.2 |
| 90 | 1.8\% | 4.6 | 7,077 | 6,096 | 8,058 | 32.9 | 28.3 | 37.5 |
| 100 | 1.6\% | 6.3 | 6,408 | 5,520 | 7,297 | 40.4 | 34.8 | 46.0 |
| 110 | 1.4\% | 8.3 | 5,517 | 4,752 | 6,282 | 45.9 | 39.5 | 52.3 |
| 120 | 0.9\% | 10.7 | 3,622 | 3,120 | 4,124 | 38.8 | 33.4 | 44.2 |
| 130 | 1.0\% | 13.5 | 3,957 | 3,408 | 4,505 | 53.4 | 46.0 | 60.8 |
| 140 | 0.7\% | 16.7 | 2,786 | 2,400 | 3,173 | 46.6 | 40.2 | 53.1 |
| 150 | 0.8\% | 20.4 | 3,176 | 2,736 | 3,617 | 65.0 | 55.9 | 74.0 |
| 160 | 0.6\% | 24.7 | 2,452 | 2,112 | 2,792 | 60.5 | 52.1 | 68.8 |
| 170 | 0.6\% | 29.4 | 2,508 | 2,160 | 2,855 | 73.7 | 63.5 | 83.9 |
| 180 | 0.6\% | 34.7 | 2,508 | 2,160 | 2,855 | 87.0 | 74.9 | 99.1 |
| 190 | 0.8\% | 40.6 | 3,065 | 2,640 | 3,490 | 124.4 | 107.1 | 141.6 |
| 200 | 0.6\% | 47.1 | 2,452 | 2,112 | 2,792 | 115.5 | 99.5 | 131.5 |
| 210 | 0.8\% | 54.3 | 3,288 | 2,832 | 3,744 | 178.4 | 153.6 | 203.1 |
| 220 | 0.6\% | 62.1 | 2,396 | 2,064 | 2,728 | 148.8 | 128.2 | 169.4 |
| 230 | 0.7\% | 70.6 | 2,898 | 2,496 | 3,300 | 204.7 | 176.3 | 233.1 |
| 240 | 0.8\% | 79.9 | 3,065 | 2,640 | 3,490 | 244.9 | 211.0 | 278.9 |
| 250 | 0.6\% | 90.0 | 2,285 | 1,968 | 2,602 | 205.5 | 177.0 | 234.0 |
| 260 | 0.6\% | 100.8 | 2,285 | 1,968 | 2,602 | 230.3 | 198.4 | 262.2 |
| 270 | 0.8\% | 112.4 | 3,288 | 2,832 | 3,744 | 369.7 | 318.4 | 421.0 |
| 280 | 0.7\% | 125.0 | 2,675 | 2,304 | 3,046 | 334.2 | 287.9 | 380.6 |
| 290 | 0.7\% | 138.3 | 2,675 | 2,304 | 3,046 | 370.0 | 318.7 | 421.3 |
| 300 | 0.6\% | 152.6 | 2,619 | 2,256 | 2,982 | 399.8 | 344.3 | 455.2 |
| 310 | 0.7\% | 167.9 | 2,842 | 2,448 | 3,236 | 477.1 | 410.9 | 543.2 |
| 320 | 0.5\% | 184.0 | 1,839 | 1,584 | 2,094 | 338.5 | 291.5 | 385.4 |
| 330 | 0.6\% | 201.2 | 2,508 | 2,160 | 2,855 | 504.6 | 434.6 | 574.6 |
| 340 | 0.5\% | 219.4 | 1,950 | 1,680 | 2,221 | 428.0 | 368.6 | 487.3 |
| 350 | 0.5\% | 238.7 | 1,839 | 1,584 | 2,094 | 438.9 | 378.0 | 499.8 |
| 360 | 0.5\% | 259.0 | 1,895 | 1,632 | 2,157 | 490.7 | 422.7 | 558.7 |
| 370 | 0.4\% | 280.4 | 1,560 | 1,344 | 1,777 | 437.5 | 376.9 | 498.2 |
| 380 | 0.3\% | 303.0 | 1,226 | 1,056 | 1,396 | 371.4 | 319.9 | 422.9 |
| 390 | 0.2\% | 326.7 | 947 | 816 | 1,079 | 309.4 | 266.5 | 352.4 |
| 400 | 0.3\% | 351.5 | 1,059 | 912 | 1,206 | 372.2 | 320.6 | 423.8 |
| 410 | 0.3\% | 377.6 | 1,059 | 912 | 1,206 | 399.8 | 344.4 | 455.3 |
| 420 | 0.3\% | 405.0 | 1,059 | 912 | 1,206 | 428.8 | 369.3 | 488.2 |
| 430 | 0.2\% | 433.6 | 669 | 576 | 761 | 289.9 | 249.7 | 330.1 |
| 440 | 0.2\% | 463.5 | 1,003 | 864 | 1,142 | 464.9 | 400.4 | 529.3 |
| 450 | 0.2\% | 494.7 | 724 | 624 | 825 | 358.4 | 308.7 | 408.0 |
| 460 | 0.2\% | 527.2 | 780 | 672 | 888 | 411.3 | 354.3 | 468.4 |
| 470 | 0.2\% | 561.2 | 669 | 576 | 761 | 375.2 | 323.2 | 427.3 |
| 480 | 0.1\% | 596.5 | 446 | 384 | 508 | 265.9 | 229.0 | 302.8 |
| 490 | 0.2\% | 633.2 | 836 | 720 | 952 | 529.3 | 455.9 | 602.7 |
| 500 | 0.1\% | 671.4 | 502 | 432 | 571 | 336.8 | 290.1 | 383.4 |
| 510 | 0.2\% | 711.1 | 669 | 576 | 761 | 475.5 | 409.6 | 541.5 |
| 520 | 0.2\% | 752.3 | 613 | 528 | 698 | 461.2 | 397.2 | 525.1 |
| 530 | 0.1\% | 795.1 | 279 | 240 | 317 | 221.5 | 190.8 | 252.2 |
| 540 | 0.1\% | 839.3 | 390 | 336 | 444 | 327.4 | 282.0 | 372.8 |
| 550 | 0.1\% | 885.2 | 557 | 480 | 635 | 493.3 | 424.9 | 561.7 |
| 560 | 0.1\% | 932.7 | 279 | 240 | 317 | 259.9 | 223.8 | 295.9 |
| 570 | 0.2\% | 981.8 | 613 | 528 | 698 | 601.8 | 518.4 | 685.3 |
| 580 | 0.1\% | 1032.6 | 446 | 384 | 508 | 460.3 | 396.5 | 524.2 |
| 590 | 0.1\% | 1085.1 | 279 | 240 | 317 | 302.3 | 260.4 | 344.3 |
| 600 | 0.0\% | 1139.3 | 167 | 144 | 190 | 190.5 | 164.1 | 216.9 |
| 610 | 0.0\% | 1195.2 | 167 | 144 | 190 | 199.8 | 172.1 | 227.5 |
| 620 | 0.0\% | 1253.0 | 111 | 96 | 127 | 139.6 | 120.3 | 159.0 |
| 630 | 0.0\% | 1312.5 | 167 | 144 | 190 | 219.4 | 189.0 | 249.8 |
| 640 | 0.1\% | 1373.8 | 279 | 240 | 317 | 382.8 | 329.7 | 435.9 |
| 650 | 0.0\% | 1437.0 | 167 | 144 | 190 | 240.2 | 206.9 | 273.5 |
| 660 | 0.0\% | 1502.0 | 167 | 144 | 190 | 251.1 | 216.3 | 285.9 |
| 670 | 0.0\% | 1569.0 | 111 | 96 | 127 | 174.9 | 150.6 | 199.1 |
| 680 | 0.0\% | 1637.9 | 167 | 144 | 190 | 273.8 | 235.8 | 311.8 |
| 690 | 0.0\% | 1708.7 | 56 | 48 | 63 | 95.2 | 82.0 | 108.4 |
| 700 | 0.0\% | 1781.5 | . | - | - | 0.0 | 0.0 | 0.0 |
| 710 | 0.0\% | 1856.3 | 56 | 48 | 63 | 103.4 | 89.1 | 117.8 |
| 720 | 0.0\% | 1933.1 | 56 | 48 | 63 | 107.7 | 92.8 | 122.7 |
| 730 | 0.0\% | 2012.0 | - | - | - | 0.0 | 0.0 | 0.0 |
| 740 | 0.0\% | 2093.0 | 56 | 48 | 63 | 116.6 | 100.5 | 132.8 |
| 750 | 0.0\% | 2176.1 | 56 | 48 | 63 | 121.3 | 104.5 | 138.1 |
| 760 | 0.0\% | 2261.3 | - | - | - | 0.0 | 0.0 | 0.0 |
| 770 | 0.0\% | 2348.7 | - | - | - | 0.0 | 0.0 | 0.0 |
| 780 | 0.0\% | 2438.2 | - | - | - | 0.0 | 0.0 | 0.0 |
| 790 | 0.0\% | 2530.0 | 56 | 48 | 63 | 141.0 | 121.4 | 160.5 |
| 800 | 0.0\% | 2624.0 | . | - | - | 0.0 | 0.0 | 0.0 |
|  | 100.0\% |  | 404,177 | 348,136 | 460,218 | 18,014 | 15,516 | 20,511 |

Appendix 41. Distribution of population and biomass, by size, using hydroacoustics at Coquitlam Reservoir, November 2005.

| Length (mm) | \% tracked | Predicted weight (g) | Population Estimate |  |  | Biomass Estimate (kg) |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | Expected | Lower 95\% | Upper 95\% | Expected | Lower 95\% | Upper 95\% |
| 10 | 0.4\% | 0.01 | 855 | 508 | 1,202 | 0.0 | 0.0 | 0.0 |
| 20 | 6.6\% | 0.1 | 12,876 | 7,646 | 18,106 | 0.8 | 0.5 | 1.1 |
| 30 | 6.8\% | 0.2 | 13,178 | 7,825 | 18,531 | 2.5 | 1.5 | 3.6 |
| 40 | 5.1\% | 0.4 | 9,858 | 5,854 | 13,863 | 4.4 | 2.6 | 6.1 |
| 50 | 3.6\% | 0.8 | 6,991 | 4,152 | 9,831 | 5.9 | 3.5 | 8.3 |
| 60 | 2.8\% | 1.4 | 5,432 | 3,226 | 7,639 | 7.8 | 4.6 | 11.0 |
| 70 | 2.5\% | 2.2 | 4,929 | 2,927 | 6,931 | 11.1 | 6.6 | 15.5 |
| 80 | 2.1\% | 3.3 | 4,074 | 2,419 | 5,729 | 13.5 | 8.0 | 18.9 |
| 90 | 1.8\% | 4.6 | 3,571 | 2,121 | 5,022 | 16.6 | 9.9 | 23.3 |
| 100 | 1.6\% | 6.3 | 3,018 | 1,792 | 4,244 | 19.0 | 11.3 | 26.8 |
| 110 | 1.6\% | 8.3 | 3,068 | 1,822 | 4,314 | 25.5 | 15.2 | 35.9 |
| 120 | 1.3\% | 10.7 | 2,616 | 1,553 | 3,678 | 28.0 | 16.6 | 39.4 |
| 130 | 1.4\% | 13.5 | 2,716 | 1,613 | 3,819 | 36.7 | 21.8 | 51.6 |
| 140 | 1.1\% | 16.7 | 2,213 | 1,314 | 3,112 | 37.0 | 22.0 | 52.1 |
| 150 | 1.2\% | 20.4 | 2,314 | 1,374 | 3,254 | 47.3 | 28.1 | 66.5 |
| 160 | 1.6\% | 24.7 | 3,018 | 1,792 | 4,244 | 74.4 | 44.2 | 104.6 |
| 170 | 0.9\% | 29.4 | 1,811 | 1,075 | 2,546 | 53.2 | 31.6 | 74.9 |
| 180 | 1.3\% | 34.7 | 2,565 | 1,523 | 3,607 | 89.0 | 52.9 | 125.2 |
| 190 | 1.1\% | 40.6 | 2,163 | 1,284 | 3,041 | 87.8 | 52.1 | 123.4 |
| 200 | 1.2\% | 47.1 | 2,364 | 1,404 | 3,324 | 111.3 | 66.1 | 156.6 |
| 210 | 1.4\% | 54.3 | 2,766 | 1,643 | 3,890 | 150.1 | 89.1 | 211.1 |
| 220 | 2.1\% | 62.1 | 4,175 | 2,479 | 5,870 | 259.2 | 153.9 | 364.5 |
| 230 | 1.6\% | 70.6 | 3,118 | 1,852 | 4,385 | 220.3 | 130.8 | 309.7 |
| 240 | 1.9\% | 79.9 | 3,672 | 2,180 | 5,163 | 293.4 | 174.2 | 412.6 |
| 250 | 1.6\% | 90.0 | 3,169 | 1,882 | 4,456 | 285.0 | 169.3 | 400.8 |
| 260 | 1.6\% | 100.8 | 3,118 | 1,852 | 4,385 | 314.3 | 186.6 | 442.0 |
| 270 | 2.1\% | 112.4 | 4,074 | 2,419 | 5,729 | 458.1 | 272.1 | 644.2 |
| 280 | 2.3\% | 125.0 | 4,426 | 2,628 | 6,224 | 553.1 | 328.4 | 777.7 |
| 290 | 1.8\% | 138.3 | 3,521 | 2,091 | 4,951 | 487.1 | 289.2 | 684.9 |
| 300 | 2.1\% | 152.6 | 4,124 | 2,449 | 5,800 | 629.5 | 373.8 | 885.2 |
| 310 | 2.3\% | 167.9 | 4,477 | 2,658 | 6,295 | 751.4 | 446.2 | 1056.7 |
| 320 | 2.1\% | 184.0 | 4,175 | 2,479 | 5,870 | 768.4 | 456.3 | 1080.5 |
| 330 | 2.5\% | 201.2 | 4,879 | 2,897 | 6,861 | 981.8 | 583.0 | 1380.6 |
| 340 | 1.7\% | 219.4 | 3,269 | 1,941 | 4,597 | 717.4 | 426.0 | 1008.8 |
| 350 | 2.0\% | 238.7 | 3,873 | 2,300 | 5,446 | 924.4 | 548.9 | 1299.8 |
| 360 | 1.7\% | 259.0 | 3,370 | 2,001 | 4,739 | 872.8 | 518.3 | 1227.3 |
| 370 | 1.5\% | 280.4 | 2,867 | 1,702 | 4,032 | 803.9 | 477.4 | 1130.5 |
| 380 | 1.6\% | 303.0 | 3,118 | 1,852 | 4,385 | 944.8 | 561.0 | 1328.5 |
| 390 | 1.6\% | 326.7 | 3,118 | 1,852 | 4,385 | 1018.7 | 604.9 | 1432.4 |
| 400 | 1.3\% | 351.5 | 2,515 | 1,493 | 3,536 | 884.1 | 525.0 | 1243.2 |
| 410 | 1.4\% | 377.6 | 2,666 | 1,583 | 3,749 | 1006.7 | 597.8 | 1415.6 |
| 420 | 1.1\% | 405.0 | 2,163 | 1,284 | 3,041 | 875.9 | 520.1 | 1231.6 |
| 430 | 1.2\% | 433.6 | 2,263 | 1,344 | 3,183 | 981.4 | 582.7 | 1380.0 |
| 440 | 1.0\% | 463.5 | 1,962 | 1,165 | 2,758 | 909.1 | 539.9 | 1278.4 |
| 450 | 0.7\% | 494.7 | 1,308 | 777 | 1,839 | 646.9 | 384.1 | 909.7 |
| 460 | 0.6\% | 527.2 | 1,107 | 657 | 1,556 | 583.4 | 346.4 | 820.4 |
| 470 | 0.8\% | 561.2 | 1,509 | 896 | 2,122 | 846.8 | 502.8 | 1190.7 |
| 480 | 0.6\% | 596.5 | 1,107 | 657 | 1,556 | 660.0 | 392.0 | 928.1 |
| 490 | 0.6\% | 633.2 | 1,207 | 717 | 1,697 | 764.4 | 453.9 | 1074.9 |
| 500 | 0.4\% | 671.4 | 855 | 508 | 1,202 | 574.1 | 340.9 | 807.3 |
| 510 | 0.8\% | 711.1 | 1,509 | 896 | 2,122 | 1073.1 | 637.2 | 1508.9 |
| 520 | 0.4\% | 752.3 | 704 | 418 | 990 | 529.8 | 314.6 | 745.0 |
| 530 | 0.6\% | 795.1 | 1,107 | 657 | 1,556 | 879.8 | 522.4 | 1237.1 |
| 540 | 0.5\% | 839.3 | 1,056 | 627 | 1,485 | 886.6 | 526.5 | 1246.7 |
| 550 | 0.5\% | 885.2 | 1,006 | 597 | 1,415 | 890.5 | 528.8 | 1252.2 |
| 560 | 0.3\% | 932.7 | 604 | 358 | 849 | 563.0 | 334.3 | 791.6 |
| 570 | 0.5\% | 981.8 | 956 | 567 | 1,344 | 938.3 | 557.2 | 1319.4 |
| 580 | 0.4\% | 1032.6 | 855 | 508 | 1,202 | 883.0 | 524.3 | 1241.6 |
| 590 | 0.5\% | 1085.1 | 956 | 567 | 1,344 | 1037.0 | 615.8 | 1458.2 |
| 600 | 0.3\% | 1139.3 | 654 | 388 | 919 | 745.0 | 442.4 | 1047.6 |
| 610 | 0.3\% | 1195.2 | 654 | 388 | 919 | 781.5 | 464.1 | 1099.0 |
| 620 | 0.4\% | 1253.0 | 704 | 418 | 990 | 882.3 | 523.9 | 1240.7 |
| 630 | 0.3\% | 1312.5 | 654 | 388 | 919 | 858.2 | 509.6 | 12068 |
| 640 | 0.3\% | 1373.8 | 654 | 388 | 919 | 898.3 | 533.4 | 1263.2 |
| 650 | 0.2\% | 1437.0 | 402 | 239 | 566 | 578.2 | 343.4 | 813.1 |
| 660 | 0.2\% | 1502.0 | 453 | 269 | 637 | 679.9 | 403.8 | 956.1 |
| 670 | 0.1\% | 1569.0 | 251 | 149 | 354 | 394.6 | 234.3 | 554.9 |
| 680 | 0.3\% | 1637.9 | 553 | 329 | 778 | 906.2 | 538.1 | 1274.3 |
| 690 | 0.1\% | 1708.7 | 251 | 149 | 354 | 429.7 | 255.2 | 604.3 |
| 700 | 0.2\% | 1781.5 | 402 | 239 | 566 | 716.8 | 425.7 | 1008.0 |
| 710 | 0.2\% | 1856.3 | 453 | 269 | 637 | 840.3 | 499.0 | 1181.6 |
| 720 | 0.2\% | 1933.1 | 453 | 269 | 637 | 875.1 | 519.7 | 1230.6 |
| 730 | 0.1\% | \% 2012.0 | 101 | 60 | 141 | 202.4 | 120.2 | 284.6 |
| 740 | 0.1\% | \% 2093.0 | 201 | 119 | 283 | 421.1 | 250.1 | 592.1 |
| 750 | 0.1\% | 2176.1 | 151 | 90 | 212 | 328.4 | 195.0 | 461.7 |
| 760 | 0.1\% | \% 22613 | 101 | 60 | 141 | 227.5 | 135.1 | 319.9 |
| 770 | 0.1\% | \% 2348.7 | 101 | 60 | 141 | 236.3 | -140.3 | 332.2 |
| 780 | 0.1\% | \% 2438.2 | 101 | 60 | 141 | 245.3 | 145.7 | 344.9 |
| 790 | 0.1\% | \% 2530.0 | 151 | 90 | 212 | 381.8 | - 226.7 | 536.8 |
| 800 | 0.4\% | \% 2624.0 | 805 | 478 | 1,132 | 2111.7 | $7 \quad 1254.0$ | 2969.4 |
|  | 100.0\% |  | 194,604 | 115,560 | 273,648 | 42,260 | 25,095 | 59,425 |

Appendix 42. Population, biomass, and production estimates of kokanee, by age and region, at Coquitlam Reservoir, May 2005.

| Age | Region | Population (\#) |  |  |  | Biomass (kg) |  |  |  | Production (\#/ ha) |  |  |  | Standing Crop (kg / ha) |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | Lower 95\% |  | Upper 95\% <br> C.I | Lower Upper 95\% |  |  |  | No. / ha | SE | $\begin{array}{r} \text { Lower } \\ 95 \% \text { C.I } \end{array}$ | $\begin{array}{r} \text { Upper } \\ 95 \% \text { C.I } \end{array}$ | $\mathrm{kg} / \mathrm{ha}$ | SE | $\begin{array}{r} \text { Lower } \\ 95 \% \text { C.I } \end{array}$ | $\begin{array}{r} \text { Upper } \\ 95 \% \text { C.I } \end{array}$ |
|  |  | Estimate | SE | C.I |  | Estimate | SE | 95\% C.I | C.I |  |  |  |  |  |  |  |  |
| Age-0 | North | 1,005 | 308 | 362 | 1,648 | 1 | 0 | 1 | 2 | 7.2 | 2.2 | 2.6 | 11.8 | 0.0 | 0.0 | 0.0 | 0.0 |
|  | Central | 53,163 | 4,797 | 42,475 | 63,851 | 74 | 7 | 59 | 89 | 88.3 | 8.0 | 70.6 | 106.1 | 0.1 | 0.0 | 0.1 | 0.1 |
|  | South | 16,220 | 17,036 | - | 54,176 | 23 | 24 | - | 76 | 35.1 | 36.9 | - | 117.4 | 0.0 | 0.1 | 0.1 | 0.2 |
|  | Whole Lake | 70,388 | 17,701 | 33,198 | 107,578 | 99 | 25 | 46 | 151 | 58.5 | 14.7 | 27.6 | 89.4 | 0.1 | 0.0 | 0.1 | 0.1 |
| Age-1 | North | 1,250 | 344 | 532 | 1,967 | 30 | 8 | 13 | 47 | 8.9 | 2.5 | 3.8 | 14.0 | 0.2 | 0.1 | 0.1 | 0.3 |
|  | Central | 29,120 | 2,344 | 23,897 | 34,342 | 702 | 56 | 576 | 828 | 48.4 | 3.9 | 39.7 | 57.1 | 1.2 | 0.1 | 1.0 | 1.4 |
|  | South | 10,325 | 17,423 | - | 49,142 | 249 | 420 | - | 1,184 | 22.4 | 37.7 | - | 106.5 | 0.5 | 0.9 | - | 2.6 |
|  | Whole Lake | 40,694 | 17,583 | 3,752 | 77,636 | 981 | 424 | 90 | 1,871 | 33.8 | 14.6 | 3.1 | 64.5 | 0.8 | 0.4 | 0.1 | 1.6 |
| Age-2 and Age-3 |  |  | 140 |  | 920 | 85 | 19 | 46 | 124 | 4.5 | 1.0 | 2.4 | 6.6 | 0.6 | 0.1 | 0.3 | 0.9 |
|  | North | 629 39,304 | 140 3.839 | 30,750 | 47,857 | 5,306 | 518 | 4,151 | 6,461 | 65.3 | 6.4 | 51.1 | 79.5 | 8.8 | 0.9 | 6.9 | 10.7 |
|  | Central <br> South | 39,304 5,898 | r 17,700 | 30,50 | 45,334 | 5,796 | 2,390 | , | 6,120 | 12.8 | 38.4 | - | 98.2 | 1.7 | 5.2 | - | 13.3 |
|  | Whole Lake | 5,898 45,830 | 18,112 | 7,776 | 83,884 | 6,187 | 2,445 | 1,050 | 11,324 | 38.1 | 15.1 | 6.5 | 69.7 | 5.1 | 2.0 | - | 9.4 |
| Total |  | 156,912 | 53,396 | 44,726 | 269,098 | 7,266 | 2,894 | 4,334 | 13,346 | 130 | 44 | 84 | 224 | 6.0 | 2.4 | 0.5 | 11.1 |

Appendix 43. Population, biomass, and production estimates of kokanee, by age and region, at Coquitlam Reservoir, November 2005.

| Age | Region | Population (\#) |  |  |  | Biomass (kg) |  |  |  | Production (\#/ ha) |  |  |  | Standing Crop (kg / ha) |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  | Lower Upper 95\% |  |  |  | No. / ha | SE | Lower$95 \% \text { C.I }$ | $\begin{aligned} & \text { Upper } \\ & 95 \% \mathrm{C} . \mathrm{I} \end{aligned}$ | kg/ha | SE | Lower$95 \% \text { C.I }$ | $\begin{aligned} & \text { Upper } \\ & 95 \% \text { C.I } \end{aligned}$ |
|  |  |  |  |  |  | Estimate | SE | 95\% C.I | C.I |  |  |  |  |  |  |  |  |
| Age-0 | North | 2,070 | - | - | - | 3 | - | - | - | 14.8 | - |  |  | 0.0 | ${ }^{-}$ | $\bigcirc$ | 0 |
|  | Central | 10,936 | 2,872 | 4,538 | 17,335 | 15 | 4 | 6 | 24 | 18.2 | 4.8 | 7.5 | 28.8 | 0.0 | 0.1 | 0.0 | 0.0 |
|  | South | 11,215 | 31,694 | - | 81,830 | 16 | 44 | - | 115 | 24.3 | 68.7 | $3^{-}$ | 177.3 | 0.0 | 0.1 | 0.0 | 0.1 |
|  | Whole Lake | 24,221 | 31,824 | 4,538 | 91,084 | 34 | 45 | 6 | 128 | 20.1 | 26.5 | 3.8 | 75.7 | 0.0 | 0.0 | 0.0 |  |
| Age-1 |  |  | - | - | - | 63 | - | - | - | 18.7 | - | - | - | 0.5 | - | - | - |
|  | Central | 9,618 | 952 | 6,897 | 11,138 | 217 | 23 | 166 | 268 | 15.0 | 1.6 | 11.5 | 18.5 | 0.4 | 0.0 | 0.3 | 0.4 |
|  | South | 18,879 | 37,797 | 6,87 | 103,090 | 455 | 911 | - | 2,484 | 40.9 | 81.9 | - ${ }^{-}$ | 223.4 | 1.0 | 2.0 | - | 5.4 |
|  | Whole Lake | 30,514 | 37,809 | 6,897 | 109,950 | 735 | 911 | 166 | 2,650 | 25.4 | 31.4 | 5.7 | 91.4 | 0.6 | 0.8 | 0.1 | 2.2 |
| Age-2 and Age-3 | North | 1,516 | - | - | - | 205 | - | - | - | 10.8 | - | - ${ }^{-}$ | - | 1.5 | - | 5 | 1 |
|  | Central | 14,687 | 1,622 | 11,073 | 18,301 | 1,983 | 219 | 1,495 | 2,471 | 24.4 | 2.7 | 18.4 | 30.4 | 3.3 | 0.4 | 2.5 | 4.1 |
|  | South | 1,221 | 33,502 | - | 75,865 | 165 | 4,523 | - | 10,242 | 2.6 | 72.6 | ${ }^{-}$ | 164.4 | 0.4 | 9.8 | 25 | 22.2 9.9 |
|  | Whole Lake | 17,424 | 33,542 | 11,073 | 87,895 | 2,352 | 4,528 | 1,495 | 11,866 | 14.5 | 27.9 | 9.2 | 73.1 | 2.0 | 3.8 | 2.5 | 9.9 |
| Total |  | 72,159 | 103,175 | 22,508 | 288,929 | 3,122 | 5,484 | 1,667 | 14,643 | 60 | 86 | 19 | 240 | 2.6 | 4.6 | 2.6 | 12.2 |


[^0]:    
    "Suadurdiexd to lish per hour using 9 gl men net
    mudrdived to lish per how
    

